

Revised Risk Assessment for the Air Characteristic Study Volume I Overview

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Office of Solid Waste
U.S. Environmental Protection Agency
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Executive Summary

The U.S. Environmental Protection
Agency (EPA), Office of Solid Waste, has
analyzed the potential direct inhalation risks
that may result from unregulated emissions
from certain waste management units. This
document (Volume I) presents an overview of
the revised risk assessment for that analysis,
also referred to as the Revised Risk
Assessment for the Air Characteristic Study.
Volume II is the Technical Background
Document and Volume III (on CD-ROM)
presents results.

The Air Characteristic Study

This report and the 1998 Air Characteristic Study are among the initial steps for the EPA in fulfilling a long standing goal to review the adequacy and appropriateness of the hazardous waste characteristics.

The first step for EPA in achieving this goal was the Hazardous Waste Characteristic Scoping Study (November 1996), in which the Agency investigated potential gaps in the characteristics. The Scoping Study identified direct inhalation risks from emissions of waste management units as one potential gap in the Resource Conservation and Recovery Act (RCRA) hazardous waste characteristics. The Agency then completed the Air Characteristic Study (May 1998) as the next step in the process. The Air Characteristic Study examined the potential direct inhalation risks due to emissions from certain waste management units. In accordance with Agency policy, the technical work performed

for the 1998 Air Characteristic Study was peer-reviewed. This report contains the revised Air Characteristic Risk Assessment based on peer-review and public comments.

Revised Risk Analysis

This study is a national analysis to evaluate the possible need for an air characteristic. As such, this study was designed to highlight areas that may require a more detailed review before any formalized regulatory development work is initiated.

The overall goal of the risk analysis is to estimate waste concentrations that could be present in certain waste management units (WMUs) and still be protective of human health. Concentrations at specified risk levels were estimated at six different distances for a subset of constituents that could be present in wastepiles, landfills, land application units, storage tanks, and aerated and nonaerated treatment tanks. The analysis is based on modeling the emissions from a waste management unit, transport through the ambient environment, and exposure to a receptor to backcalculate to a threshold concentration in waste below which the risk to human health would fall below a preestablished threshold. To accomplish this, we characterized waste sources, applied peerreviewed and commonly used emissions and dispersion models, and established a Monte Carlo analysis to capture variabilities in receptor characteristics, such as exposure parameters and location around a facility. Chronic exposures were evaluated for 104 of

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the 105 constituents, and acute and subchronic exposures were considered for 35 and 64 constituents, respectively. In addition, protective concentrations in waste were estimated for five receptor categories: an adult resident, a child resident with exposure starting between 0 and 3 years old, a child resident with exposure starting between 4 and 10 years old, a child resident with exposure starting between 11 and 18 years old, and an off-site worker.

Distances of 25, 50, 75, 150, 500, and 1.000 meters were used as the basis for backcalculated risk-based waste concentrations. The resulting waste concentrations were considerably higher for receptors at the 500- and 1,000-m distances. A sensitivity analysis conducted on the dispersion component of this analysis indicated that there is a sharp decline in air concentration after the 150-m distance. The 25-m distance produces the lowest waste concentrations but is also an unlikely exposure scenario. The 50-, 75- and 150-m results were very similar to each other (within a factor of 2 to 3). This report displays results for only the 25-, 150- and 1,000-m distances. Results for the remaining distances are provided in Volume III: Results.

Results of the risk analysis indicate that the lowest estimated protective waste concentrations (e.g., highest risk) were for the aerated and nonaerated treatment tanks. Aeration increases the potential for a chemical to be emitted to the air, which results in a higher emission rate per unit area for these

tanks relative to the other units. Nonaerated tanks are typically bigger than aerated tanks, resulting in similar total emissions. In general, the estimated protective waste concentrations for treatment tanks were lower than the other units by about an order of magnitude or more. Following aerated and nonaerated treatment tanks, the WMU ranking was storage tanks, land application units, landfills, and wastepiles.

Of the receptors evaluated, the protective waste concentrations for adult residents were lowest (i.e., highest risk), followed by the child residents, from youngest to oldest. The estimated waste concentrations for the offsite worker were about an order of magnitude higher than those for residents. The differences in the results for the resident scenarios can be attributed to the variation in assumed exposure duration. The exposure duration used in the risk modeling was greatest for the adult, followed by the child residents, and finally the off-site worker. For the chronic exposures, it appeared that the most important factor affecting the results was the chemical's toxicity. The chemicals with the lowest protective waste concentrations, and so highest risk, were among the most toxic.

No clear pattern emerged from the chronic, subchronic, and acute results. Subchronic and acute results may be lower or higher than chronic results depending on the chemical, and the difference ranges from negligible up to 2 orders of magnitude in either direction. The most likely reason for this is that the hazard posed by a chemical is likely to vary with exposure duration, i.e., some chemicals have greater hazard at chronic exposures; others at acute and subchronic exposures.

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Note that one chemical of the original 105, 3,4-dimethylphenol, was addressed, but risks could not be quantified because data were insufficient to develop a health benchmark.

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Integrating the Revised Risk Assessment with the 1998 Analyses

In order to determine the need for an Air Characteristic, the Agency conducted two other analyses in 1998 along with the risk assessment. These analyses on regulatory coverage and constituent occurrence were to ascertain the current management of the 105 constituents. Integrating the results from these two analyses with the risk assessment results would help the Agency identify the nature and extent of gaps in regulatory coverage and the significance of the resulting human health risks.

This step was repeated for this task. The results from the revised risk assessment were combined with the results of the regulatory gaps analysis and the occurrence analysis from the 1998 Air Characteristic Study. This comparison showed that 16 constituents were neither associated with a listing nor on the Toxicity Characteristic (TC) list under RCRA. Two of these constituents had concentrations in tanks less than 100 ppm. In addition, 2 of these 16 constituents were not on the Clean Air Act's hazardous air pollutants (HAPs) list.

Three constituents had estimated protective waste concentrations lower than the TC or TC-derived waste concentration. Two constituents had TC levels that may not be

protective of air pathway risks for tanks, and two constituents had waste concentrations more stringent than TC levels for land-based units. The magnitude of the difference between the TC and the estimated air characteristic waste concentrations (C_w) varied according to the waste management unit and the constituent.

Land disposal restrictions (LDRs) and the protective concentrations in waste were compared, and results indicated that the treatment standards are not always below the levels at which there are potential air risks. Two constituents had concentrations in waste for chronic exposures that were below the LDR treatment levels. No constituents had concentrations in waste that were below the LDR treatment levels for acute or subchronic exposures.

Next Steps

Should EPA decide this analysis identifies constituents and waste management units of potential significance as unregulated emissions of possible concern, EPA has a range of options. EPA could decide to further study and potentially address these issues through regulation under the CAA, RCRA, or both. Further analysis would be needed before any new regulatory action could be promulgated.

1.0 Introduction

The U.S. Environmental Protection Agency (EPA), Office of Solid Waste (OSW), has analyzed the potential risks to human health posed by the inhalation of vapor (gaseous) and particulate (nongaseous) air emissions from a set of chemicals and metals when managed in certain waste management units (WMUs). An analysis of these risks was initially performed in 1998 as part of the *Air Characteristic Study* (U.S. EPA, 1998a). In accordance with Agency policy, the risk assessment conducted for the 1998 Air Characteristic Study was peer reviewed to ensure that science was used credibly and appropriately in the work performed. Based on comments made by the peer reviewers, EPA has revised the original risk assessment.

This report presents the revised risk assessment in three volumes. This document is Volume I, the Overview. This volume provides a discussion of the changes made from the 1998 Air Characteristic Study, a general overview of the risk assessment, a summary of results of the risk assessment, and the integration of the revised risk assessment results with the May 1998 regulatory gaps and occurrence analyses. A detailed description of the methodologies, data, and supporting analyses used for the risk assessment may be found in Volume II, *Revised Risk Analysis for the Air Characteristic Study: Technical Background Document.* The complete results of the analysis are presented in Volume III, *Revised Risk Analysis for the Air Characteristic Study: Results* (on CD-ROM).

1.1 Purpose and Requirements of the Air Characteristic Study

This report and the 1998 Air Characteristic Study are among the initial steps for EPA in fulfilling a long-standing goal to review the adequacy and appropriateness of the hazardous waste characteristics. The first step in achieving this goal was the Hazardous Waste Characteristic Scoping Study (U.S. EPA, 1996), which the Agency completed November 15, 1996, under a deadline negotiated with the Environmental Defense Fund. This study was conducted to identify potential gaps in the current hazardous waste characteristics, as well as other modifications and updates that are necessary to ensure that the definition of characteristics is complete, up-to-date, and based on state-of-the-art methodologies. Based on the initial bounding analysis of potential risks due to air emissions done as part of the Scoping Study, as well as follow-up analysis on potential gaps in regulatory coverage under the Clean Air Act (CAA) and Subpart CC of the Resource Conservation and Recovery Act (RCRA), OSW identified air emissions from WMUs as one of the areas meriting further analysis.

The Air Characteristic Study addresses this area by examining the potential direct inhalation risks due to emissions from certain WMUs. On May 15, 1998, in accordance with a consent decree, EPA completed the first portion of the study. According to the consent decree with EDF, a second part of the Air Characteristic Study, covering surface impoundments

receiving wastewaters that never exhibited a characteristic, will be completed March 26, 2001. The purpose of the 1998 Air Characteristic Study, as outlined by the consent decree, was to investigate gaps in the current hazardous waste characteristics and CAA programs. In addition, resulting potential risks to human health posed by the inhalation of air emissions from wastes managed in certain WMUs were to be investigated.

The 1998 Air Characteristic Study has three components: an evaluation of the coverage and potential regulatory gaps in RCRA Subtitle C and the CAA, a risk analysis of air emissions from WMUs, and an evaluation of the occurrence of these constituents in nonhazardous industrial waste. The risk assessment component has undergone a peer review, and EPA has made a number of changes to the risk assessment based on peer reviewer comments. In addition, other revisions have been made based on public comments and improvements initiated by the Agency. Since the other components of the May 1998 Air Characteristic Study have not been revised, those analyses are not covered in this document. The original results of these analyses are used in this report to present the significant findings from the integration of the revised risk assessment results with the regulatory gaps and occurrence analyses.

1.2 Overview of Risk Assessment

The risk assessment described in this document is a national analysis designed to assess the potential human health risk attributable to inhalation exposures when certain chemicals and metals are managed as waste in certain types of WMUs. The purpose of the analysis is to determine which chemicals and waste management units are of potential national concern purely from a risk perspective; it is not intended to draw conclusions concerning regulatory coverage. This information, combined with preliminary information on regulatory coverage and on the presence of these chemicals in nonhazardous waste, will be useful in determining the need for expanded regulatory coverage. Specifically, the purpose of this study is to provide technical information on the potential risk from WMU emissions to help EPA determine the need to expand regulatory coverage in the future.

The analysis presented in this report addresses specific chemicals that when managed as a waste may pose a risk through direct inhalation exposures. Tables 1-1 and 1-2 list the chemicals and metals included in this analysis. The analysis is structured so that the results of the risk assessment are the concentrations of each constituent that can be present in each type of WMU and still be protective of human health. The protective concentrations in waste were developed for three types of receptors: adult residents, child residents, and workers. Three risk endpoints— chronic (over 1 year), subchronic (1 month), and acute (1 day)— were evaluated.

The protective waste concentrations were estimated by modeling the emissions from a waste management unit, the transport through the ambient environment, and the exposure to a receptor to backcalculate a threshold concentration in a waste below which the risk to human health would fall below a pre-established threshold. The waste management scenario modeled in this analysis is storage, disposal, or treatment of industrial waste streams in RCRA subtitle D WMUs.

Table 1-1. Constituents Modeled for All WMUs

Constituent	CAS No.
Acetaldehyde [ethanal]	75-07-0
Acetone [2-propanone]	67-64-1
Acetonitrile [methyl cyanide]	75-05-8
Acrolein	107-02-8
Acrylonitrile	107-13-1
Allyl chloride	107-05-1
Arsenic	7440-38-2
Barium	7440-39-3
Benzene	71-43-2
Beryllium	7440-41-7
Bromodichloromethane [dichlorobromomethane]	75-27-4
Bromoform [tribromomethane]	75-25-2
Bromomethane [methyl bromide]	74-83-9
1,3-Butadiene	106-99-0
Cadmium	7440-43-9
Carbon disulfide	75-15-0
Carbon tetrachloride	56-23-5
Chlorobenzene	108-90-7
Chlorodibromomethane [dibromochloromethane]	124-48-1
Chloroform	67-66-3
Chloromethane [methyl chloride]	74-87-3
Chloroprene [2-chloro-1,3-butadiene]	126-99-8
Chromium VI	7440-47-3
Cobalt	7440-48-4
Cumene [isopropyl benzene]	98-82-8
Cyclohexanol	108-93-0
1,2-Dibromo-3-chloropropane	96-12-8
1,2-Dichlorobenzene [o-dichlorobenzene]	95-50-1
1,4-Dichlorobenzene [p-dichlorobenzene]	106-46-7
Dichlorodifluoromethane [CFC-12]	75-71-8
1,2-Dichloroethane [ethylene dichloride]	107-06-2
1,1-Dichloroethylene [vinylidene chloride]	75-35-4
1,2-Dichloropropane [propylene dichloride]	78-87-5
cis-1,3-Dichloropropylene	10061-01-5
trans-1,3-Dichloropropylene	10061-02-6
1,4-Dioxane [1,4-diethyleneoxide]	123-91-1
Epichlorohydrin [1-chloro-2,3-epoxypropane]	106-89-8
1,2-Epoxybutane	106-88-7
2-Ethoxyethanol [ethylene glycol monoethyl ether]	110-80-5
2-Ethoxyethanol acetate [2-EEA]	111-15-9
Ethylbenzene	100-41-4
Ethylene dibromide [1,2-dibromoethane]	106-93-4
	(continued)

(continued)

Table 1-1. (continued)

Constituent	CAS No.
Ethylene oxide	75-21-8
Formaldehyde	50-00-0
Furfural	98-01-1
Hexachloroethane	67-72-1
<i>n</i> -Hexane	110-54-3
Lead	7439-92-1
Manganese	7439-96-5
Mercury	7439-97-6
Methanol	67-56-1
2-Methoxyethanol	109-86-4
2-Methoxyethanol acetate [2-MEA]	110-49-6
Methyl tert-butyl ether	1634-04-4
Methylene chloride [dichloromethane]	75-09-2
Methyl ethyl ketone [2-butanone][MEK]	78-93-3
Methyl isobutyl ketone [hexone] [4-methyl-2-pentanone]	108-10-1
Methyl methacrylate	80-62-6
Naphthalene	91-20-3
Nickel	7440-02-0
2-Nitropropane	79-46-9
N-Nitrosodi- <i>n</i> -butylamine	924-16-3
N-Nitrosodiethylamine	55-18-5
N-Nitrosopyrrolidine	930-55-2
Propylene oxide	75-56-9
Pyridine	110-86-1
Styrene	100-42-5
1,1,1,2-Tetrachloroethane	630-20-6
1,1,2,2-Tetrachloroethane	79-34-5
Tetrachloroethylene [perchloroethylene]	127-18-4
Toluene	108-88-3
1,1,1-Trichloroethane [methyl chloroform]	71-55-6
1,1,2-Trichloroethane [vinyl trichloride]	79-00-5
Trichloroethylene	79-01-6
Trichlorofluoromethane [trichloromonofluoromethane]	75-69-4
1,1,2-Trichloro-1,2,2-trifluoroethane [freon 113]	76-13-1
Triethylamine	121-44-8
Vanadium	7440-62-2
Vinyl acetate	108-05-4
Vinyl chloride	75-01-4
Xylenes, mixed isomers [xylenes, total]	1330-20-7

Table 1-2. Constituents Modeled for Tanks Only

	GAGN
Constituent	CAS No.
Acrylamide	79-06-1
Acrylic acid	79-10-7
Aniline	62-53-3
Benzidine	92-87-5
Benzo(a)pyrene	50-32-8
2-Chlorophenol [o-chlorophenol]	95-57-8
Cresols, total	1319-77-3
7,12-Dimethylbenz[a]anthracene	57-97-6
N,N-Dimethyl formamide	68-12-2
3,4-Dimethylphenol	95-65-8
2,4-Dinitrotoluene	121-14-2
1,2-Diphenylhydrazine	122-66-7
Ethylene glycol	107-21-1
Hexachlorobenzene	118-74-1
Hexachloro-1,3-butadiene [hexachlorobutadiene]	87-68-3
Hexachlorocyclopentadiene	77-47-4
Isophorone	78-59-1
3-Methylcholanthrene	56-49-5
Nitrobenzene	98-95-3
Phenol	108-95-2
Phthalic anhydride	85-44-9
2,3,7,8-TCDD [2,3,7,8-tetrachlorodibenzo- <i>p</i> -dioxin]	1746-01-6
o-Toluidine	95-53-4
1,2,4-Trichlorobenzene	120-82-1

Emissions, transport, and exposure were modeled somewhat differently for the three risk endpoints (chronic, subchronic, and acute). For emissions and transport, different averaging times were used for each endpoint (1 year for chronic, 1 month for subchronic, and 1 day for acute) to generate emission rates and dispersion factors. For exposure, subchronic and acute exposures were modeled deterministically, using the point of maximum exposure at a specific distance. Chronic exposures were modeled probabilistically using a Monte Carlo approach to capture variation in receptor location and exposure factors. The WMUs assessed are aerated treatment tanks, nonaerated treatment tanks, storage tanks, landfills, waste piles, and land application units. The risk assessment was structured to capture national variations in environmental settings. In addition, Monte Carlo analysis was used in the modeling to include the variations in receptor characteristics such as exposure parameters and location around the facility.

1.3 Organization of Report

The remainder of this report is organized as follows. Section 2 summarizes changes made from the 1998 Air Characteristic Study. Section 3 provides a general overview of the risk analysis. Section 4 presents the revised risk analysis results. Section 5 presents the integration of the revised risk assessment results with the May 1998 regulatory gaps and occurrence analyses. References are provided in Section 6, and supporting analyses are included in Appendix A.

1.4 Companion Documents

Volume II of this report, the *Technical Background Document*, provides a detailed description of the methodologies, data, and supporting analyses used for the risk assessment.

Volume III of this report, *Results* (provided on CD-ROM), presents the detailed results of the risk analysis.

2.0 Revisions to the Risk Assessment Framework

The analytical approach used for this analysis differs in important ways from the approach used for the May 1998 Air Characteristic Study. Changes have been made to the risk assessment to improve the robustness of the analysis, reduce uncertainty, and make corrections to the 1998 study. The changes reflect comments made by peer-reviewers and public commenters and other improvements made by the Agency. Several aspects of the Air Characteristic Study risk assessment were modified, including source characterization, emissions modeling, air dispersion modeling, health benchmarks, and exposure and risk modeling. These changes are discussed in the following sections.

2.1 Source Characterization

Several changes were made to improve the source characterization. The source characterization is the information about waste management unit (WMU) dimensions and operations that defines how Industrial D waste is managed. It is important to accurately establish these characteristics since they influence the rate of emissins and amount of dispersion of a constituent.

Changes in source characterizations have affected all the WMU categories. These changes are discussed for each WMU in the following sections. Except for tanks, source characterizations were and still are based on the Subtitle D Survey (Schroeder et al., 1987); however, the actual number of units included in the analysis from that survey has increased slightly. Those increases are also discussed in the following sections.

2.1.1 Landfills

In both the May 1998 study and the current study, landfills were modeled assuming that the landfill is divided into an equal number of cells. The cell size is determined by dividing the total landfill area by the landfill life, 20 years, creating 20 cells for each landfill. One cell operates for 1 year for the life of the landfill. Each cell is assumed to be covered at the end of the year, preventing further emissions from that cell. Thus, emissions are only occurring from one cell at any given time, or from an area equal to one-twentieth of the total area. Emissions from the open cell are modeled as an emission rate per unit area (g/m²-s), and total emissions in g/s are then calculated by multiplying this per-unit-area emission rate by the area from which emissions occur. In the May 1998 study, the total area, instead of the cell area, was used to calculate total emissions. This error has been corrected in the current study. This correction reduces total emissions by a factor of 20 and air concentration (and therefore risk) by about a factor of 10 (because dispersion is not linear on source area). As a result of this change, the protective waste concentrations for landfills are higher by about a factor of 10.

In the May 1998 study, 790 landfills were modeled from the Subtitle D survey data. This reflected a total of 827 landfills reported, 37 of which were culled for various reasons. Eleven of the 37 were culled based on results of previous groundwater modeling work by EPA not related to the Air Characteristic Study. Those 11 sites are, however, relevant to the Air Characteristic Study. Because retaining as many sites as possible is desirable, those 11 sites were included this time, resulting in a database of 801 landfills for the current study. The culls made for the current study are detailed in Volume II, Section 3.1.

2.1.2 Land Application Units

The Subtitle D survey does not provide data on application frequency for land application units (LAUs). In the May 1998 study, application frequency was assumed to be four times per year. Sensitivity analysis shows that the application frequency has a significant impact on emissions even when total annual waste quantity is held constant: the more frequent the applications, the greater the emissions. This is due in part to the fact that tilling is presumed to occur whenever waste is applied, and tilling increases emissions by disturbing the waste. Therefore, we reviewed several data sources in an effort to better characterize application frequency for LAUs. These sources included Land Treatment Practices in the Petroleum Industry (Environmental Research & Technology, 1983); Review and Evaluation of Current Design and Management Practices for Land Treatment Units Receiving Petroleum Wastes (Martin et al., 1986); and Handbook of Land Treatment Systems for Industrial and Municipal Wastes (Reed and Crites, 1984). Data in these sources were used to establish a relationship between the number of applications per year and the annual waste quantity managed. This relationship was applied to the LAUs in the Subtitle D survey to establish a distribution of application frequencies relevent to Industrial D LAUs. An application frequency of 24 times per year was selected for use in this study, reflecting a central tendency value from the distribution (see Volume II, Section 4.5.1 for more details).

The increase in application frequency from 4 applications per year to 24 applications per year should increase emissions (and therefore risk) and decrease the protective waste concentration.

In the May 1998 study, 308 land application units were modeled from the Subtitle D survey data. This reflected a total of 354 land application units reported, 46 of which were culled for various reasons. Thirty-seven of the sites culled were culled based on previous groundwater modeling work by EPA not related to the Air Characteristic Study. Those 31 sites are relevant to the Air Characteristic Study. Because retaining as many sites as possible is desirable, those sites were included this time. Many of these sites had been culled because the reported waste quantity and area implied an unrealistically large application rate (greater than 10,000 tons/acre/yr). Those sites were retained in this study, and new waste quantities were imputed that did not violate this criterion (see Volume II, Section 3.1.3 for more details). This resulted in a database of 345 land application units for the current study. The culls made for the current study are detailed in Volume II, Section 3.1.

2.1.3 Wastepiles

The Subtitle D survey does not provide data on wastepile height. In the May 1998 study, wastepiles were modeled based on two assumed heights (2 and 5 meters). However, wastepile height is related to wastepile area, waste quantity, and retention time. Therefore, it is more realistic to evaluate the characteristics of each wastepile and assign a height individually. Wastepile area and waste quantity are reported in the Subtitle D survey, but retention time is not. Therefore, to tailor the wastepile heights to known data, a relationship between wastepile area, waste quantity, and height was developed (see Volume II, Section 3.1.1 for more details). In this study, each wastepile modeled was assigned a height based on that relationship. Height is used only in the dispersion modeling and affects air concentration; the greater the height, the greater the dispersion, and so the lower the air concentration at a particular location. The dispersion model is not sensitive to small changes in height; therefore, to simplify dispersion modeling without sacrificing accuracy, a set of six discrete heights, covering the range of heights calculated for all the wastepiles, was used. These heights were 1, 2, 4, 6, 8, and 10 meters.

Approximately 77 percent of the wastepiles modeled were assigned a height of 1 m, and 95 percent of the wastepiles were assigned a height of 4 m or less. Therefore, most of the wastepiles are modeled at a lower height in the current study than in the May 1998 study. As a result, air concentration (and therefore risk) will tend to be greater and the protective waste concentration lower. This difference is significant: about a factor of 2 to 10 relative to the 2-m wastepiles in the May 1998 study and a factor of 2 to 25 relative to the 5-m wastepiles in the May 1998 study.

In the 1998 study, 742 wastepiles were modeled from the Subtitle D survey data. This reflected a total of 853 wastepiles reported, of which 111 were culled for various reasons (most because they are Bevill facilities, which are exempt from Subtitle C regulation and would therefore never be subject to an Air Characteristic under Subtitle C). Three of the sites culled were culled based on previous groundwater modeling work by EPA not related to the Air Characteristic Study. Those three sites are relevant to the Air Characteristic Study. Because retaining as many sites as possible is desirable, those three sites were included this time, resulting in a database of 745 wastepiles for the current study. The culls made for the current study are detailed in Volume II, Section 3.1.

2.1.4 Tanks

The tank source category has been revised extensively, with respect to both how tanks are characterized and the categories of tanks modeled.

Because the Subtitle D survey did not contain data on tanks, they were characterized in the May 1998 study using two model tanks placed at 29 locations. A full distribution of tanks, using a database of many actual tank facilities (as was done for the other WMUs), would provide a more representative result. However, data on Industrial D tanks do not exist; therefore, in the current study, tanks were characterized using tank data from the 1986 National Survey of Hazardous Waste Treatment, Storage, Disposal, and Recycling Facilities (TSDR) Database (U.S. EPA, 1987) (see Volume II, Section 3.4 for more details). These data provide a distribution of tanks to represent the range of tank configurations used in the United States. This is an important

step in improving the analysis because some tank characteristics are critical parameters in the emissions modeling. However, the TSDR tank database did not include data on all parameters needed for the emission and dispersion modeling; therefore, data from site visits to tanks done by the Agency in 1985 and 1986 in support of the development of RCRA Air Emission Standards were used to develop some of the tank-specific parameter values to characterize tank engineering and operating parameters (see Volume II, Section 3.4.2 for more details). This introduces some uncertainty into the tank characterization; however, we believe this uncertainty to be less than the uncertainy arising from the use of only two model tanks to characterize the universe of Industrial D tanks.

In the May 1998 study, four categories of tanks were modeled: aerated tanks with and without biodegradation and storage tanks with and without biodegradation. These categories do not capture nonaerated treatment tanks. Based on the TSDR tank data, nonaerated treatment tanks appear to differ from storage tanks (which are also nonaerated) in important ways, particularly with respect to the distribution of area. In addition, storage tanks are not designed for biodegradation, so the category of storage tank with biodegradation is not representative of real tanks. The tank categories modeled in the current study include aerated treatment tanks, nonaerated treatment tanks, and storage tanks. Some of the aerated treatment tanks were modeled with biodegradation, while others were modeled without biodegradation, depending on the treatment process reported. Nonaerated treatment tanks, like storage tanks, are typically not optimized for biodegradation; therefore, both nonaerated treatment tanks and storage tanks were modeled with no biodegradation.

Figure 2-1 shows the distribution of area for the three tank categories modeled in this study and shows where the two model tanks used in the 1998 study fall relative to those distributions. Table 2-1 shows exactly where the two model tanks fall in the new tank distributions. Because the two model tanks both fall relatively high in the new distributions, the use of the new distributions will tend to decrease tank size (and therefore risk) and increase the 90th percentile protective waste concentration. The elimination of biodegradation from many of the tanks will increase emissions (and therefore risk) and tend to decrease the protective waste concentration relative to tank types with biodegradation from the 1998 study.

2.2 Emissions Modeling

Changes in emissions modeling have affected all of the land-based WMUs. Tank emissions modeling was not changed; the new distribution of tanks is modeled in the same manner as in the May 1998 study with regard to emissions estimates. Changes in emissions modeling for land-based units were considered for both volatile and particulate emissions.

For particulate emissions, most of the parameter values used to compute particulate emission rates are site-specific. Some vary with the waste (e.g., silt content of the waste), others with the waste management unit (e.g., roughness height of unit, vegetative cover on unit), and still others with the location of the unit (e.g., meteorological parameters like precipitation data and windspeed). The 1998 study did account for the variability of the meteorological parameters by varying these based on assigned location of the unit; however, variability in the other parameters was not captured. For the current study, consideration was given to developing distributions of the other parameter values in the particulate emissions model for use in the

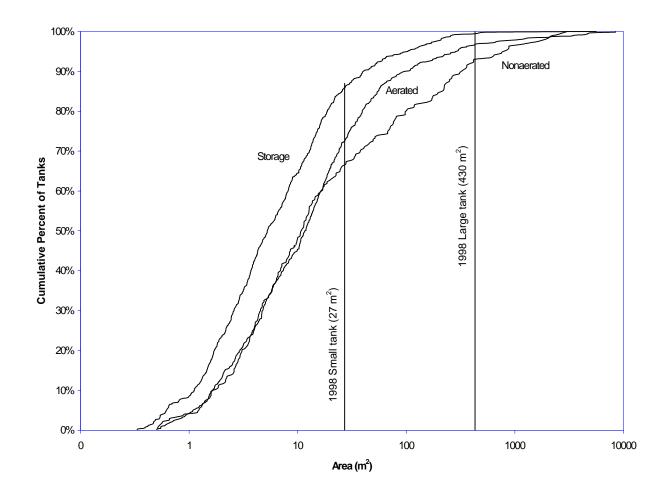


Figure 2-1. Cumulative tank distributions used in the current study.

Table 2-1. Percentiles of New Tank Distributions Associated with Two Model Tanks

	Percent of new tanks that are s	maller than the model tank:
Type of Tank	Small model tank (27 m²)	Large model tank (430 m²)
Aerated treatment	73%	97%
Nonaerated treatment	66%	93%
Storage	86%	100%

Monte Carlo model, in order to capture variability in the protective waste concentration due to variation in these parameters and to provide a more complete description of the variability of the protective waste concentrations. However, no data were identified that would support development of distributions, so no change was made to the particulate emissions modeling. The exclusion of such distributions only affects the distribution of results for metals; particulate emissions for volatile constituents are trivial compared to volatile emissions, so such a refinement would not have a significant effect on the results for volatile constituents.

Modeling of volatile emissions from all land-based units was modified with respect to the treatment of adsorbtion. Adsorption/absorption is the tendency of a chemical or liquid media to attach or bind to the surface or fill the pores of particles in the soil or waste and therefore not volatilize into the air. This tendency of a chemical to adsorb to or absorb in particles is important to consider in estimating the concentration of the chemical on particles emitted to the air due to wind erosion. The CHEMDAT8 model estimates emissions from land-based WMUs using a simple emissions model that accounts for contaminant partitioning between a liquid waste matrix and the air, diffusion of vapors through a porous media, and contaminant loss through biodegradation. This model accounts for adsorbtion when the waste concentration entered is a liquid-phase concentration; however, it does not account for adsorbtion when a total waste concentration (i.e., liquid and solid phase) is entered. The assumption of an entered waste concentration in liquid phase was based on the petroleum wastes for which CHEMDAT8 was originally developed and may not apply to the chemicals considered in this analysis. Therefore, a method for including adsorptive partitioning for total waste concentrations was developed and used to modify CHEMDAT8 for the current study. The changes to the CHEMDAT8 code are shown in detail in Section 4.3 of Volume II. This change should tend to decrease emissions and risk and increase protective waste concentration; the extent of the decrease in emissions will be constituent-specific, depending on the constituent's tendency to adsorb to particles.

The LAU emissions model was changed substantially. Instead of quarterly meteorological data, which were used in the May 1998 study, monthly meteorological data were used. Monthly meteorological data are more consistent with the application rate used in this study (24 applications per year). In addition, changes were made to the approach for estimating long term emission rates for LAUs in the current study. The May 1998 version of the model used steady state assumptions to estimate long-term emission rates. This presumed that all chemicals reached steady state emissions immediately (i.e., concentration remaining in the unit remains constant over time because waste additions and losses balance each other). While most chemicals will reach steady state within 1 or 2 years, some chemicals take longer than that or may never reach steady state. In order to better address the time to reach steady state in the current study, emissions were estimated using a pseudo-steady state approach, in which a series of steady state solutions was calculated for many short time periods, and the resulting emission rates were averaged to estimate long-term emissions. Specifically, emissions were estimated on a monthly basis for 40 years, and monthly emission rates for year 40 were averaged to estimate long-term annual emission rates. The actual length of time to reach steady state is constituentspecific. Forty years was chosen as a sufficiently long time for all chemicals that would ever reach steady state to do so. For those constituents that reach steady state sooner, there is no difference between using the first year after steady state is reached (typically year 2 or 3) and year 40. However, for constituents that do take many years to reach steady state, this approach provides a more realistic estimate of emissions.

An error in soil biodegradation rates that affects chronic, subchronic, and acute volatile emission rates for LAUs and chronic volatile emissions for wastepiles was identified and corrected for the current study (the differences between chronic, subchronic, and acute are discussed in Section 3.1). In the May 1998 study, the soil biodegradation rates were erroneously labeled as half lives and therefore used incorrectly. While half life is related to the first-order biodegradation rate, the two are not interchangeable. When this error was discovered, all soil biodegradation rates were verified against the original data source (Howard et al., 1991). The current study correctly uses the verified biodegradation rates from Howard et al. (1991). The impact of this error is chemical-specific. Table 2-2 summarizes the direction of the error for the chemicals modeled in land-based units. For about half of these, the biodegradation rate was too low, resulting in an overestimate of emissions. Emissions modeled with the correct biodegradation rate will be lower, resulting in less risk and a higher protective waste concentration. Most of the remaining chemicals were relatively unaffected by this correction. For only one chemical was the incorrect biodegradation rate too high, resulting in an underestimate of emissions. The corrected emissions for this chemical will be higher, resulting in more risk and a lower protective waste concentration.

2.3 Air Dispersion Modeling

Several changes in dispersion modeling were implemented in the current study and make the dispersion modeling more accurate than in the May 1998 study.

In the May 1998 study, wet and dry depletion of the atmospheric concentrations (plume depletion) of vapors and particulates were not considered due to the great increase in run time of ISCST3 for area sources when depletion is modeled (run times with depletion for area sources are typically 15 to 30 times longer than run times without depletion) and the short timeframe for completing the study. However, plume depletion can have a significant effect on air concentration, especially for particulates, and many of the peer-review comments identified this as a serious shortcoming of the May 1998 study. Therefore, for this analysis, with more time available, the issue of depletion was revisited. In addition, since May 1998, it had come to light for other EPA work (the Hazardous Waste Identification Rule, or HWIR) that the precipitation data in the hourly meteorological data used in the dispersion model were incomplete (i.e., some hours had missing precipitation data, resulting in total precipitation less than actual precipitation), which could affect the amount of wet depletion occurring. Work had already been done for the HWIR project to interpolate missing precipitation data for many of the meteorological locations modeled in the Air Characteristic Study.

Several sensitivity analyses were conducted to assess the importance of wet and dry depletion for particulates and wet depletion for vapors (ISCST3 cannot model dry depletion of vapors; therefore, this could not be considered. However, dry depletion of vapors is expected to be negligible). These sensitivity analyses showed that for both vapors and particulates, wet depletion did not have a significant impact on air concentrations (differences were less than 2 percent), even using the more complete interpolated precipitation data developed for HWIR. Dry depletion of particulates, on the other hand, did have a significant effect on air concentration (differences ranged up to about 40 percent). Therefore, dry depletion of particles was included in the current study for all land-based units. This change will reduce the air concentration of

Table 2-2. Chemical-specific Effects of Biodegradation Rate Correction

		New soil	Old soil
CAS	Chemical	biodegradation rate (sec ⁻¹)	biodegradation rate (sec ⁻¹)
Higher emis	ssions, lower C _w		
	Dichloropropane, 1,2-	6.2E-09	1.1E-07
Lower emis	sions, higher C _w		
75070	Acetaldehyde	1.1E-06	1E-20
-	Acetone	1.1E-06	6.1E-10
	Acetonitrile	2.9E-07	2.4E-09
107028	Acrolein	2.9E-07	2.4E-09
107131	Acrylonitrile	3.5E-07	2.0E-09
	Allyl chloride	5.7E-07	1.2E-09
71432	Benzene	5.0E-07	1.4E-09
75274	Bromodichloromethane	4.5E-08	1E-20
106990	Butadiene, 1,3-	2.9E-07	1E-20
67663	Chloroform	4.5E-08	2.4E-09
98828	Cumene	1.0E-06	7.0E-10
108930	Cyclohexanol	4.5E-08	1E-20
106467	Dichlorobenzene, p-	4.5E-08	1E-20
10061015	Dichloropropylene, cis-1,3-	7.1E-07	9.8E-10
10061026	Dichloropropylene, trans-1,3-	7.1E-07	9.8E-10
106898	Epichlorohydrin	2.9E-07	2.4E-09
106887	Epoxybutane, 1,2-	6.2E-07	1E-20
111159	Ethoxyethanol acetate, 2-	2.9E-07	1E-20
110805	Ethoxyethanol, 2-	2.9E-07	2.4E-09
100414	Ethylbenzene	8.0E-07	8.7E-10
75218	Ethylene oxide	6.8E-07	1E-20
50000	Formaldehyde	1.1E-06	6.1E-10
67561	Methanol	1.1E-06	6.1E-10
110496	Methoxyethanol acetate, 2-	2.9E-07	1E-20
109864	Methoxyethanol, 2-	2.9E-07	1E-20
74839	Methyl bromide	2.9E-07	2.4E-09
74873	Methyl chloride	2.9E-07	2.4E-09
78933	Methyl ethyl ketone	1.1E-06	6.1E-10
108101	Methyl isobutyl ketone	1.1E-06	6.1E-10
80626	Methyl methacrylate	2.9E-07	2.4E-09
1634044	Methyl tert-butyl ether	4.5E-08	1E-20
75092	Methylene chloride	2.9E-07	2.4E-09
91203	Naphthalene	1.7E-07	4.2E-09
110543	n-Hexane	5.0E-07	1E-20
924163	N-Nitrosodi-n-butylamine	4.5E-08	1E-20
75569	Propylene oxide	6.5E-07	1E-20
110861	Pyridine	1.1E-06	6.1E-10
100425	Styrene	2.9E-07	2.4E-09
630206	Tetrachloroethane, 1,1,1,2-	1.8E-07	5.8E-09
79345	Tetrachloroethane, 1,1,2,2-	1.8E-07	3.8E-09
108883	Toluene	3.6E-07	1.9E-09

(continued)

Table 2-2. (continued)

CAS	Chemical	New soil biodegradation rate (sec ⁻¹)	Old soil biodegradation rate (sec ⁻¹)
76131	Trichloro-1,2,2-trifluoroethane, 1,1,2-	2.2E-08	1E-20
	Vinyl acetate	1.1E-06	1E-20
1330207	<u> </u>	2.9E-07	2.4E-09
Relatively u	<u> </u>		
7440382	Arsenic	0	1E-20
7440393	Barium	0	1E-20
7440417	Beryllium	0	1E-20
7440439		0	1E-20
75150	Carbon disulfide	0	1E-20
56235	Carbon tetrachloride	2.2E-08	3.1E-08
108907	Chlorobenzene	5.3E-08	1.3E-08
124481	Chlorodibromomethane	4.5E-08	1.6E-08
126998	Chloroprene	4.5E-08	1.6E-08
7440473	Chromium (total)	0	1E-20
7440484	Cobalt	0	1E-20
96128	Dibromo-3-chloropropane, 1,2-	4.5E-08	1.6E-08
95501	Dichlorobenzene, o-	4.5E-08	1.6E-08
75718	Dichlorodifluoromethane	4.5E-08	1.6E-08
107062	Dichloroethane, 1,2-	4.5E-08	1.6E-08
75354	Dichloroethylene, 1,1-	4.5E-08	1.6E-08
123911	Dioxane, 1,4-	4.5E-08	1.6E-08
106934	Ethylene dibromide	4.5E-08	1.6E-08
98011	Furfural	0	1E-20
67721	Hexachloroethane	4.5E-08	1.6E-08
7439921	Lead	0	1E-20
7439965	Manganese	0	1E-20
7439976	Mercury	0	1E-20
7440020	Nickel	0	1E-20
79469	Nitropropane, 2-	4.5E-08	1.6E-08
55185	N-Nitrosodiethylamine	4.5E-08	1.6E-08
930552	N-Nitrosopyrrolidine	4.5E-08	1.6E-08
127184	Tetrachloroethylene	2.2E-08	3.1E-08
75252	Tribromomethane	4.5E-08	1.6E-08
71556	Trichloroethane, 1,1,1-	2.9E-08	2.4E-08
79005	Trichloroethane, 1,1,2-	2.2E-08	3.2E-08
79016	Trichloroethylene	2.2E-08	3.1E-08
75694	Trichlorofluoromethane	2.2E-08	3.1E-08
121448	Triethylamine	0	1E-20
7440622	Vanadium	0	1E-20
75014	Vinyl chloride	4.5E-08	1.6E-08

particulates, reducing risk and increasing the protective waste concentration for land-based units. The change is only significant for metals, however, because particulate emissions of volatile constituents are negligible compared to volatile emissions.

As discussed earlier, several changes to the source characterization were made that required new dispersion modeling. These changes included the addition of more specific heights for wastepiles and the recharacterization of tanks.

Dispersion modeling for wastepiles was modified to better capture the effect of wastepile height on ground-level concentrations. Section 2.1.3 discusses these changes. The difference in the results is significant—about a factor of 2 to 10 relative to the 2-m wastepiles in the May 1998 study and a factor of 2 to 25 relative to the 5-m wastepiles in the May 1998 study.

Dispersion modeling for tanks was also modified to capture the range of area/height combinations reflected in the new tank characterizations. A total of 33 area/height combinations were modeled for tanks, compared to only 2 area/height combinations (corresponding to the two model tanks) in the May 1998 study. The overall effect of the new area-height combinations compared to the ones used last year is not clearly in one direction, as the effect of the change in areas has effects in the opposite direction of the effect of the change in heights. As shown in Figure 2-1, the two model tanks used in the May 1998 study fall fairly high on the distribution of tanks used in the current study. Therefore, many of the tanks modeled in the current study are smaller in area, which will tend to result in lower air concentrations, lower risk, and higher protective waste concentrations relative to the May 1998 study. However, the two heights modeled in the May 1998 study were also high relative to the distribution of heights modeled in the current study, which has the opposite effect: the generally lower heights will tend to increase air concentration and risk, and lower protective waste concentration.

Finally, an error in the interpolation of dispersion coefficients for wastepiles in the May 1998 study was discovered and corrected for the current study. In the May 1998 study, the areas used for the interpolation were those for tanks, not wastepiles, which resulted in interpolated UACs (and therefore air concentration and risk) that are too low by a factor of 2 to 3.

2.4 Human Health Benchmarks

Twenty-eight of the inhalation benchmarks used in the May 1998 study have been changed. The changes are summarized in Table 2-3. More detailed information is available in Volume II. Section 6.0.

In some cases, new IRIS or other published information became available during the past year that suggested a change in the inhalation benchmark for this study.

The progression of values from chronic to subchronic to acute benchmarks was also reviewed, especially when the chronic value exceeded the subchronic or acute value. The anticipated progression would reflect that high concentrations of a chemical can be tolerated without ill effect for shorter periods of exposure than for longer periods of exposure. Therefore,

Table 2-3. Summary of Issues and Changes for Inhalation Benchmarks
Used in the Air Characteristic Study

CAS	Name	Issue	Resolution
75-05-8	Acetonitrile	New IRIS RfC=0.06 mg/m³, appropriate to use as subchronic	Revise chronic and subchronic RfCs
7440-38-2	Arsenic	CalEPA acute REL updated	Revise acute RfC
75-15-0	Carbon disulfide	Chronic RfC target organ should be neurological; CalEPA acute REL updated	Revise chronic RfC target organ and acute RfC
7440-47-3	Chromium VI	New IRIS RfC= 1E-4 mg/m³ (particulates); revised intermediate MRL= 5E-4 mg/m³ (particulates)	Revise chronic and subchronic RfCs
1319-77-3	Cresols (total)	Received public comment on chronic RfC; subchronic lower than chronic (due to calculation error)	No revision on chronic RfC; revise subchronic RfC (1.2E-3 mg/m³)
108-93-0	Cyclohexanol	New FR RfC=2E-5 mg/m ³	Revise chronic RfC; recalculate subchronic (2E-4 mg/m³)
106-46-7	Dichlorobenzene, 1,4-	Chronic RfC target organ should be liver	Revise chronic RfC target organ
107-06-2	Dichloroethane, 1,2-	Acute RfC lower than subchronic	Revise subchronic RfC (acute MRL = chronic MRL, therefore should also = subchronic = 0.81 mg/m ³)
123-91-1	Dioxane, 1,4-	Acute RfC lower than subchronic; CalEPA acute REL updated	Update acute RfC (still incorrect progression - see text)
106-89-8	Epichlorohydrin	CalEPA acute REL updated	Revise acute RfC
111-15-9	Ethoxyethanol acetate, 2-	CalEPA acute REL updated	Revise acute RfC
110-80-5	Ethoxyethanol, 2-	Acute RfC lower than subchronic; CalEPA acute REL updated; chronic RfC target organ should be male reproductive and hematological	Update acute RfC (still incorrect progression - see text); revise chronic RfC target organ
78-59-1	Isophorone	New FR RfC=1.2E-2 mg/m ³	Revise chronic RfC; recalculate subchronic (1.2E-1 mg/m³)
7439-97-6	Mercury	CalEPA acute REL updated	Revise acute RfC
67-56-1	Methanol	CalEPA acute REL updated	Revise acute RfC
109-86-4	Methoxyethanol, 2-	Acute RfC lower than subchronic; CalEPA acute REL updated	Update acute RfC (still incorrect progression - see text)
78-93-3	Methyl ethyl ketone	CalEPA acute REL updated	Revise acute RfC
75-09-2	Methylene chloride	Revised acute MRL= 3 ppm (10 mg/m ³)	Revise acute RfC
91-20-3	Naphthalene	New IRIS RfC= 3E-3 mg/m ³	Revise chronic RfC; recalculate subchronic RfC (3E-2 mg/m³)
7440-02-0	Nickel	CalEPA acute REL updated	Revise acute RfC
108-95-2	Phenol	Public comment on chronic RfC; new FR RfC=6E-3 mg/m³; CalEPA acute REL updated	Revise chronic RfC as per FR; recalculate subchronic RfC (6E-2 mg/m³); revise acute RfC
75-56-9	Propylene oxide	CalEPA acute REL updated	Revise acute RfC
100-42-5	Styrene	CalEPA acute REL updated	Revise acute RfC
1746-01-6	TCDD, 2,3,7,8-	URF available (3.3E+1 per μg/m³) in HEAST	Add URF
127-18-4	Tetrachloroethylene	Acute RfC lower than subchronic; cancer benchmarks available - Superfund URF (5.8E-7 per μg/m³) and CSF (2E-3 per mg/kg/d)	No revision of acute RfC; revise URF & CSF
108-88-3	Toluene	Revised acute MRL = 4 ppm (15 mg/m ³)	Revise acute RfC

(continued)

Table 2-3. ((continued)	۱
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CAS	Name	Issue	Resolution
7440-62-2	Vanadium	Public comment on chronic RfC; acute RfC lower than subchronic	Revise subchronic RfC (subchronic = chronic = 7E-5 mg/m³); recalculate acute (7E-4 mg/m³)
1330-20-7	Xylenes (total)	Calculation error for chronic RfC	Revise chronic RfC (4E-1 mg/m³)

CSF = cancer slope factor
FR = Federal Register
MRL = minimal risk level
REL = reference exposure level
RfC = reference concentration

URF = unit risk factor

chronic noncarcinogenic benchmarks should be lower than subchronic benchmarks, and subchronic benchmarks should be lower than acute benchmarks (note that for chronic benchmarks, this comparison can only be meaningfully made for noncarcinogens, since the chronic carcinogenic slope factor cannot be directly compared to a subchronic or acute benchmark). This review resulted in some modifications; however, data were not available to correct all instances in which the progression from chronic to subchronic to acute was not as expected. Subchronic and acute benchmarks are typically obtained from different sources and based on different underlying studies than chronic benchmarks are. Inconsistencies in how the benchmarks were developed or the underlying studies used often accounts for the discrepancy in expected progression. In many cases, no set of benchmarks could be found that displayed the expected progression.

Finally, public commenters on the May 1998 study specifically identified benchmarks for four constituents that should be reviewed: cobalt, cresols, phenol, and vanadium. These reviews resulted in changes as well.

In addition to changes in the basic health benchmarks described above, a change was made in how the cancer slope factors were used for children. Slope factors are developed for adults, using an assumed body weight of 70 kg. In the May 1998 study, slope factors were adjusted for children based on actual body weight. However, based on peer-review comments and further discussion with Agency experts in cancer dose-response, this adjustment has been eliminated from the current study. Cancer slope factors are used as presented for children, without adjustment. It should be noted, however, that the differences between an adult's and a child's physiology are accounted for in this study by adjusting the appropriate exposure factors. This is discussed in the following section.

2.5 Exposure and Risk Modeling

Four changes were made to the exposure and risk model used in this analyis:

- # Update of the exposure factor distributions
- # New child exposure approach

- # Change in worker scenarios
- # Change in checks for nonlinearity.

Exposure factors used in this study include body weight, inhalation rate, and exposure duration. Distributions of these exposure factors were updated from those used in the May 1998 study. Previously, data for males were used. However, this does not account for differences between males and females. Males typically have higher body weights and inhalation rates than females, as well as higher inhalation rates per unit of body weight. For this update, data on both males and females were used to better capture the potential effects on the whole population, not just males. Because females have a lower inhalation rate per unit of body weight than males, the effect should be to lower the overall distribution of risk and increase the protective waste concentration.

Three child age groups, or cohorts, were used to model child exposures: 0 to 3, 4 to 10, and 11 to 18 years of age. These cohorts are unchanged from the May 1998 study and reflect the age cohorts for which inhalation rate data are available. In the May 1998 study, the results were presented as a single "child" receptor. For each iteration of the Monte Carlo analysis, a starting age for exposure was chosen at random from among 0, 4, and 11 years (the three cohort starting ages), with the probability of each of the three starting ages being chosen proportional to the total number of years in the cohort. A single exposure duration was associated with each of the three starting ages, based on the median for that age cohort, and this was not varied for a particular starting age. None of the three exposure durations, when coupled with their associated starting age, resulted in exposure past age 18. This approach does not fully capture the impact of different starting ages (since these were restricted to three) or the full variability of exposure duration.

For this study, the child exposure approach was modified to better capture variations in age at start of exposure and exposure duration. Results were calculated and saved separately for receptors falling into each of the three age cohorts at the start of exposure (note that exposure may last longer than just the range of ages in a cohort); these are presented as "child 0-3 years," "child 4-10 years," and "child 11-18 years." For each iteration for a cohort, exposure begins at a starting age selected at random within the cohort (with each year of age within the cohort having equal probability of being selected). An exposure duration is also selected at random for each iteration from a distribution for the cohort (so there are three exposure duration distributions, one per cohort). Exposure is then started at the selected starting age and continues through succeeding age cohorts as necessary until the exposure duration selected for that starting age is reached. Depending on the starting age and exposure duration selected, exposure may continue into adulthood.

Both on-site workers and off-site workers were included in the May 1998 study. For this study, the on-site workers are no longer included in the receptors modeled. Accordingly, because concentration at 0 m from the WMU was only used for the on-site worker scenario, it has been dropped. Off-site workers have been retained and are evaluated for all receptor locations, as before.

A units conversion error in the calculation of a hazard quotient for lead was identified in the May 1998 model and corrected for this study. Specifically, a units conversion factor to

convert air concentration from $\mu g/m^3$ to mg/m^3 was omitted. As a result, the hazard quotients for lead were too high by a factor of 1,000 and the protective waste concentrations too low by the same amount. Due to the modeling of different receptors for lead (children age 0 to 3 years and 3 to 7 years), the hazard quotient equation for lead was separate from the equation used for all other chemicals, so this error did not affect any other chemicals.

Finally, the approach to adjusting the results for nonlinearities in the emissions modeled has been changed. In the May 1998 study, two types of adjustments were made that have been dropped for the current study. These are discussed below.

- In the May 1998 study, waste concentrations were backcalculated using both an aqueous-phase emission rate (modeled using Raoult's law). Typically, the aqueous-phase emission rates are much higher than the organic-phase emissions rates, resulting in lower waste concentrations based on aqueous-phase emission rates, but for a few chemicals that was not the case. In those cases, the backcalculated concentration was adjusted in the May 1998 study to be based on the organic-phase emission rate. However, the fact that greater emissions occur from the organic phase does not alter the fact that the aqueous phase is a far more likely scenario for the waste management units modeled in this study. The Agency decided that this adjustment was unnecessarily worst-case; therefore, this adjustment was dropped in the current study, and all results are based on the organic-phase emission rates. Results that would be lower if based on the organic-phase emission rates are footnoted.
- # In the May 1998 study, the backcalculated waste concentration based on the aqueous-phase emission rate was compared to either the soil saturation concentration (for land-based units) or the solubility at neutral pH and a temperature of 20-25 °C. These are the theoretical maximum concentrations at which aqueous phase wastes can exist; at higher concentrations, the waste is organic phase. If the backcalculated waste concentration based on aqueous-phase emission rates exceeded the soil saturation concentration or solubility, then it was adjusted to be based on the organic-phase emission rate instead. However, the soil saturation concentration and solubility are both dependent on site- and wastespecific conditions such as temperature and pH. Therefore, a backcalculated waste concentration near the soil saturation concentration or solubility calculated for this study may be possible in some situations and not in others. Rather than artificially restrict the results to standard conditions, in the current study all results are based on the aqueous-phase emission rates. If this backcalculated concentration exceeds the soil saturation concentration or solubility calculated for this study, the result is footnoted, and the footnote identifies whether pure organic-phase component (i.e., 1 million ppm modeled as organic phase) results in a risk greater or less than the cutoff risk of 10⁻⁵ or the cutoff HQ of 1.

3.0 Summary of Risk Assessment Modeling Approach and Data Sources

The analysis described in this section and in the Technical Background Document and appendixes is designed as a national analysis to assess the potential risk attributable to inhalation exposures when certain chemicals and metals are managed as a waste in certain types of waste management units. Of particular interest are chemicals and metals managed as wastes that are not regulated under RCRA as hazardous wastes. The purpose of this analysis is to determine

which chemicals and waste management units are of potential national concern purely from a risk perspective; it is not intended to draw conclusions concerning regulatory coverage. This information, combined with preliminary information presented in the May 1998 Air Characteristic Study on regulatory coverage and on the presence of these chemicals in nonhazardous waste, will be useful in determining the possible need to expand regulatory coverage in the future.

This section provides a general overview of the approach and primary data sources used and discusses the major components of the analysis—emissions modeling, dispersion modeling, and exposure modeling/risk estimation. Technical details on the models and a complete set of inputs and associated references are provided in Volume II.

3.1 Overview of Modeling Approach

The overall goal of this risk analysis is to estimate the concentrations of constituents that can be present in a waste management unit (WMU) and remain protective of human health. These protective waste concentrations were calculated for 104 constituents¹ including volatiles, semi-volatiles, and metals. These

The Air Characteristic Study addresses:

- # 105 constituents
- # 4 WMU types
 - landfill
 - land application unit (LAU)
 - wastepile (WP)
 - tank
- # 5 receptors
 - adult resident, exposure starting age 19 years
 - child resident, exposure starting age 0-3 years
 - child resident, exposure starting age 4-10 years
 - child resident, exposure starting age 11-18 years
 - off-site worker
- # direct inhalation only
- # volatiles and particulates
- # 6 distances from the site
- # 3 risk endpoints or averaging times
 - chronic (over 1 year)
 - subchronic (1 month) LAU, WP
 - acute (1 day) LAU, WP

¹ 105 were addressed but one constituent, 3,4 dimethylphenol, did not have an inhalation benchmark.

constituents were selected for their potential to result in risk from inhalation exposure. Workers, adults, and children were evaluated for three different types of exposures or risk endpoints: chronic (over 1 year), subchronic (1 month), and acute (1 day). Estimating protective concentrations required a multistep modeling process that could relate the concentrations in ambient air at a receptor point that could create a health effect to a concentration in the waste management unit. To achieve this, the analytical approach for this analysis is based on three primary components:

- # Emissions modeling—characterizing emissions from a WMU
- # Dispersion modeling—describing the transport of these emissions through the ambient environment
- # Exposure modeling/risk estimation—estimating exposure to a receptor and then backcalculating to arrive at a waste concentration (C_w) that presents a risk equal to a prespecified risk level (e.g., 1 in 1 million, or 1E-6).

To illustrate the scenario that was modeled for this study, Figure 3-1 is a conceptual diagram of a waste site. Constituents managed in the WMU can be released as gases if they volatilize and as particulates if the constituent attaches to solid particles in the waste. Once the constituent is released from the site, the ambient air provides a medium for the transport of the airborne constituent. The direction the constituent travels and its concentration in the air are determined by meteorological conditions in the surrounding area such as wind direction, air temperature, and atmospheric stability at the time it is released. Because meteorological patterns are dynamic, the concentration of the constituents in the air varies over time and people who live and work at various locations around the WMU have different inhalation risks. The risk to an individual from the release of a constituent also depends upon characteristics of that individual such as body weight, inhalation rate, and the length of time that individual remains in the area around the WMU. These last characteristics are the reason that this assessment considers the

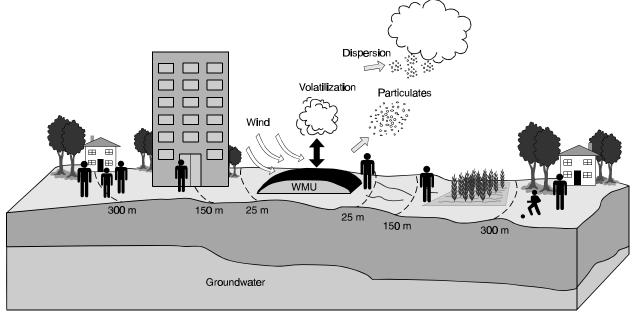


Figure 3-1. Conceptual diagram of a waste site.

exposure to multiple types of receptors: adult residents, child residents of various ages, and workers.

In order to model the scenario described above, the preliminary requirements for the analysis included:

- # Emissions models for the various WMUs to provide estimates of gas and particle releases from the unit
- # A dispersion model capable of modeling area sources for chronic (over 1 year), subchronic (1 month), and acute (1 day) releases
- # An exposure model for locating receptors proximate to the WMUs and estimating their exposure
- # A risk model that combines the exposure characteristics of different types of receptors with constituent-specific toxicity benchmarks.
- # The ability to backcalculate C_w from a prespecified risk level (e.g., 1E-6).

For each constituent and each WMU type, EPA wanted to be able to specify a C_w that would not exceed a target risk level (e.g., 1 in 100,000, or 1E-5) in more than a specified percentage (e.g., 10 percent) of the cases being modeled. Therefore, a probabilistic modeling approach, which would produce a distribution of C_w's, was needed, as opposed to a deterministic approach, which would only produce a point estimate. A deterministic analysis produces a point estimate because it uses a single value for each parameter in the analysis. A probabilistic approach considers the variability in the inputs required to estimate the concentration nationally. This type of approach produces a distribution of results because the method iterates through the analysis more than once, allowing the input parameters in the analysis to take on different values for each iteration from a distribution of values. For this analysis, EPA used a Monte Carlo simulation. This is a type of probabilistic analysis that can be used when the distribution of some or all input variables is known or can be estimated. A large number of iterations of the calculations are performed (i.e., 1,000), with a value for each input variable selected at random from the variable's distribution and the result (in this case, C_w) calculated for each iteration. The results of each iteration are combined into a distribution of C_w. It was assumed that the modeled cases represent the national distribution of risk-specific concentrations.

The probabilistic approach described above was used to model chronic exposures. A deterministic approach designed to produce a more high-end point estimate was used to model acute and subchronic exposures. The acute/subchronic approach uses the maximum exposure point at any given distance, so no variability in receptor location is accounted for. It also uses the meteorological conditions that produce the maximum air concentration for a 24-hour or 30-day time period over 5 years of meteorological data. The results from the acute/subchronic analysis are comparable to the 100th percentile of the distribution generated for the chronic analysis. It should be noted that acute/subchronic exposures were only assessed for land application units (LAUs) and wastepiles, which may have episodic loading events. There are a variety of other

differences in the acute/subchronic approach in how the emission rates and dispersion factors were calculated; these are described in more detail in Sections 3.2.2 and 3.2.3.

To estimate volatile emissions from each type of WMU, EPA's CHEMDAT8 model was used. For the landfill, LAU, and wastepile, the concentration of hazardous constituent in the surface layer of the soil (hereafter referred to as soil concentration) was estimated using a mass balance approach (i.e., competing pathways such as volatilization, adsorption, and biodegradation are accounted for). Particulate emissions due to wind erosion were modeled for land-based units (landfills, LAUs, and wastepiles). Landfills and LAUs were modeled as ground-level sources using the Cowherd model (U.S. EPA, 1985b, 1988). Wastepiles were modeled as elevated sources using the AP-42 model for wind erosion from aggregate storage piles (U.S. EPA, 1985a). To obtain the emission rate of constituent sorbed to particulate matter, the emission rate of particulate matter was multiplied by the soil concentration calculated by CHEMDAT8. This was done to account for the portion of the original constituent concentration that would remain in the waste after volatilization and biodegradation losses, and so would realistically be available for emission in the particulate phase.

The modeling assumes waste is continuously added to landfills and tanks, while LAUs and wastepiles have noncontinuous, episodic waste loadings. To capture potential peaks in emissions immediately after episodic loading events, acute and subchronic exposures were evaluated for LAUs and wastepiles.

Dispersion modeling was performed for each WMU using EPA's Industrial Source Complex Model Short-Term (ISCST3) to develop unitized air concentrations (UACs) for vapors and particulates. UACs are dispersion coefficients based on a unit emission (i.e., $1 \,\mu g/m^2 \cdot s$) for use in a backcalculation. UACs varied depending on the averaging time (i.e., chronic, subchronic, or acute), the size of the WMU, the distance and direction of the receptor from the WMU, and the associated meteorological station. Dispersion modeling for vapors did not account for depletion, as sensitivity analysis showed that depletion of vapors has a negligible impact on air concentration of vapors. Dispersion modeling for particulates accounted for dry depletion of particles, since a sensitivity analysis showed that dry depletion has a potentially significant impact on air concentrations of particulates. Wet depletion of particulates was not accounted for in the dispersion modeling, as sensitivity analysis showed that wet depletion has little impact on air concentration.

The air concentration at any specific receptor is the product of the emission rate (in $\mu g/m^2 \cdot s$) and appropriate UAC (in $[\mu g/m^3]/[\mu g/m^2 \cdot s]$). Air concentrations were estimated for chronic, subchronic, and acute exposures (using averaging times of 1 year, 1 month, or 1 day), based on a combination of volatile and particulate emissions.

Many previous risk analyses have used the maximum point of exposure at some prespecified distance from the WMU as the point for analysis. Such an approach is usually criticized as being overly conservative because it does not consider the possibility of no one living at that exact point. Because individuals may potentially be located in any direction and at various distances from a facility, this analysis developed an explicit way to incorporate this consideration. First, a sensitivity analysis was conducted to determine a reasonable distance at which to bound the analysis. This sensitivity analysis showed that, beyond 1,000 m, most air

concentrations are a small percentage (less than 10 percent) of the concentration at the point of maximum exposure. Therefore, 1,000 m was used as the outer bound on the distance of receptors included in this analysis. A receptor grid was set up to allow individuals to reside in any of 16 directions and at distances of 25, 50, 75, 150, 500, and 1,000 m from the edge of the unit.

For this analysis, five receptors were included: an adult resident, a child resident with exposure starting between 0 and 3 years old, a child resident with exposure starting between 4 and 10 years old, a child resident with exposure starting between 11 and 18 years old, and an off-site worker. These receptors could be located in any of 16 directions and at distances of 25, 50, 75, 150, 500, and 1,000 m from the edge of the unit. Each distance was evaluated separately and the location of a receptor was allowed to vary among any of the 16 directions. The 16 directions were equally weighted, so there is equal probability of a receptor's being located anywhere around the WMU. For acute and subchronic exposures, receptors were modeled at 25, 50, and 75 m because it was assumed that the greatest possibility of acute exposure would be closest to the site.

3.2 Conducting the Analysis

As discussed earlier, the analysis consists of three main parts: emissions modeling, dispersion modeling, and exposure modeling/risk estimation. Figure 3-2 shows the model framework. Emissions and dispersion modeling was performed first and the results used as inputs to the exposure modeling/risk estimation. In addition, a database containing characterizations of WMUs was used. The goal of the analysis is to backcalculate a waste concentration that will result in a specified risk. Because risk is assumed to be linear with waste concentration under most circumstances, a waste concentration was generated by forwardcalculating a risk associated with a unit concentration in the waste (i.e., 1 mg/kg for land-based units and 1 mg/L for tanks), then scaling the unit concentration using the ratio of target risk to calculated risk. The assumption of linearity is accurate for the dispersion modeling and the exposure and risk modeling. The emissions model is linear for land-based units and tanks without biodegradation. The emissions model for tanks with biodegradation is nonlinear at the concentration where biodegradation shifts from first order to zero order. The results for tanks with biodegradation were backcalculated using first-order emission rates; however, if this result exceeded the concentration at which biodegradation becomes zero order, the result was adjusted to be based on zero-order emission rates. Even when the emissions model is linear, it is possible, using this approach, to backcalculate waste concentrations that exceed the solubility or soil saturation concentration for the chemical. Results that exceed the solubility or soil saturation concentration under neutral conditions are footnoted in the result tables (soil saturation concentration and solubility can vary according to site-specific temperature and pH conditions).

Emissions modeling was performed for all WMUs and all chemicals, assuming a unit concentration of the chemical in the waste (1 mg/kg for land-based units or 1 mg/L for tanks). These emissions were used as inputs to Step 2 of the exposure modeling/risk estimation portion of the model.

Dispersion modeling was performed for 76 representative WMU areas and height combinations and 29 meteorological locations, assuming a unit emission rate of 1 μ g/m²-s. This

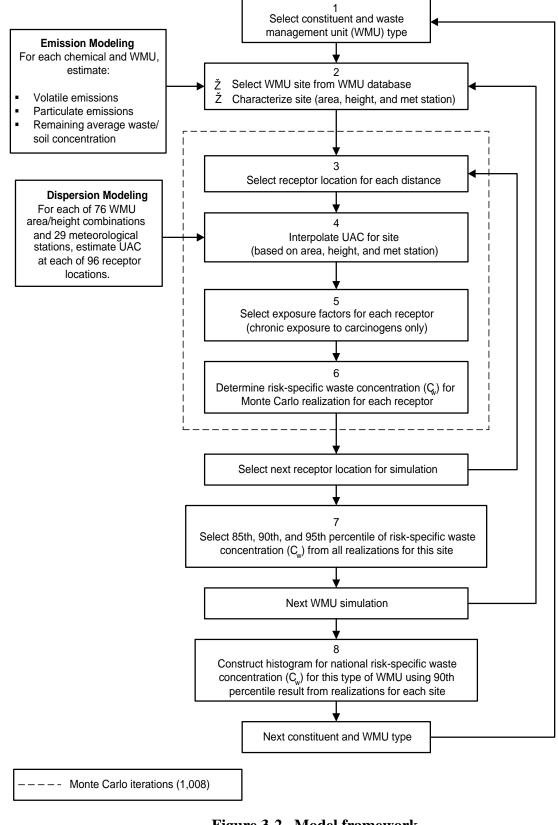


Figure 3-2. Model framework.

produced vapor and particle-phase UACs for each area/height combination, meteorological station, and receptor location, which were used as the basis from which to interpolate in Step 4 of the exposure modeling/risk estimation portion of the model.

The analytical framework shown in Figure 3-2 consists of a series of steps and loops. In Step 1, a chemical and WMU type (e.g., landfills) were selected (thus, all landfills were analyzed as a group for each chemical, and so on).

In Step 2, a WMU was selected from the data file for that unit type. For example, for landfills, the database has a data record containing the facility identification and WMU characteristics such as surface area, depth, and waste quantity managed per year for each of 801 landfill units. The database also has a sampling weight for each facility that defines how many facilities nationally were represented by that facility. An assigned meteorological station was added to the database based on locational information for each WMU. The model simulation starts with the first record and moves to each successive record. For each WMU record, the associated emission rate for that WMU and chemical was obtained from the emission modeling results.

In Step 3, receptor locations were selected by choosing at random one of the 16 directions modeled in the dispersion modeling. Receptors were modeled in that direction at each of six distances from the site.

In Step 4, a UAC was interpolated for the WMU. Due to the long run time of ISCST3 for area sources, UACs were modeled for only 76 selected WMU area/height combinations for each meteorological station and receptor location. To calculate a UAC corresponding to the WMU's actual area and height, EPA first chose the modeled height closest to the actual unit height, then interpolated between the UACs for the two closest of the areas modeled. For example, the first three areas modeled for wastepiles were 20, 162, and 486 m². These were modeled at heights of 1, 2, 4, 6, and 8 m. For a WMU with an actual area of 100 m² and an actual height of 3.5 m, the UAC was interpolated from the UACs for 20 m²/4 m high and 162 m²/4 m high. For a WMU with an actual area of 200 m² and an actual height of 6.9 m, the UAC was interpolated from the UACs for 162 m²/6 m high and 486 m²/6 m high.

In Step 5, for chronic exposures to carcinogens, values of exposure factors such as body weight, inhalation rate, and exposure duration were chosen at random from distributions of these parameters (developed from data in the *Exposure Factors Handbook*, U.S. EPA, 1997c and 1997d) to capture the variability in exposure factors for a given receptor. These exposure factors differ for different receptor types (such as adults, children, and workers). Noncarcinogens were not assessed in this manner because the health benchmarks, such as EPA's reference concentration (RfC), are expressed in terms of ambient concentration and cannot be adjusted for variations in these exposure factors. Similarly, acute and subchronic health benchmarks are expressed as ambient exposure concentrations and cannot be adjusted for variability in exposure factors.

In Step 6, the emission rate, UACs, and, if applicable, the exposure factors, were combined with the health benchmark for the chemical to estimate risk (for chronic exposure to carcinogens) or hazard quotient (for acute and subchronic exposures, and chronic exposures to

noncarcinogens) associated with the unit concentration modeled. This risk was then compared to the target risk of 1 in 1 million, 1 in 100,000, or 1 in 10,000 (i.e., 1E-6, 1E-5, or 1E-4) for carcinogens, and the ratio was used to scale the unit concentration to a concentration in the waste (C_w) that would result in the target risk at that receptor. A similar technique was used for scaling the hazard quotient for noncarcinogens.

Steps 3 through 6, which form the core of the Monte Carlo simulation, were then repeated 1,008 times for each WMU, resulting in a distribution of $C_{\rm w}$ for that WMU for each receptor (adult, child, or worker) at each distance from the site (25, 50, 75, 150, 500, and 1,000 m) for a specific risk criteria (i.e., 1E-4, 1E-5, or 1E-6 for carcinogens and 10, 1, or 0.25 for noncarcinogens). Once 1,008 iterations had been performed for a WMU, various percentiles were selected from the distribution to characterize it. These percentiles represent the percentage of receptors protected at the WMU.

Steps 2 through 6 were then repeated to obtain distributions of C_w for each WMU in the database. These distributions are somewhat different for carcinogens and noncarcinogens and for chronic, subchronic, and acute exposures. For chronic exposure to carcinogens, they represent both the potential variability in location around a WMU, as well as the variability in exposure duration, inhalation rate, and body weight for each receptor type. For noncarcinogens and for subchronic and acute exposures, variability in these exposure factors is not considered because the measure of risk is a ratio of air concentrations. For chronic exposures to noncarcinogens, the distributions represent the variability in location around the WMU at a specific distance. For subchronic and acute exposures, only point estimates were made at various distances using the receptor located at the point of maximum air concentration for that distance.

The cumulative distribution of C_w for each WMU is presented as the percentage of receptors that are at or below the risk criteria for any C_w (see Figure 3-3, left side). For example, 90 percent of all adult residents at a distance of 150 meters have a predicted risk at or below 1 in 100,000 (1E-5) if the concentration of the chemical (e.g., cumene) in the landfill is 1 mg/kg (see point a). A second landfill may have a 90 percent protection level for all adult residents at 150 m at a concentration of 10 mg/kg (point b), and a third landfill at a concentration of 100 mg/kg (point c). Thus, in Step 7, for each WMU, the distribution shows the percent of potential receptors at or below a specified risk level for each concentration of constituent in the WMU (C_w) for each distance and each receptor.

In Step 8, once all WMUs of a certain type had been modeled, the distributions of $C_{\rm w}$ for all individual WMUs of the same type (e.g., landfills) were combined to produce a cumulative distribution that presents the variability in $C_{\rm w}$ across all units of a certain type. For a given percentage of protected receptors (e.g., 90 percent) as described above, the $C_{\rm w}$ was combined across all WMUs of a specified type (e.g., landfills) to provide a distribution of the percentage of sites considered protective at that level, as shown in Figure 3-3 (right side). Figure 3-3, for example, shows the cumulative distribution of $C_{\rm w}$ at a 90 percent protection level across all landfills. From this distribution, the 90th percentile $C_{\rm w}$ value for all 90 percent protection levels across all landfills could be estimated. As described above, three landfills that give a 90 percent protection level (i.e., at 1E-5) for a resident at 150 meters from the unit boundary have corresponding $C_{\rm w}$ values of 1 mg/kg, 10 mg/kg, and 100 mg/kg (see points labeled a, b, and c).

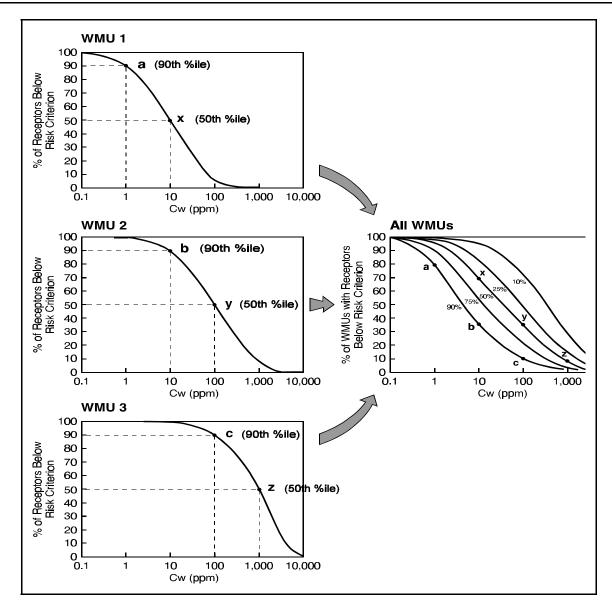


Figure 3-3. Combination of results for individual WMUs into a distribution across all WMUs.

These values plus similar values from all other landfills constitute the cumulative distribution. The $C_{\rm w}$ value that is protective of 90 percent of receptors across 90 percent of the sites is referred to in this study as the 90/90 protection level. These distributions were developed for each unit type, each receptor type, each risk criteria, and each distance from the WMU.

These cumulative distributions are intended to encompass the variability across WMUs. Thus, the variability in WMU characteristics and in meteorological settings is included in these distributions.

This process was repeated from Step 1 for each chemical and WMU type analyzed in this study.

3.2.1 Data Sources

The Industrial D Survey database (Shroeder et al., 1987) was the primary source of data on WMUs used in this analysis. This database provides information on each of the WMUs assessed, with the exception of tanks. Tank data are from EPA's National Survey of Hazaradous Waste Treatment, Storage, Disposal and Recycling Facilities (TSDR Survey, U.S. EPA, 1987). The Industrial D Survey database contains information on the size and capacity of a statistical sample of each WMU type, general location information, and statistical weights for each facility in the sample. The statistical sample was designed to represent all industrial waste management units not regulated under the RCRA hazardous waste program at the time the survey was conducted in 1987. The weights in the database indicate the number of facilities represented by each facility in the sample. For this assessment, it is assumed that the data contained in this database provide an appropriate representation of the characteristics of each WMU type and of the general location of these types of facilities with respect to climate regions of the country.

Meteorological stations provided temperature and windspeed data as inputs to the emissions model and a large set of inputs for the dispersion model. Although meteorological data are available at over 200 meteorological stations in the United States (see, for example, *Support Center for Regulatory Air Models (SCRAM) Bulletin Board* at http://www.epa. gov/scram001), various resource constraints prevented the use of all available data sets in this analysis. Therefore, a set of 29 stations was used that had been selected as representative of the nine general climate regions in the contiguous United States in an assessment for EPA's Superfund Soil Screening Level (SSL) program (EQM, 1993).

In EPA's Superfund study, it was determined that 29 meteorological stations would be a sufficient sample to represent the population of 200 meteorological stations and predict mean dispersion values with a high (95 percent) degree of confidence. The 29 meteorological stations were distributed among the nine climate regions based on meteorological representativeness and variability across each region. Large-scale regional average conditions were used to select the actual stations.

The 29 meteorological stations are listed in Section 5 of Volume II. To assign each Industrial D or TSDR facility to a meteorological station, EPA used a geographic information system (GIS) to construct areas around each station that encompass the areas closest to each station. The boundaries of these areas were then adjusted to ensure that each boundary encloses an area that is most similar in meteorological conditions to those measured at the meteorological station. First, the boundaries were adjusted to correspond to Bailey's ecological divisions (Bailey et al., 1994), which are defined primarily on physiography and climate. The boundaries were further adjusted for coastal (including Great Lakes) areas and the central valley of California to ensure that these stations were used only in regions with similar meteorology. Based on zip codes in the Industrial D Survey database and EPA IDs in the TSDR database, the sites were then overlaid on this GIS coverage, and meteorological station assignments were then exported for use in the modeling exercise. Several sites in Alaska, Hawaii, and Puerto Rico were deleted from the analysis at this point because the 29 meteorological stations are limited to the continental United States.

Figure 3-4 shows the final meteorological station boundaries used for the study along with the zip code centroid locations for the Industrial D sites.

3.2.2 Emissions Modeling

Both volatile emissions (for all WMU types) and particulate emissions due to wind erosion (for land-based WMUs) were included in the risk analysis. To assess these two types of emissions, three parameters had to be modeled: volatile emission rate, long-term average soil concentration in the unit (for LAUs, landfills, and wastepiles), and particulate matter emission rate.

EPA's CHEMDAT8 model was selected as the model to estimate volatile emissions rates and long-term average soil concentrations in the WMU. The CHEMDAT8 model was originally developed in projects funded by EPA's Office of Research and Development (ORD) and Office of Air Quality Planning and Standards (OAQPS) to support National Emission Standards for Hazardous Air Pollutants (NESHAP) from sources such as tanks, surface impoundments, landfills, wastepiles, and land application units for a variety of industry categories including chemical manufacturers, pulp and paper manufacturing, and petroleum refining. It also has been used to support the emissions standards for hazardous waste treatment, storage, and disposal facilities (TSDF) (U.S. EPA, 1991) regulated under Subpart CC rules of RCRA, as amended in 1984. The CHEMDAT8 model is publicly available and has undergone extensive review by both EPA and stakeholder representatives. The CHEMDAT8 spreadsheet model and model documentation may be downloaded at no charge from EPA's web page (http://www.epa.gov/ttn/chief/software.html).

The CHEMDAT8 model considers most of the competing removal pathways that might limit air emissions, including adsorption, hydrolysis (for tanks only), and biodegradation. Adsorption/absorption is the tendency of a chemical or liquid media to attach or bind to the surface or fill the pores of particles in the soil or waste and therefore not volatilize into the air. This tendency to adsorb to or absorb in particles is an important process for estimating the concentration of the chemical on particles emitted to the air due to wind erosion. CHEMDAT8 in its original form models adsorption for land-based units by presuming that the entered waste concentration is in liquid phase. Because waste concentrations are more typically measured as total concentration (liquid plus solid phase), CHEMDAT8 was modified to model adsorption explicitly for an entered total waste concentration for land-based units. Biodegradation is the tendency of a chemical to be broken down or decomposed into less complex chemicals by organisms in the waste or soil. Similarly, hydrolysis is the tendency of a chemical to be broken down or decomposed into less complex chemicals by reaction with water. Chemicals that decompose due to biodegradation or hydrolysis have lower potential for emission to the air as gases or particles than those that do not. Loss of contaminant by leaching or runoff is not included in the CHEMDAT8 model. Both leaching and runoff are a function of a chemical's tendency to become soluble in water and follow the flow of water (e.g., due to rainfall) down through the soil to groundwater (leaching) or downhill to surface water (runoff). These two mechanisms would also make less chemical available for emission to the air as a gas or as particles. As such, CHEMDAT8 is considered to provide reasonable to slightly high (environmentally protective) estimates of air emissions from the land-based units.

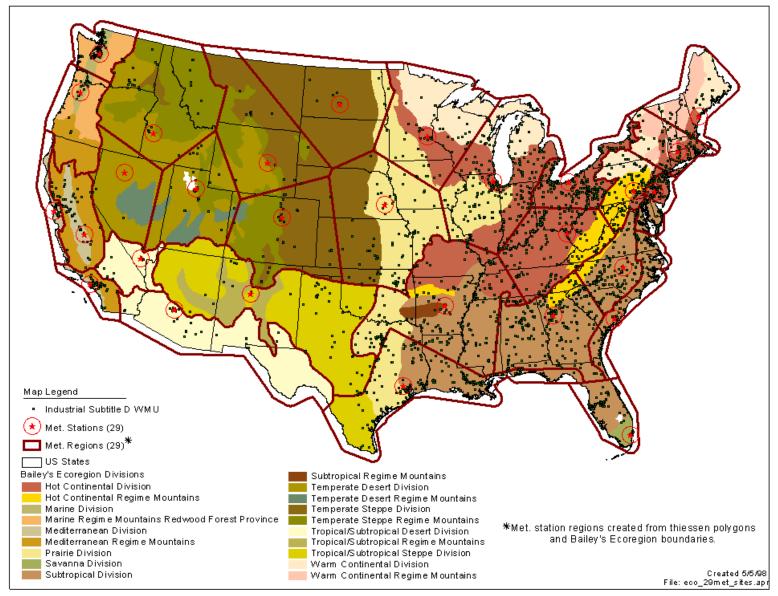


Figure 3-4. Meteorological station regions.

Two different models were used to model wind erosion: one for wastepiles (elevated sources) and one for landfills and land application units (ground-level sources). The Cowherd model (U.S. EPA, 1985b and 1988) was selected for modeling wind erosion emissions from ground-level sources, and the AP-42 model for wind erosion from aggregate storage piles (U.S. EPA, 1985a) was selected for modeling wind erosion emissions from wastepiles. Newer versions of both of these models are available; however, the newer versions are event-based algorithms that require extensive site-specific data that were not available for the sites modeled in this analysis. The versions used probably result in somewhat higher particulate emissions estimates than the event-based algorithms would. This overestimation of particulate emissions is not significant for volatile chemicals, as particulate emissions were found to be a negligible fraction (less than 2 percent in most cases) of total emissions for the volatile chemicals modeled in land-based units. The protective waste concentrations (C_w's) for metals other than mercury (which do not volatilize and are therefore based solely on particulate emissions) may be somewhat lower as a result of this overestimation of emissions.

Both volatile and particulate emissions were estimated for the landfill, land application unit, and wastepile, while only volatile emissions were estimated for tanks.

3.2.2.1 Estimating Volatile Emissions. The modeling scenario and critical parameters required for each type of WMU are provided in the following subsections. A more detailed discussion of the emissions modeling is provided in Volume II.

The input parameters used for the CHEMDAT8 land-based unit emissions model are presented in Table 3-1.² Of these parameters, two are actually flags to determine which model equations to apply: the aqueous waste flag and the biodegradation flag. The most important flag for emission estimates is probably the aqueous waste flag. This flag tells the CHEMDAT8 model which equilibrium partitioning model to use between the liquid and gas phases. For organic wastes, the model uses Raoult's law and the liquid-to-air partition coefficient becomes proportional to the contaminant's partial vapor pressure. For aqueous wastes, the model uses Henry's law and the liquid-to-air partition coefficient becomes proportional to the contaminant's Henry's law coefficient.

All land-based WMUs were modeled twice; once assuming unit concentration (concentration set to 1 mg/kg, assuming Henry's law applies) and once assuming pure component (concentration set to 1E+6 mg/kg, assuming Raoult's law applies). The results presented in Section 4 and in Volume III are based on the aqueous phase emission rates (unit concentration and Henry's law). The pure component emission rates were used only to identify chemicals for which greater emissions occur from the organic phase than from the aqueous phase (which is rare) or to identify chemicals for which the aqueous-based results exceeded soil saturation concentrations or solubility limits at neutral pH and standard temperature and to note for these whether the target risk or hazard quotient would be exceeded modeling pure component.

²The data entry form in the CHEMDAT8 model refers to oil rather than waste; the term waste is used here for clarity.

Table 3-1. CHEMDAT8 Land-Based Unit Model Input Requirements

Input Parameter	Data Source/Assumption
Loading (g waste/cm³ soil)	Waste quantity and/or density from Ind. D Survey
Concentration in waste (ppmw)	1 for unit concentration run; 1E+6 for pure component run
Depth of tilling (or unit) (cm)	Assumed or set by capacity
Total porosity	Assumed default value of 0.5
Air porosity (0 if unknown)	Assumed default value of 0.25
Molecular weight of waste (g/mol)	18 for unit concentration run; 147 for pure component run
Aqueous waste flag:	For aqueous waste, enter 1 For organic waste, enter 0
Time of calculation (days)	Dependent on type of WMU
Biodegradation Flag:	For biodegradation, enter 1 For no biodegradation, enter 0
Temperature (°C)	Set by location of WMU
Windspeed (m/s)	Set by location of WMU
Area (m²)	Input from Ind. D Survey
Fraction organic carbon	Assigned randomly from distribution

Three other parameters are critical for land-based units: the annual waste quantity, the temperature, and the biodegradation rate. The annual waste quantity, along with assumptions regarding the frequency of waste addition and the dimensions of the WMU, combine to influence a number of model input parameters including loading, concentration of contaminant in the waste, depth of the unit (or tilling), operational life, and surface area of the WMU.

Temperature is important because it affects the air diffusivity, which affects the volatilization rate and may affect the biodegradation rate (biodegradation rates were independent of temperature above 5°C and were set to zero below 5°C). Temperature is the only meteorological data input that potentially impacts the emissions results for the CHEMDAT8 model for the land-based WMU. The CHEMDAT8 model is insensitive to windspeeds for long-term emission estimates from land-based units.

The process of biodegradation is important because it lowers both the emission rate and the average soil concentration. Consequently, biodegradation is an important input parameter, and the biodegradation rate constants used in the model are critical parameters. Biodegradation was treated differently for the various WMUs. Landfills are not designed for biodegradation, and waste in wastepiles managed over short periods will not be affected substantially. Therefore, both the landfill emission runs and the short-term wastepile emission runs did not include biodegradation losses. First-order biodegradation was included in the LAU emission runs and long-term wastepile emission runs.

Table 3-2 presents the required CHEMDAT8 input parameters for tanks. Three types of parameters are critical: factors affecting turbulence, waste characteristics affecting biodegradation, and meteorological inputs.

Factors that affect the relative surface area of turbulence and the intensity of that turbulence are important in determining the fate of chemicals in tanks. The tank model has several input parameters that impact the degree and intensity of the turbulence created by the aeration (or mixing). The tank model is most sensitive to the fraction aerated.

Waste characteristics that influence the rate of biodegradation are important in determining emissions from both aerated and storage tanks. As shown in Table 3-2, these parameters include active biomass concentration, total solids in, total organics in, and total biorate. Biodegradation was modeled for aerated tanks reporting biological treatment. Aerated tanks reporting other types of treatment, nonaerated treatment tanks, and storage tanks were modeled with no biodegradation.

Unlike the biodegradation rate model that was used for the land-based units, the biodegradation rate model used in CHEMDAT8 for tanks depends on the amount of active biomass in the WMU. Therefore, the active biomass concentration is a critical parameter for

Table 3-2. CHEMDAT8 Tank Model Input Requirements

Input Parameter	Date Source/Assumption							
	Unit Design							
Flow rate (m ³ /s)	Survey							
Depth (m)	Imputed based on volume							
Average surface area (m ²)	Imputed based on volume and depth							
Height above ground (m)	Imputed based on depth							
A	eration Parameters							
Fraction agitated	Estimated distribution							
Total power (hp)	Imputed based on volume							
Number of impellers	Imputed based on total power							
Impeller diameter (cm)	Estimated constant = 61							
Impeller speed (rad/s)	Estimated constant = 130							
Power efficiency (unitless)	Estimated constant = 0.83							
O ₂ transfer rate (lbO ² /h-HP)	Estimated constant = 3							
Submerged air flow (m³/s)	Estimated constant $= 0$							
W	aste Characteristics							
Active biomass conc. (kg/m³)	Estimated distribution, depends on treatment code							
Total solids in (kg/m³)	Estimated distribution							
Total organics (COD) In (g/m³)	Estimated distribution							
Total biorate (mg/g-h)	Estimated constant = 19							
N	Meteorological Data							
Temp (°C)	Imputed based on meteorological station							
Windspeed (m/s)	Imputed based on meteorological station							

aerated tanks. Because this parameter can vary widely for different types of tanks, biomass concentrations were set on a tank-by-tank basis for aerated tanks using process code information (WMU codes) from the TSDR Survey.

Meteorological inputs are also important for the tank emission model. For nonaerated treatment tanks and storage tanks, the emission estimates are impacted by both temperature and wind speed. Because the emissions for aerated tanks are predominantly driven by the turbulent area and associated mass transfer coefficients, the emissions from the aerated tanks are not strongly impacted by the wind speed. Aerated tank emissions are impacted by temperature. Annual average temperatures were used as input to the model based on tank locations.

The following sections describe the emissions assumptions used for determining volatile emissions for each WMU type.

Landfills. For landfills, annual average emissions were estimated from the active landfill cell assuming the active landfill cell could hold 1 year's worth of waste material. The emissions for the active cell were made assuming that the cell is instantaneously filled and that no waste cover is applied for the first year. Therefore, a full year's worth of waste was available for emissions to the air each year. Once the cell is covered at the end of a year, no additional emissions of gases or particles were modeled from that cell. Because landfills are not constructed for the purpose of biodegrading wastes, as are land application units or biologically active tanks, and because conditions are not controlled to foster biodegradation in landfills, biodegradation was not modeled in landfills.

The annual average emission rate and waste concentration for the active

Assumptions for Modeling Volatile Emissions from Landfills

- # Landfill operates for 20 years filling 20 equal cells sequentially.
- # The active cell is modeled as being instantaneously filled at time t=0 and remains open for 1 year.
- # Emissions are calculated only for one cell for 1 year (after 1 year, cells are either depleted of the constituent or capped).
- # Waste is homogeneous with an initial concentration of 1 mg/kg.
- # The waste matrix may be aqueous (Henry's law partitioning applies) or organic (Raoult's law partitioning applies).
- # Annual average temperature is used (determined by assigned meteorological station).
- # Acute and subchronic exposures were not modeled.

landfill cell were estimated using annual average meteorological data. A sensitivity analysis showed no difference in emissions estimates using seasonal meteorological data (less than 2 percent error for most chemicals). The annual average emission rates were used for chronic risk calculations. Acute and subchronic risks were not considered for landfills.

The average concentration of the waste in the landfill cell was estimated from the emission fraction by assuming first-order contaminant (concentration) disappearance. The details of this calculation are provided in Volume II, Section 4. The relationship between the emission rate and the waste concentration was needed to estimate a concentration in the WMU that corresponded to a specific risk or hazard quotient (HQ) for a receptor.

Land Application Units. For land application units, the volatile emissions were estimated assuming waste additions 24 times per year and first-order biodegradation for temperatures greater than 5°C. The emissions were estimated using monthly average meteorological data. LAU emissions are time-dependent (depending on how recently waste was added) but were modeled as pseudo-steady-state (i.e., steady-state emissions were modeled for a series of short intervals: these estimates were then averaged to produce a long-term emission rate). The average emission rate and associated waste/soil mixture concentrations were estimated for each bimonthly period (i.e., the time between applications). These computations were carried out for 40 years, and the average emission rate and soil/waste concentration for year 40 was used to estimate the longterm annual average emission rates and soil/waste concentrations for each contaminant. The 40-year time period is

Assumptions for Modeling Volatile Emissions from LAUs

- # Waste application occurs 24 times per year.
- # Emissions are modeled as pseudo-steady-state.
- # Emissions in year 40 are used to estimate long-term emissions.
- # Waste is homogeneous with an initial concentration of 1 mg/kg.
- # The waste matrix may be aqueous (Henry's law partitioning applies) or organic (Raoult's law partitioning applies).
- # Monthly average temperature was used for chronic exposure (determined by assigned meteorological station).
- # Maximum or minimum average monthly temperature was used for acute and subchronic exposures (whichever gave higher emissions).
- # Biodegradation occurs at temperatures greater than 5° C.

long enough to result in steady-state emissions for most chemicals and is longer than most of the exposure durations used in the analysis. The annual average emission rates were used for chronic risk calculations. For acute and subchronic risk calculations, emissions were calculated based on the first 24 hours (for acute) or the first 30 days (for subchronic) after waste was added. In the absence of biodegradation, higher temperatures would produce higher volatile emissions. However, when biodegradation is modeled, it slows to zero at temperatures below 5°C, thus increasing volatile emissions at low temperatures. Therefore, for both acute and subchronic exposures, emissions were modeled at the maximum monthly temperature and the minimum monthly temperature, and the one that produced higher emissions was used.

Wastepiles. The wastepile was assumed to remain at a constant volume. Annual waste additions were therefore matched with a corresponding quantity of waste removed. The average residence time of the waste (based on the size of the wastepile and the annual waste quantity) was used to estimate the emission rate and waste concentration across the wastepile. Monthly average emission/waste concentration estimates were made using monthly meteorological data and first-order biodegradation for temperatures greater than 5°C. The resulting monthly average emissions and waste concentrations were then arithmetically averaged to estimate the long-term annual average emission rates and waste concentrations for each contaminant. The annual average emission rates were used for chronic risk calculations. For acute and subchronic risk calculations, emissions were calculated based on the first 24 hours (for acute) or the first 30 days

(for subchronic) after waste was added. Because wastepiles are not designed for biodegradation, there may be a lag time after waste is added before enough acclimated biomass accumulates for biodegradation to begin. This time is typically several months, longer than the 1-day or 30-day periods modeled for acute and subchronic exposures. Therefore, biodegradation was not considered for wastepiles for acute and subchronic exposures. In the absence of biodegradation, higher temperatures produce higher emissions. Therefore, for both acute and subchronic exposures, emissions were modeled at the maximum monthly temperature.

Tanks. For all tanks, the emissions were estimated assuming units were well-mixed and were operating at steady state. The tanks were assumed to have a constant influent and were assumed to operate at a constant temperature. Annual average temperatures and windspeeds were used to estimate the operating conditions for the tanks. A biodegradation rate model using Monod kinetics was used to estimate biodegradation rates for aerated treatment tanks expected to have biodegradation (based on process codes). Biodegradation was not modeled for nonaerated treatment tanks and storage tanks.

Due to the nonlinearity of the biodegradation rate model used in the tank emission estimates, direct backcalculation of an acceptable waste concentration may not be appropriate for some compounds. Unlike the emission

Assumptions for Modeling Volatile Emissions from Wastepiles

- # Wastepile operates with a fixed volume.
- # Waste is homogeneous with an initial concentration of 1 mg/kg.
- # The waste matrix may be aqueous (Henry's Law partitioning applies) or organic (Raoult's law partitioning applies).
- No specific operating life was assumed for wastepiles.
 Residence time of waste in the pile was unit specific.
- # Monthly average temperature was used for chronic exposure (determined by assigned meteorological station).
- # Maximum monthly temperature was used for acute and subchronic exposures.
- # Biodegradation occurs at temperatures greater than 5°C for chronic exposures.
- # No biodegradation was assumed for acute and subchronic exposures.

Assumptions for Modeling Volatile Emissions from Tanks

- # Tanks operate at steady state.
- # Tank is well mixed.
- # Waste has an influent concentration of 1 mg/L.
- # The waste matrix may be aqueous (Henry's Law partitioning applies) or organic (Raoult's law partitioning applies).
- # Annual average temperature was used for chronic exposure (determined by assigned meteorological stations).
- # Operating life is not an explicit input; assumed to be long enough to reach steady state.
- # Biodegradation rate is first order with respect to biomass concentrations.
- # Biodegradation rate follows Monod kinetics with respect to contaminant concentrations.
- # Hydrolysis rate is first order with respect to contaminant concentrations.
- # Acute and subchronic exposures were not modeled.

results from the land-based units, the contaminant concentration used in the analysis may impact the predicted "normalized" emission rate (i.e., the emission rate in g/m²-s per mg/L of contaminant). Therefore, the tanks with biodegradation were run at a low concentration (i.e., 0.001 mg/L) and at a high concentration (i.e., the constituent's solubility). The most appropriate backcalculated emission value was then selected based on the concentration range of the backcalculated values and the constituent's biodegradation characteristics (see Volume II, Section 7.9, for further details).

3.2.2.2 Development of Particulate Emissions. Particulate emissions due to wind erosion were modeled for land-based units (landfills, land application units, and wastepiles). Particulate emissions from truck movement and other activities at the WMUs were not modeled. These activities are likely to result in short bursts of particulate emissions and should be modeled using an event-based emissions model. Such models require more site-specific information than was available for the sites modeled in this analysis.

Landfills and LAUs were modeled differently than wastepiles because they are ground-level sources and wastepiles are elevated sources. For both types of WMU, the models described in this section and in Volume II predict the emission rate of particulate matter released from a site due to wind erosion. To obtain the emission rate of constituent sorbed to particulate matter, the emission rate of particulate matter must be multiplied by the soil or waste concentration.

Landfills and Land Application Units. Wind erosion emissions from landfills and LAUs were modeled using the Cowherd model (U.S. EPA, 1985b). A newer version of Cowherd's model is available in U.S. EPA (1988).

Inputs and Intermediate Values Used for Wind Erosion from Landfills and LAUs											
Symbol	Parameter	Units	Value	Source							
V	Vegetative cover	fraction	0	Assumption							
z_0	Roughness height	cm	1	U.S.EPA (1985b)							
\mathbf{U}^*	Threshold friction velocity	m/s	0.5	Assumed unlimited reservoir							

However, the newer version is an event-based model that requires detailed site-specific information unavailable for this analysis. Therefore, it was not used. The older Cowherd model tends to slightly overestimate emissions relative to the event-based version. Although the degree to which it overestimates is not known, it is expected to be relatively small. Because particulate emissions are negligible compared to volatile emissions for the volatile chemicals modeled, this is only of concern for the metals (other than mercury), which are based only on particulate emissions.

The Cowherd model estimates the emission of respirable particles (i.e., PM₁₀) due to wind erosion from a ground-level surface with an unlimited reservoir of erodible particles. Surfaces are defined as having a limited or unlimited reservoir based on threshold friction velocity (U*). Surfaces with a U* greater than 0.5 m/s are considered limited; those with U* less than 0.5 m/s are considered unlimited (U.S. EPA, 1988). Threshold friction velocity is a measure of the windspeed at the ground surface that would be required to remove particles from the surface. Examples of limited reservoirs include nonhomogeneous surfaces with stones, clumps of

vegetation, or other nonerodible elements or crusted surfaces. Further, wind erosion is considered unlikely to occur from surfaces with full vegetative cover.

A detailed explanation of inputs used in the calculation of particulate emissions is presented in Section 4 of Volume II including vegetative cover, roughness height, average annual windspeed, and threshold friction velocity.

Wastepiles. Wind erosion emissions from wastepiles were modeled using an equation from AP-42 (U.S. EPA, 1985a) for estimating emissions from wind erosion from active storage piles. The equation gives emissions of total suspended particulates (TSP). Typically, an equation-specific particle size multiplier is applied to reduce the emissions to a desired size category, in this case, PM₁₀. No particle size multipliers are given for this equation in AP-42; however, Cowherd (U.S. EPA, 1988) gives a particle size multiplier of 0.5 for use with this equation, and this was used.

Important input parameters for this calculation include silt content of waste (i.e., percent with small particle size), number of days with greater than 0.01 inches of rainfall, and percent of time that windspeed exceeds 5.4 m/s. Data on the silt content of the wastes being modeled were not available. A median silt content of 12 percent was used based on "miscellaneous fill material" from AP-42 (U.S. EPA, 1985a). The number of precipitation days and the frequency of windspeed greater than 5.4 m/s were location-specific; values were obtained from NOAA (1992) and are summarized in Section 4 of Volume II.

3.2.3 Dispersion Modeling

Dispersion modeling was used to estimate air concentrations to which the various human receptors were exposed. A dispersion model (ISCST3) was run to calculate air concentrations associated with a standardized unit emission rate (1 $\mu g/m^2$ -s) to obtain a unitized air concentration (UAC), also called a dispersion factor, which is measured in $\mu g/m^3$ per $\mu g/m^2$ -s. Total air concentration estimates are then developed by multiplying the constituent-specific emission rates derived from CHEMDAT8 by this dispersion factor.

Running ISCST3 to develop a dispersion factor for each of the approximately 3,400 individual WMUs modeled in this study would have been very time consuming due to the run time of the area source algorithm in ISCST3. In

Assumptions Made for Dispersion Modeling

- # Dry depletion was activated in the dispersion modeling for particulates. Depletion was not considered for vapors.
- # An area source was modeled for all WMUs.
- # To minimize error due to site orientation, a square area source with sides parallel to X- and Y-axes was modeled.
- # Modeling was conducted using a unit emission rate of 1 μ g/s-m².
- # Receptor points were placed on 25, 50, 75, 150, 500, and 1,000 m receptor squares starting from the edge of the source with 16 receptor points on each square.
- # The rural option was used in the dispersion modeling since the types of WMUs being assessed are typically in nonurban areas.
- # Flat terrain was assumed.

addition, modeling for many different locations requires extensive preprocessing to generate the detailed meteorological data needed for each location modeled. Therefore, a database of

dispersion factors was developed by running ISCST3 for many separate scenarios designed to cover a broad range of unit characteristics, including:

- # both ground-level and elevated sources
- # 14 surface area sizes for landfills and land application units, 29 surface area—height combinations for waste piles, and 33 surface area-height combinations for tanks
- # 6 receptor distances from the unit (25, 50, 75, 150, 500, and 1,000 meters) placed in 16 directions in relation to the edge of the unit.

Based on the size and location of a specific unit, an appropriate dispersion factor was interpolated from the database of dispersion factors using the closest location and the two closest unit sizes.

In addition, WMUs were assigned to and dispersion modeling was performed for 29 meteorological stations. These were chosen from the more than 200 available to represent the nine general climate regions of the continental U.S.

Each UAC in the database is specific to one meteorological station, one area-height combination, one distance from the unit, and one direction from the unit.

3.2.3.1 Model Selection

A number of dispersion models are available through the EPA Support Center for Regulatory Air Models (SCRAM) Bulletin Board (http://www.epa.gov/scram001/). These dispersion models were developed for a variety of applications, and each has its own strengths and weaknesses. This analysis required a model with the capability to model (1) area sources; (2) ground-level and elevated sources; (3) off-site impacts; (4) vapors and particulates; and (5) annual, monthly, and daily averaging times.

ISCST3 (U.S. EPA, 1995) was selected for all aspects of this analysis because it met all the criteria. This model, however, requires considerable run time, which limited the number of meteorological stations included in this analysis.

3.2.3.2 <u>Meteorological Stations</u>. As stated in Section 3.2.1, due to the considerable run time of ISCST3 for area sources, a set of 29 meteorological stations selected in an assessment for EPA's Superfund Soil Screening Level (SSL) program (EQM, 1993) as being representative of the nine general climate regions of the continental United States was used.

The dispersion modeling was conducted using 5 years of representative meteorological data from each of the 29 meteorological stations. Five-year wind roses representing the frequency of wind directions and windspeeds for the 29 meteorological stations were analyzed. These show that the 29 meteorological stations represent a variety of wind patterns. The wind roses are presented in Appendix C of Volume II.

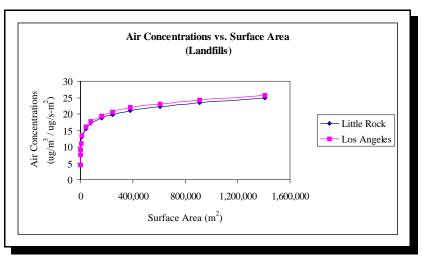
Wind direction and windspeed are typically the most important factors for dispersion modeling analysis. Wind direction determines the direction of the greatest impacts. Windspeed is inversely proportional to ground-level air concentrations, so that the lower the windspeed, the higher the air concentration. Mixing height and stability class are other meteorological conditions that influence dispersion. Mixing height determines the heights to which pollutants can be diffused

Shape of Wind Rose for 29 Meteorological Stations								
Shape of Wind Rose	No. of Stations							
Narrowly distributed	10							
Moderately distributed	4							
Evenly distributed	6							
Bimodally distributed	9							

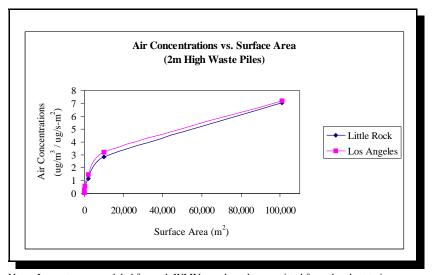
vertically. Stability class is also an important factor in determining the rate of lateral and vertical diffusion. The more unstable the air, the greater the diffusion.

3.2.3.3 Source

Release Parameters. In the modeling analysis, four types of WMUs were considered (landfill, land application unit, wastepile, and tank). Because ISCST3 is sensitive to the size of the area source, the relationship between air concentrations and size of the area source was analyzed. The results show that, for relatively small area sources, air concentrations increase significantly as the size of the area source increases (see boxes). For large area sources, this increase in air concentrations is not as significant. To address this model sensitivity yet avoid modeling approximately 3,400 separate WMUs, EPA developed area strata that represented the distribution of the surface area for each of the WMU types. The surface areas were then used in the dispersion modeling to provide UACs for each of the surface areas for use in the analysis. For elevated



Note: Largest areas modeled for each WMU type have been omitted from the chart to improve clarity.



Note: Largest areas modeled for each WMU type have been omitted from the chart to improve clarity.

sources, area-height combinations were modeled that best covered the range of area-height combinations found in the database. For any specific WMU, a UAC was estimated using an interpolation routine that used the UACs immediately above and below the actual area of the unit. For elevated sources, the UACs associated with the modeled height closest to the actual WMU height were used. The interpolation routine provides a technique for minimizing the number of ISCST3 runs required while also minimizing the error associated with the difference between the UACs for preselected areas and the UAC for the actual area of the WMU.

Landfills and LAUs were modeled as ground-level area sources while wastepiles and tanks were treated as elevated area sources. Fourteen surface areas were selected for modeling for landfills and LAUs. Twenty-nine surface area-height combinations were selected for wastepiles and 33 area-height combinations for tanks. The areas were selected using a modified version of a statistical method called the Dalenius Hodges procedure (see Appendix A, Volume II for details). This procedure divided into strata the skewed distribution of areas found in the Industrial D survey database so that all WMUs in the database would be adequately represented. The median area in each stratum was then used in the dispersion modeling. This procedure is described in more detail in Section 3 of Volume II.

The selected area-height combinations for the four types of WMUs were modeled with 29 representative meteorological locations in the continental United States to estimate UACs. The 5-year average UACs at all receptor points were calculated for the long-term or chronic exposure scenario. They were used as inputs to the Monte Carlo analysis and as input to the interpolation routine discussed above.

A similar methodology and assumptions were used to model dispersion for acute and subchronic exposures. Since the ISCST3 model uses hourly meteorological data, the outputs from the model can be used to develop any averaging times equal to or greater than 1 hour. One set of ISCST3 runs (for all area-height combinations and 29 meteorological stations) was performed for both acute and subchronic analyses, resulting in 5 years of hourly average concentrations at each receptor. For each area, meteorological location, and receptor location, the maximum air concentration for any 24-hour period over the 5 years was selected for acute exposures. Then, for each area and meteorological station, the maximum 24-hour air concentration among all receptor locations at each distance modeled was selected, and this was used as the UAC for that area and meteorological station for acute exposure. The same method was used to determine the subchronic UAC, except that the maximum 30-day period over the 5 years was used instead of the maximum 24-hour period. It was assumed that the greatest potential for acute exposure would be closest to the site, therefore, the receptors points were placed at 25, 50, and 75 meters from the edge of the WMU, in 16 directions at each distance.

3.2.4 Exposure Modeling/Risk Estimation

The previous sections described how emissions and UACs were developed. This section describes the models used to combine those results with exposure factor distributions to calculate risk or hazard quotient and risk-specific waste concentration (C_w).

For carcinogens, a Monte Carlo analysis was performed in which the location of the receptor and various exposure factors (body weight, inhalation rate, and exposure duration) were

varied. For each constituent and WMU type combination, a separate Monte Carlo simulation was run for each WMU in the Industrial D Survey or TSDR Survey database. The emission rate for the specific constituent from the specific WMU was used as an input to the Monte Carlo simulation and was not varied across iterations within a simulation. Approximately 1,000 iterations were performed for each WMU, resulting in a distribution of waste concentrations (C_w) that would result in the specified risk criteria. This distribution captures the range in waste concentration attributable to the variability in potential location and in exposure factors associated with each receptor. From this distribution, the 85th, 90th, and 95th percentiles were selected to characterize the distribution. These percentiles represent the percentage of receptors that are protected at the risk criterion for a specific WMU.

When the Monte Carlo simulation had been run for all the WMUs, a cumulative distribution across all facilities for each protection level (85th, 90th, or 95th percent of receptors) was obtained for each receptor at each distance. This distribution reflects the variability across facilities. In developing this distribution, the results were weighted using the facility weights from the Industrial D Survey data. These weights indicate the number of facilities in the United States represented by a particular facility in the Industrial D Survey database. The resulting cumulative distribution accounts for variability across all facilities represented, not just those actually modeled. The TSDR survey data used to characterize tanks did not include facility weights; therefore, the tank distributions are not weighted.

Hazard quotients for noncarcinogens depend only on air concentration and the health benchmark (a reference concentration). Therefore, exposure factors are not used and of the variables varied in the Monte Carlo analysis, only the location of the receptor is relevant. Because the location of the receptor is such a simple distribution, a Monte Carlo analysis was unnecessary for noncarcinogens; the distribution of hazard quotient (and therefore C_w) based on the distribution of the location of the receptor can be obtained analytically by calculating hazard quotient (and C_w) for each of the 16 receptor locations (or directions around the site) and taking the desired percentiles from those 16 values.

Exposure and risk modeling for subchronic and acute exposures differed somewhat from the modeling for chronic exposures in several respects. All acute and subchronic health benchmarks are analagous to chronic noncarcinogen benchmarks, so exposure factors were not used. In addition, receptor location around the site was not varied for acute and subchronic exposures. Therefore, a Monte Carlo analysis was not performed for subchronic and acute exposures. Instead, a point estimate of C_w was calculated for each WMU in the Industrial D database using the single sector that resulted in the maximum air concentration. This point estimate represents the maximum, or 100^{th} percentile, concentration and therefore is most comparable to the 100^{th} percentile of the distribution generated by the Monte Carlo model for chronic exposures. The point estimate can be interpreted as the level at which 100 percent of receptors are protected at a particular WMU. A distribution across all WMUs of a specific type was generated from these point estimates, and the 90^{th} percentile of that distribution is presented in the results for subchronic and acute exposures.

3.2.4.1 Obtain Health Benchmarks. For chronic exposures, standard health benchmarks (cancer slope factors for carcinogens and reference concentrations for noncarcinogens) were obtained for each constituent (these are shown in Table 3-3). Chronic

benchmarks for 15 chemicals were developed explicitly for the Air Characteristic Study. However, a benchmark could not be developed for one chemical (3,4-dimethylphenol) due to lack of appropriate data, so, although EPA has addressed this chemical, risks for it could not be quantified.

Information on acute and intermediate/subchronic inhalation benchmark values and occupational exposure limits was collected for use in the analysis. These data are also shown in Table 3-3.

3.2.4.2 <u>Calculate Risk or Hazard Quotient</u>. The risk or hazard quotient associated with a unit waste concentration was calculated for each iteration based on the calculated air concentration and the exposure factors selected for the iteration.

Carcinogens. Adult receptors modeled include adult residents and off-site workers. Risk for adults is calculated using long-term average air concentration that is constant over the entire exposure duration. The inhalation rate, exposure frequency, and exposure duration differ for residents and workers. Body weight is the same for all adults, whether resident or worker. All exposure factors for adults are held constant over the entire exposure duration.

Three child age groups, or cohorts, were used to model child exposures: 0 to 3, 4 to 10, and 11 to 18 years of age. These cohorts reflect

Calculation of Risk for Carcinogens for Adults

$$Risk_{calc'd} = \frac{C_{air} * CSF * IR * ED * EF}{BW * AT * 365 d/yr}$$

where

Risk_{calc'd} = individual risk associated with unit waste

concentration (per mg/kg)

 C_{air} = air concentration associated with a unit

waste concentration ([mg/m³]/[mg/kg])

CSF = cancer slope factor (per mg/kg-d)

IR = inhalation rate (m³/d) ED = exposure duration (yr) EF = exposure frequency (d/yr)

BW = body weight (kg)

AT = averaging time (yr) = 70

the age cohorts for which inhalation rate data are available. Results were calculated and saved for receptors falling into each of these three age cohorts at the start of exposure and presented as child 1 (0-3), child 2 (4-10), and child 3 (11-18). An exposure duration was selected randomly for each of the three starting age cohorts from a distribution specific to each starting age cohort. For each age cohort, exposure begins at a starting age selected at random within the cohort and then continues through succeeding age cohorts and into adulthood as necessary until the exposure duration selected for that starting age cohort is reached.

Annual risk for each year of exposure (from starting age to starting age plus exposure duration) was calculated and summed over the exposure duration for each child receptor. If the child reached age 19 before the exposure duration ended, adult exposure factors were used for the remainder of the exposure duration. This approach requires both body weight and inhalation rate distributions by year of age; however, only body weight is available by year. Inhalation rate is available only for the age groups used to define the cohorts (0-3, 4-10, and 11-18 years). Because inhalation rate data could not be disaggregated to individual years of age, we retained

Section 3.

Table 3-3. Inhalation Health Benchmarks Used in the Air Characteristic Analysis

CAS#	Name	Chronic RfC (mg/m³)	Refa	Inhal URF (µg/m³)-1	Inhal CSF (mg/kg/day) ⁻¹	Refa	Subchronic RfC (mg/m³)	Refª	Acute RfC (mg/m³)	Refa
75-07-0	Acetaldehyde	9.0E-03	I	2.2E-06	7.7E-03	ı	9.0E-02	С		
67-64-1	Acetone	3.1E+01	Α	NA	NA		3.1E+01	Α	6.2E+01	Α
75-05-8	Acetonitrile	6.0E-02	I	NA	NA		6.0E-02	I		
107-02-8	Acrolein	2.0E-05	I	NA	NA		2.0E-04	С	1.1E-04	Α
79-06-1	Acrylamide	NA		1.3E-03	4.6E+00	I				
79-10-7	Acrylic acid	1.0E-03	I	NA	NA		3.0E-03	Н	6.0E+00	CA
107-13-1	Acrylonitrile	2.0E-03	I	6.8E-05	2.4E-01	I	2.0E-02	С	2.2E-01	Α
107-05-1	Allyl chloride	1.0E-03	I	NA	NA		1.0E-02	Н		
62-53-3	Aniline	1.0E-03	I	NA	NA		1.0E-02	Н		
7440-38-2	Arsenic	NA		4.3E-03	1.5E+01	I			4.0E-04	CA
7440-39-3	Barium	5.0E-04	Н	NA	NA		5.0E-03	Н		
71-43-2	Benzene	NA		8.3E-06	2.9E-02	I	1.3E-02	Α	1.6E-01	Α
92-87-5	Benzidine	NA		6.7E-02	2.3E+02	I				
50-32-8	Benzo(a)pyrene	NA		8.8E-04	3.1E+00	Ν				
7440-41-7	Beryllium	2.0E-05	I	2.4E-03	8.4E+00	I				
75-27-4	Bromodichloromethane	NA		1.8E-05	6.2E-02	D				
75-25-2	Bromoform (Tribromomethane)	NA		1.1E-06	3.9E-03	I				
106-99-0	Butadiene, 1,3-	NA		2.8E-04	1.8E+00	I				
7440-43-9	Cadmium	NA		1.8E-03	6.3E+00	I				
75-15-0	Carbon disulfide	7.0E-01	I	NA	NA		7.0E-01	Н	2.0E+01	CA
56-23-5	Carbon tetrachloride	NA		1.5E-05	5.3E-02	1	3.1E-01	Α	1.3E+00	Α
126-99-8	Chloro-1,3-butadiene, 2- (Chloroprene)	7.0E-03	Н	NA	NA		7.0E-02	Н		
108-90-7	Chlorobenzene	2.0E-02	Н	NA	NA		2.0E-01	SF		

Table 3-3. (continued)

CAS#	Name	Chronic RfC (mg/m³)	Refa	Inhal URF (µg/m³) ⁻¹	Inhal CSF (mg/kg/day) ⁻¹	Refa	Subchronic RfC (mg/m³)	Refa	Acute RfC (mg/m³)	Refa
124-48-1	Chlorodibromomethane	NA		2.4E-05	8.4E-02	D				
67-66-3	Chloroform	NA		2.3E-05	8.1E-02	I	2.4E-01	Α	4.9E-01	Α
95-57-8	Chlorophenol, 2-	1.4E-03	D	NA	NA					
7440-47-3	Chromium VI	1.0E-04	I	1.2E-02	4.2E+01	I	5.0E-04	Α		
7440-48-4	Cobalt	1.0E-05	D	NA	NA		3.0E-05	Α		
1319-77-3	Cresols (total)	4.0E-04	D	NA	NA		1.2E-03	С		
98-82-8	Cumene	4.0E-01	I	NA	NA		4.0E+00	С		
108-93-0	Cyclohexanol	2.0E-05	FR	NA	NA		2.0E-04	С		
96-12-8	Dibromo-3-chloropropane, 1,2-	2.0E-04	I	6.9E-07	2.4E-03	Н	2.0E-03	С		
95-50-1	Dichlorobenzene, 1,2-	2.0E-01	Н	NA	NA		2.0E+00	Н		
106-46-7	Dichlorobenzene, 1,4-	8.0E-01	I	NA	NA		2.5E+00	Н	4.8E+00	Α
75-71-8	Dichlorodifluoromethane	2.0E-01	Н	NA	NA		2.0E+00	Н		
107-06-2	Dichloroethane, 1,2-	NA		2.6E-05	9.1E-02	I	8.1E-01	С	8.1E-01	Α
75-35-4	Dichloroethylene, 1,1-	NA		5.0E-05	1.8E-01	I	7.9E-02	Α		
78-87-5	Dichloropropane, 1,2-	4.0E-03	I	NA	NA		1.3E-02	Н	2.3E-01	Α
10061-01-5	Dichloropropene, cis-1,3-	2.0E-02	I	3.7E-05	1.3E-01	Н	2.0E-02	Н		
10061-02-6	Dichloropropene, trans-1,3-	2.0E-02	I	3.7E-05	1.3E-01	Н	2.0E-02	Н		
57-97-6	Dimethylbenz(a)anthracene, 7,12-	NA		2.4E-02	8.4E+01	D				
68-12-2	Dimethylformamide, N,N-	3.0E-02	I	NA	NA		3.0E-02	Н		
95-65-8	Dimethylphenol, 3,4-	NA		NA	NA					
121-14-2	Dinitrotoluene, 2,4-	NA		1.9E-04	6.8E-01	D				
123-91-1	Dioxane, 1,4-	8.0E-01	D	NA	NA		8.0E+00	С	6.0E+00	CA
122-66-7	Diphenylhydrazine, 1,2-	NA		2.2E-04	7.7E-01	ı				
106-89-8	Epichlorohydrin	1.0E-03	I	1.2E-06	4.2E-03	I	1.0E-02	Н	3.0E+00	CA
106-88-7	Epoxybutane, 1,2-	2.0E-02	1	NA	NA		2.0E-01	С		

(continued)

Table 3-3. (continued)

CAS#	Name	Chronic RfC (mg/m³)	Ref ^a	Inhal URF (µg/m³) ⁻¹	Inhal CSF (mg/kg/day) ⁻¹	Refa	Subchronic RfC (mg/m³)	Refª	Acute RfC (mg/m³)	Refª
111-15-9	Ethoxyethanol acetate, 2-	3.0E-01	D	NA	NA				3.0E-01	CA
110-80-5	Ethoxyethanol, 2-	2.0E-01	I	NA	NA		2.0E+00	Н	9.0E-01	CA
100-41-4	Ethylbenzene	1.0E+00	I	NA	NA		8.7E-01	Α		
106-93-4	Ethylene dibromide	2.0E-04	Н	2.2E-04	7.7E-01	I	2.0E-03	Н		
107-21-1	Ethylene glycol	6.0E-01	D	NA	NA		6.0E+00	С	1.3E+00	Α
75-21-8	Ethylene oxide	NA		1.0E-04	3.5E-01	Н	1.6E-01	Α		
50-00-0	Formaldehyde	NA		1.3E-05	4.6E-02	I	1.2E-02	Α	6.1E-02	Α
98-01-1	Furfural	5.0E-02	Н	NA	NA		5.0E-01	Н		
87-68-3	Hexachloro-1,3-butadiene	NA		2.2E-05	7.7E-02	I				
118-74-1	Hexachlorobenzene	NA		4.6E-04	1.6E+00	I				
77-47-4	Hexachlorocyclopentadiene	7.0E-05	Н	NA	NA		7.0E-04	Н		
67-72-1	Hexachloroethane	NA		4.0E-06	1.4E-02	I	5.8E+01	Α	5.8E+01	Α
110-54-3	Hexane, -	2.0E-01	I	NA	NA		2.0E-01	Н		
78-59-1	Isophorone	1.2E-02	FR	NA	NA		1.2E-01	С		
7439-96-5	Manganese	5.0E-05	I	NA	NA		5.0E-04	С		
7439-97-6	Mercury	3.0E-04	I	NA	NA		3.0E-04	Н	2.0E-03	CA
67-56-1	Methanol	1.3E+01	D	NA	NA				3.0E+01	CA
110-49-6	Methoxyethanol acetate, 2-	3.0E-02	D	NA	NA					
109-86-4	Methoxyethanol, 2-	2.0E-02	I	NA	NA		2.0E-01	Н	2.0E-02	CA
74-83-9	Methyl bromide (Bromomethane)	5.0E-03	I	NA	NA		1.9E-01	Α	1.9E-01	Α
74-87-3	Methyl chloride (Chloromethane)	NA		1.8E-06	6.3E-03	Н	4.1E-01	Α	1.0E+00	Α
78-93-3	Methyl ethyl ketone	1.0E+00	I	NA	NA		1.0E+00	Н	3.0E+01	CA
108-10-1	Methyl isobutyl ketone	8.0E-02	Н	NA	NA		8.0E-01	Н		
80-62-6	Methyl methacrylate	7.0E-01	I	NA	NA		7.0E+00	С		
1634-04-4	Methyl tert-butyl ether	3.0E+00	I	NA	NA		2.1E+00	Α	6.1E+00	Α

Table 3-3. (continued)

CAS#	Name	Chronic RfC (mg/m³)	Refª	Inhal URF (µg/m³)-1	Inhal CSF (mg/kg/day) ⁻¹	Refª	Subchronic RfC (mg/m³)	Refª	Acute RfC (mg/m³)	Refª
56-49-5	Methylcholanthrene, 3-	NA		2.1E-03	7.4E+00	D				
75-09-2	Methylene chloride	3.0E+00	Н	4.7E-07	1.6E-03	I	3.0E+00	Н	1.0E+01	Α
91-20-3	Naphthalene	3.0E-03	I	NA	NA		3.0E-02	С		
7440-02-0	Nickel	NA		2.4E-04	8.4E-01	1	2.0E-03	С	1.0E-02	CA
98-95-3	Nitrobenzene	2.0E-03	Н	NA	NA		2.0E-02	Н		
79-46-9	Nitropropane, 2-	2.0E-02	I	2.7E-03	9.4E+00	Н	2.0E-02	Н		
55-18-5	Nitrosodiethylamine	NA		4.3E-02	1.5E+02	I				
924-16-3	Nitrosodi-n-butylamine	NA		1.6E-03	5.6E+00	I				
930-55-2	N-Nitrosopyrrolidine	NA		6.1E-04	2.1E+00	I				
108-95-2	Phenol	6.0E-03	FR	NA	NA		6.0E-02	С	6.0E+00	CA
85-44-9	Phthalic anhydride	1.2E-01	Н	NA	NA		1.2E-01	Н		
75-56-9	Propylene oxide	3.0E-02	I	3.7E-06	1.3E-02	I	3.0E-02	Н	6.0E+00	CA
110-86-1	Pyridine	7.0E-03	0	NA	NA					
100-42-5	Styrene	1.0E+00	I	NA	NA		3.0E+00	Н	2.0E+01	CA
1746-01-6	TCDD, 2,3,7,8-	NA		3.3E+01	1.6E+05	Н				
630-20-6	Tetrachloroethane, 1,1,1,2-	NA		7.4E-06	2.6E-02	I				
79-34-5	Tetrachloroethane, 1,1,2,2-	NA		5.8E-05	2.0E-01	I	2.7E+00	Α		
127-18-4	Tetrachloroethylene	3.0E-01	Α	5.8E-07	2.0E-03	SF	3.0E+00	С	1.4E+00	Α
108-88-3	Toluene	4.0E-01	I	NA	NA		4.0E+00	С	1.5E+01	Α
95-53-4	Toluidine, o-	NA		6.9E-05	2.4E-01	D				
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2-	3.0E+01	Н	NA	NA		3.0E+01	Н		
120-82-1	Trichlorobenzene, 1,2,4-	2.0E-01	Н	NA	NA		2.0E+00	Н		
71-55-6	Trichloroethane, 1,1,1-	1.0E+00	SF	NA	NA		3.8E+00	Α	1.1E+01	Α
79-00-5	Trichloroethane, 1,1,2-	NA		1.6E-05	5.6E-02	I				
79-01-6	Trichloroethylene	NA		1.7E-06	6.0E-03	SF	5.4E-01	Α	1.1E+01	Α

(continued)

Table 3-3. (continued)

CAS#	Name	Chronic RfC (mg/m³)	Refª	Inhal URF (µg/m³)-1	Inhal CSF (mg/kg/day) ⁻¹	Refa	Subchronic RfC (mg/m³)	Refª	Acute RfC (mg/m³)	Refa
75-69-4	Trichlorofluoromethane	7.0E-01	Н	NA	NA		7.0E+00	Н		
121-44-8	Triethylamine	7.0E-03	I	NA	NA		7.0E-02	С		
7440-62-2	Vanadium	7.0E-05	D	NA	NA		7.0E-05	С	7.0E-04	С
108-05-4	Vinyl acetate	2.0E-1	- 1	NA	NA		2.0E-01	Н		
75-01-4	Vinyl chloride	NA		8.4E-05	3.0E-01	Н	7.7E-02	Α	1.3E+00	Α
1330-20-7	Xylenes (total)	4.0E-01	Α	NA	NA		3.0E+00	Α	4.3E+00	Α

CAS = Chemical Abstract Service.

CSF = Cancer slope factor. NA = Not available.

RfC = Reference concentration.

URF = Unit risk factor.

^aReferences:

I = IRIS (U.S. EPA, 1999)

H = HEAST (U.S. EPA, 1997a)

A = Agency for Toxic Substances Disease Registry; minimal risk levels (MRLs) (ATSDR, 1999)

SF = Superfund Risk Issue Paper (U.S. EPA, 1996, U.S. EPA, nd)

FR = 63 FR 64371-402 (U.S. EPA, 1998)

N = NCEA Risk Assessment Issue Paper (U.S. EPA, 1994b)

D = Developed for this study

O = Other source (see Volume II, Sections 6.1.1 and 6.1.2)

C = Calculated from chronic RfC value

CA = Cal EPA 1-h acute inhalation reference exposure levels (RELs) (CalEPA, 1998)

E = Acute exposure guideline level (AEGL) (U.S. EPA, 1997b)

year-by-year body weights and used the inhalation rate for the cohort associated with each year of age for that year. Thus, the inhalation rate is a constant for all ages within an age cohort and changes only when the receptor ages from one cohort to the next. Both EPA and a statistician experienced in working with EFH exposure factor data (L. Myers, RTI, personal communication with Anne Lutes, RTI, March 16, 1998) preferred this approach over the alternative of pooling body weights to the age cohort age ranges because it retains the most detail from the available data without sacrificing statistical rigor.

Noncarcinogens. Because the hazard quotient equation for noncarcinogens does not consider exposure factors, there is no difference in results for different receptors at the same location (e.g., adult resident, child resident, and offsite worker). Therefore, only an adult resident was modeled for noncarcinogens.

Calculation of Hazard Quotient for Noncarcinogens

$$HQ_{calc'd} = \frac{C_{air}}{RfC}$$

where

HQ_{calc'd} = hazard quotient associated with unit waste concentration (per mg/kg)

 C_{air} = air concentration associated with a unit waste concentration ([mg/m³]/[mg/kg])

RfC = reference concentration (mg/m^3)

When a particular constituent had both carcinogenic and noncarcinogenic effects, the carcinogenic risk was used, because it is generally more protective.

3.2.4.3 <u>Risk-Specific Waste Concentration</u>. The final step in each iteration was to backcalculate the risk-specific waste concentration from the risk or hazard quotient corresponding to a unit waste concentration. Because risk is linear with respect to waste concentration in the models used in this analysis, this may be done by a simple ratio technique.

As mentioned, risk is assumed to be linear with waste concentration. The assumption of linearity is accurate for the dispersion modeling and the exposure and risk modeling. However, the emissions model is linear only for land-based units and for tanks with no biodegradation; for tanks with biodegradation, the emissions model is nonlinear with respect to biodegradation. At low concentrations, biodegradation in tanks is first order. However, at concentrations in excess of the half-saturation level, biodegradation becomes zero order. In order to address this, emissions were modeled in the aqueous phase at 0.001 mg/L to capture first-order biodegradation and at the solubility to capture zero-order biodegradation. These emission rates then were normalized to a unit concentration by dividing by 0.001 or the solubility. When the backcalculated waste concentration based on first-order biodegradation exceeded the half-saturation constant, suggesting that biodegradation would be zero order, it was recalculated based on the normalized solubility limit emission rate.

The results for all WMU types presented in Section 4 and Volume III were calculated as described above using aqueous-phase emission rates. Most of the waste streams managed in the types of units modeled are expected to contain constituents in the aqueous, rather than the organic, phase; therefore, this is the most realistic scenario. However, results based on organic-phase emissions are of interest in two circumstances: when organic-phase emissions are higher than aqueous-phase emissions, and when backcalculated results based on aqueous-phase

emissions exceed physical limitations on the aqueous phase, such as the soil saturation concentration or solubility. These are discussed below.

Most chemicals are better able to volatilize from an aqueous medium than from an organic medium; therefore, for most chemicals, the aqueous-phase emission rates are considerably higher than organic-phase emission rates. However, for a few chemicals (most notably formaldehyde), the organicphase emissions are higher than the aqueous-phase emissions and protective waste concentrations based on organicphase emissions would be lower than protective waste concentrations based on aqueous-phase emissions. When this is the case, the results based on aqueousphase emissions are footnoted to indicate this. This does not invalidate the aqueous-phase results, as that is still the most likely waste matrix.

Some of the backcalculated waste concentrations based on aqueous-phase emissions exceed the soil saturation concentration or solubility under standard conditions. These are the theoretical maximum possible aqueous-phase concentration in soil or water, respectively; once this is exceeded, free (organic-phase) product will occur in the soil or wastewater. In tanks, free organic phase product will either sink, yielding aqueous-phase emissions from a

Modifications to Methodology for Lead. Human health risk assessment for lead is unique. Instead of developing an RfC in the traditional manner, all identified sources of lead exposure (including background) are used to predict blood lead (PbB) levels in the exposed individuals. The predicted PbB levels are compared to a target PbB. PbB levels have long been used as an index of body lead burdens and as an indicator of potential health effects.

The Integrated Exposure Uptake Biokinetic Model (IEUBK) (U.S. EPA, 1994a) was developed to predict PbB levels for an individual child or a population of children. The model was specifically designed to evaluate lead exposure in young children (birth to 7 years of age) because this age group is known to be highly sensitive to lead exposure. Therefore, only two receptors were modeled for lead: children aged 0 to 3 years and 3 to 7 years. Adults (including workers) and older children were excluded from the analysis for lead because those age groups are considered less sensitive to lead than 0- to 7-year-olds (and, in fact, the pharmacokinetic relationships in the IEUBK model are only valid for 0- to 7-year-olds).

For this analysis, the IEUBK model was used to identify air concentrations that would result in a less than 5 percent probability of having a PbB level higher than the target PbB. That concentration in air was then used in place of an RfC in the calculations. Because the IEUBK model cannot be run in a backcalculation mode, different air concentrations were modeled until one was found that satisfied the 95 percent protection level desired. A target blood lead level of $10~\mu\text{g/dL}$ was selected because that level has been identified as a level of concern by the Centers for Disease Control (CDC) (U.S. EPA, 1994a).

The IEUBK model inputs are summarized in Volume II. They are inhalation rate, body weight, media concentrations (including soil, indoor dust, water, and food), and indoor air concentration as a percentage of outdoor air concentration.

concentration equal to the solubility, or float on the surface, yielding emissions from the organic phase at a concentration of pure component. The soil saturation concentration and solubility under standard temperature and pH conditions (20-25 C and neutral pH) have been estimated for each chemical in the analysis, but these are somewhat site- and waste-specific values. Therefore, a backcalculated concentration may exceed them in some situations but not in others. See Section 7.10.2 of Volume II for details on how the soil saturation concentration and solubility were estimated. When the backcalculated concentration based on aqueous-phase emissions exceeded the typical soil saturation concentration or solubility calculated for this analysis, the result was footnoted to indicate whether pure component (i.e., a concentration of 10^6 mg/kg or

mg/L) would result in a risk exceeding the target risk when modeled using organic-phase emission rates.

3.3 Analysis of Variability and Uncertainty

The purpose of this section is to discuss the methods that are used in this study to capture variability and uncertainty. Variability and uncertainty are discussed separately because they are fundamentally different.

This discussion describes the treatment of variability in some parameters used to describe human receptors and their behavior. Treatment of variability using a Monte Carlo simulation forms the basis for the risk distributions. Uncertainty necessitates the use of assumptions, default values, and imputation techniques in this study. Table 3-4 presents the major categories of variability and uncertainty and how they have been addressed in this study. The columns in the table show scenario uncertainty, model uncertainty, parameter uncertainty, and parameter variability. The rows present the five main model components in the analysis: source characterization, the emissions model, the dispersion model, the exposure model, and the risk model.

3.3.1 Variability

In conducting a national risk assessment, numerous parameters will vary across the nation. Variability is often used interchangeably with the term "uncertainty," but this is not strictly correct. Variability is tied to variations in physical, chemical, and biological processes and cannot be reduced with additional research or information. Although variability may be known with great certainty (e.g., age distribution of a population

Variability arises from true heterogeneity in characteristics such as body weight differences within a population or differences in contaminant levels in the environment.

Uncertainty represents lack of knowledge about factors, such as the nature of adverse effects from exposure to constituents, which may be reduced with additional research.

may be known and represented by the mean age and its standard deviation), it cannot be eliminated and needs to be treated explicitly in the analysis. Spatial and temporal variability in parameter values used to model exposure and risk account for the distribution of risk in the exposed population.

In planning this analysis, it was important to specifically address as much of the variability as possible, either directly in the Monte Carlo analysis or through disaggregation of discrete parts of the analysis. For example, use of a refined receptor grid accounts for spatial variability in concentrations around an WMU. Variability in WMU characteristics is accounted for using a large database of individual WMUs that represent the range of possible WMU characteristics.

3.3.2 Uncertainty

Uncertainty is a description of the imperfection in knowledge of the true value of a particular parameter. In contrast to variability, uncertainty is reducible by additional information

Table 3-4. Summary of Variability and Uncertainty in the Study

	Scenario Uncertainty	Model Uncertainty	Parameter Uncertainty	Parameter Variability
Source Characterization	Ancillary site operations not addressed (e.g., emissions from truck traffic or unloading operations)		Quality of survey data not addressed (e.g., age and representativeness of data, missing data) Imputation of parameter values not directly surveyed based on statistical inference using data for similar WMUs	Facility-specific location, waste volume, dimensions, engineering design parameters used to address variability in WMU parameters Biodegradtion variations with site-specific factors not addressed
Emissions Model	Variations in operation practices not addressed	Instantaneous release model used for acute and subchronic peak releases Competing release mechanisms (e.g., runoff, erosion, leaching) not addressed	Dependencies of biodegradation, volatility, and temperature addressed through sensitivity analysis and use of seasonal temperature variations Differences in biodegradation rates between soil and aqueous systems not addressed	Facility-specific locations and meteorology used to address variability in WMU parameters
Dispersion Model		Model error increased by about 2% by not using wet deposition/depleti on option Photochemical reactions and degradation not addressed 30-day and 1-day averages used for subchronic and acute exposures, respectively	Sensitivity analyses conducted on a number of parameters including shape and orientation of WMU, meteorologic data, and receptor grid	29 meteorologic stations used to represent climate regions 14 surface areas used to represent distribution of surface area for landfills, LAUs 29 surface areas/ heights combinations used for wastepiles 33 surface areas/height combinations used for tanks
Exposure Model	Indirect exposures not addressed		Sensitivity analysis conducted for receptor grid	16 receptor locations at each distance used in Monte Carlo analysis Exposure factor distributions developed and used in Monte Carlo analysis
Risk Model			Health benchmark uncertainty not addressed (e.g., high to low dose extrpolation, animal to human extrapolation)	Variabilility in individual dose response not addressed

gathering or analysis activities (better data, better models). EPA typically classifies the major areas of uncertainty in risk assessments as scenario uncertainty, model uncertainty, and parameter uncertainty. Scenario uncertainty refers to missing or incomplete information needed to fully define exposure and dose. Model uncertainty is a measure of how well the model simulates reality. Finally, parameter uncertainty is the lack of knowledge regarding the true value of a parameter that is used in the analysis. While some aspects of uncertainty were directly addressed in the analysis, much of the uncertainty associated with this analysis could only be addressed qualitatively.

Sources of scenario uncertainty include the assumptions and modeling decisions that are made to represent an exposure scenario. Because we lack information or resources to define and model actual exposure conditions, uncertainty is introduced into the analysis. Despite the complexity of this analysis, it was necessary to exclude or simplify actual exposure conditions. For example, this analysis only addresses inhalation exposures; indirect exposure pathways were excluded. Professional judgement, often coupled with using the results of sensitivity analysis, is used to decide which parameters to include in describing exposure conditions and behaviors. These judgements are imperfect and uncertainty is introduced.

To reduce model uncertainty, EPA generally selected models that are considered state-of-the-art. Model uncertainty is associated with all models used in all phases of a risk assessment. These include the animal models used as surrogates for testing human carcinogenicity, dose-response models used in extrapolations, and computer models used to predict the fate and transport of chemicals in the environment. Computer models are simplifications of reality, requiring exclusion of some variables that influence predictions but cannot be included in models due either to increased complexity or to a lack of data on a particular parameter. The risk assessor needs to consider the importance of excluded variables on a case-by-case basis because a given variable may be important in some instances and not in others. A similar problem can occur when a model that is applicable under average conditions is used when conditions differ from the average. In addition, choosing the correct model form is often difficult when conflicting theories seem to explain a phenomenon equally well. Modeling uncertainty is not addressed directly in this study but is discussed qualitatively.

Parameter uncertainty occurs when (1) there is a lack of data about the parameters used in the equations, (2) the data that are available are not representative of the particular instance being modeled, or (3) parameter values cannot be measured precisely and/or accurately either because of equipment limitations or because the quantity being measured varies spatially or temporally. Random, or sample errors, are a common source of parameter uncertainty that is especially critical for small sample sizes. More difficult to recognize are nonrandom or systematic errors that result from bias in sampling, experimental design, or choice of assumptions.

4.0 Summary of Risk Assessment Results

This section presents an overview of the results of the risk analysis that evaluated the direct inhalation risks from waste management unit (WMU) emissions. These results present waste concentration levels (C_w 's) that protect 90 percent of receptors at distances of 25, 150, and 1,000 m from the edge of the WMU across 90 percent of the sites (90/90 protection levels) at a risk level of 10^{-5} or an HQ of 1. This subset of the results was selected for presentation purposes only and does not imply that these are the results that would be used for an air characteristic. The detailed results that are summarized here, as well as results for alternative risk levels, additional distances, and additional protection levels, are presented in Volume III, *Results*, on CD-ROM.

4.1 Overview of Results

The most protective (i.e., lowest) $90/90 \, C_w$ values for adults across all WMUs ranged from 0.005 ppm to 1 million ppm across all chemicals modeled. The lowest value, 0.005 ppm, was for 2,3,7,8-TCDD in tanks. It should be noted that this value exceeds TCDD's solubility limit (0.00002 ppm) at a neutral pH and temperature of 25°C. As discussed in Section 3.2.4.3, the solubility limit (or soil saturation concentration for land-based units) is a site-specific value, as it varies with pH and temperature. Due to this uncertainty, all the results in this report are shown with the solubility or soil saturation concentration at neutral pH and a temperature of 20-25°C only for comparison purposes. The value for the next lowest chemical (Nitrosodiethylamine) was 0.1 ppm. Chemicals with a C_w of 1 million ppm did not have any concentration that would meet the specified risk level of 10^{-5} or HQ =1.

Figures 4-1 through 4-6 show the number of chemicals at a receptor distance of 150 m with $C_{\rm w}$ in each order of magnitude range from 0.001 to 1 million ppm for aerated treatment tanks, nonaerated treatment tanks, storage tanks, landfills, LAUs, and wastepiles, respectively. For tanks and landfills, only chronic exposures were modeled, and these results are shown. For LAUs and wastepiles, subchronic and acute results are also shown. Of the 104 chemicals modeled, over half of those less then 1 million ppm fall in the 10 to 10,000 ppm range for tanks and in the 100 to 100,000 ppm range for land-based units. From Figures 4-1 and 4-2 it appears that at least 7 of the 105 chemicals in this study may present a significant potential risk via inhalation at very low concentrations (i.e., <1 ppm) when managed in treatment tanks, and another 28 may be of concern at relatively low concentrations (i.e., < 100 ppm). Figures 4-3 to 4-6 suggest that few chemicals are of concern at low levels (e.g. < 100 ppm) when managed in other WMUs.

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Note that 1 chemical of the original 105—3,4-dimethylphenol—was addressed, but risks could not be quantified because data were inadequate to develop a health benchmark for it.

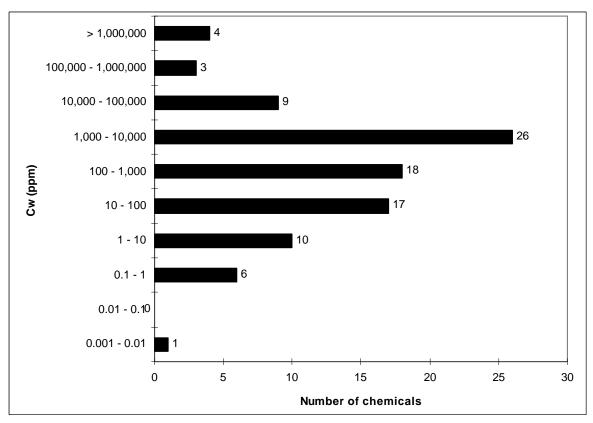


Figure 4-1. Histogram of most protective 90/90 $C_{\rm W}$ for aerated treatment tanks.

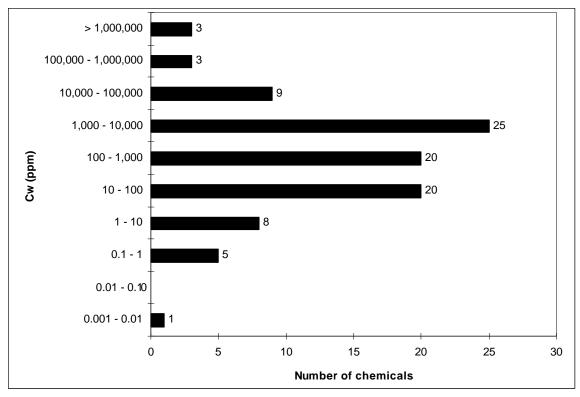


Figure 4-2. Histogram of most protective 90/90 $C_{\rm W}$ for nonaerated treatment tanks.

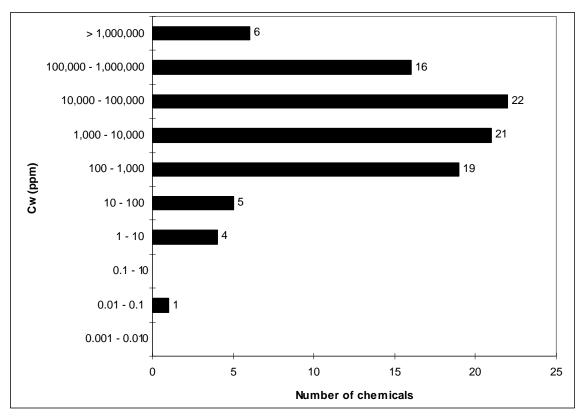


Figure 4-3. Histogram of most protective 90/90 $C_{\rm W}$ for storage tanks.

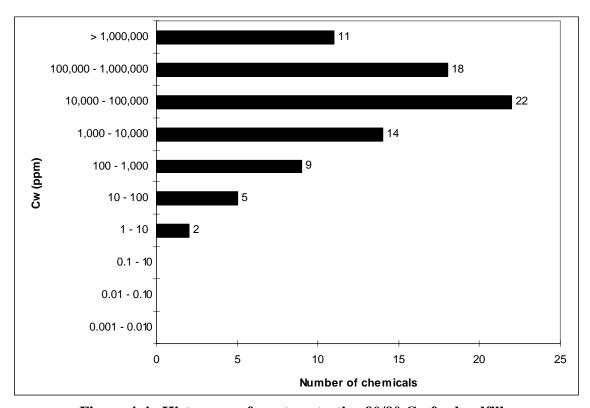


Figure 4-4. Histogram of most protective 90/90 $C_{\rm W}$ for landfills.

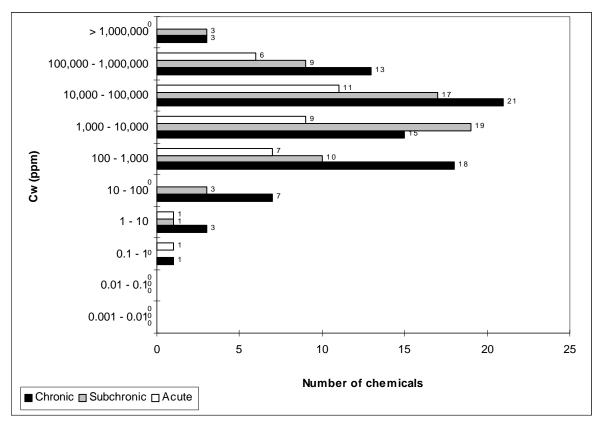


Figure 4-5. Histogram of most protective 90/90 $C_{\rm W}$ for LAUs.

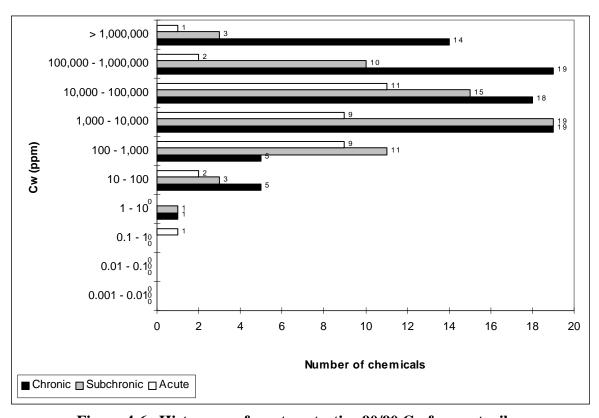


Figure 4-6. Histogram of most protective 90/90 $C_{\rm W}$ for wastepiles.

Tables 4-1 through 4-6 summarize the chronic 90/90 values for all chemicals and receptors for landfills, land application units, wastepiles, aerated treatment tanks, nonaerated treatment tanks, and storage tanks, respectively. These tables show the 90/90 $C_{\rm w}$ for adult residents and for child residents age 0 to 3 years at the start of exposure for chronic exposures at a range of distances. The acute and subchronic results are summarized later in the section and are shown in Tables 4-9 through 4-12. Acute and subchronic exposures were modeled for LAUs and wastepiles to capture the potential peaks in emissions immediately after episodic loading events.

Generally speaking, volatile emissions far exceeded particulate emissions for volatile and semivolatile chemicals, accounting for over 98 percent of total emissions in most cases. Therefore, the particulate emissions are important for metals (except mercury), which do not volatilize, and much less important for the other chemicals included in this study.

The results appear to be closely associated with three primary factors: (1) the physical-chemical properties and toxicity of the chemicals, (2) the WMU type, and (3) exposure factors. The effects of each of these factors are discussed in sections 4.2 through 4.4 with respect to the chronic results. A comparison of chronic, subchronic, and acute results follows in section 4.5.

Table 4-7 presents the physical-chemical properties and toxicity of each chemical. With respect to the health benchmarks shown, note that higher CSFs indicate greater toxicity and lower RfCs reflect greater toxicity.

4.2 Effect of Chemical Properties and Health Benchmarks

The most important factors affecting the results are the physical-chemical properties and toxicity of the chemical. These chemical properties, which include solubility, Henry's law constant (which reflects the tendency to volatilize from water), and biodegradation rate, interact in a complex manner to produce the final result. A chemical must both be emitted in significant quantities and be fairly toxic to result in high risk (and therefore have a low $C_{\rm w}$). A chemical may be highly toxic, but if it is not emitted in significant quantities, it will result in little risk. Because of this, it appears that the

Ten Chemicals with Highest Risk Across All WMU Types

2,3,7,8-TCDD
Nitrosodiethylamine
Nitrosodi-n-butylamine
Acrolein
Nitropropane, 2Butadiene, 1,3Hexachlorocyclopentadiene
Mercury

Vinyl chloride Ethylene Dibromide

physical-chemical properties seem to be more important for chemicals with low risk, while the toxicity seems to be more important for chemicals with high risk.

Chemicals with the highest risk (see box), and so lowest $C_{\rm w}$, appear to be driven by toxicity. These top-10 chemicals are not the most volatile chemicals of the 104; 2 of them, hexachlorocyclopentadiene and 2,3,7,8,-TCDD, are even defined as semivolatiles. Seven of the 10 chemicals are fairly toxic carcinogens with cancer slope factors greater than most of the carcinogens evaluated for this study. The 3 noncarcinogens are also quite toxic and have very low RfCs, among the 10 lowest in this study. The RfCs for these three constituents were equal to or less than 3E-4 mg/m³. It should be noted that 2 of the 10 chemicals with the highest risk have

Table 4-1. Chronic 90/90 $C_{\rm W}$ at 25 m to 1,000 m, Risk = $10^{-5}/{\rm HQ}$ = 1 for Landfills (mg/kg)

			Adult Resid	ent*	Ch	ild Resident 0-3	yr	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg
75-07-0 Acetaldehyde	Vol	4E+03 °	3E+04 °	6E+05 a,c	3E+03 °	3E+04 °	5E+05 a,c	3.3E+0
67-64-1 Acetone	Vol	no risk b,c	no risk b,c	no risk b,c		same as adult		2.3E+0
75-05-8 Acetonitrile	Vol	7E+03	6E+04	no risk b		same as adult		2.3E+0
107-02-8 Acrolein	Vol	1E+00	1E+01	2E+02		same as adult		5.0E+0
79-06-1 Acrylamide	Semi		NM			NM		1.5E+0
79-10-7 Acrylic acid	Semi		NM			NM		2.4E+0
107-13-1 Acrylonitrile	Vol	7E+01	6E+02	1E+04	6E+01	5E+02	1E+04	1.8E+0
107-05-1 Allyl chloride	Vol	1E+01	1E+02	2E+03 ^a		same as adult		1.7E+0
62-53-3 Aniline	Semi		NM			NM		1.0E+0
7440-38-2 Arsenic	Metal	5E+03	4E+04	no risk	4E+03	4E+04	9E+05	1.0E+0
7440-39-3 Barium	Metal	2E+05	no risk	no risk		same as adult		1.0E+0
71-43-2 Benzene	Vol	3E+02	2E+03 ^a	5E+04 ^a	2E+02	2E+03 ^a	4E+04 a	1.8E+0
92-87-5 Benzidine	Semi		NM			NM		2.4E+0
50-32-8 Benzo(a)pyrene	Semi		NM			NM		1.5E+0
7440-41-7 Beryllium	Metal	9E+03	8E+04	no risk ^a	8E+03	7E+04	no risk ^a	1.0E+0
75-27-4 Bromodichloromethane	Vol	4E+02	3E+03	7E+04 ^a	3E+02	3E+03	6E+04 ^a	6.3E+0
75-25-2 Bromoform (Tribromomethane)	Vol	2E+04 ^a	2E+05 b	no risk b	2E+04 ^a	1E+05 b	no risk	4.5E+0
106-99-0 Butadiene, 1,3-	Vol	1E+00	9E+00	2E+02	9E-01	9E+00	2E+02	1.1E+0
7440-43-9 Cadmium	Metal	1E+04	1E+05	no risk	1E+04	9E+04	no risk ^a	1.0E+0
75-15-0 Carbon disulfide	Vol	1E+04 ^a	9E+04 ^b	no risk b		same as adult		1.3E+0
56-23-5 Carbon tetrachloride	Vol	1E+02	9E+02	2E+04 ^a	1E+02	8E+02	2E+04 a	2.7E+0
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	1E+02	1E+03	3E+04 ^a		same as adult		1.8E+0
108-90-7 Chlorobenzene	Vol	2E+03	1E+04 ^a	3E+05 b		same as adult		2.0E+0
124-48-1 Chlorodibromomethane	Vol	5E+02	5E+03 ^a	1E+05 ^b	5E+02	4E+03 ^a	9E+04 b	2.7E+0
67-66-3 Chloroform	Vol	9E+01	8E+02	2E+04 a	8E+01	7E+02	1E+04 ^a	5.7E+0
95-57-8 Chlorophenol, 2-	Semi		NM			NM		2.2E+0
7440-47-3 Chromium VI	Metal	2E+03	2E+04	4E+05	2E+03	1E+04	3E+05	1.0E+0

(continued)

Table 4-1. (continued)

		Adult Resident*		Ch	ild Resident 0-3	yr	_	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg)
7440-48-4 Cobalt	Metal	3E+03	3E+04	6E+05		same as adult		1.0E+06
1319-77-3 Cresols (total)	Semi		NM			NM		5.2E+03
98-82-8 Cumene	Vol	6E+03 ^b	5E+04 ^b	no risk		same as adult		1.9E+03
108-93-0 Cyclohexanol	Vol	6E+00	5E+01	9E+02		same as adult		1.6E+04
96-12-8 Dibromo-3-chloropropane, 1,2-	Vol	5E+04 b	4E+05 b	no risk	4E+04 b	4E+05 b	no risk b	1.8E+03
95-50-1 Dichlorobenzene, 1,2-	Vol	5E+04 ^b	4E+05 ^b	no risk b		same as adult		2.2E+03
106-46-7 Dichlorobenzene, 1,4-	Vol	2E+05 b	no risk b	no risk b		same as adult		1.0E+03
75-71-8 Dichlorodifluoromethane	Vol	2E+03 ^a	2E+04 ^a	3E+05 b		same as adult		1.2E+03
107-06-2 Dichloroethane, 1,2-	Vol	1E+02	1E+03	2E+04 ^a	1E+02	9E+02	2E+04 ^a	3.4E+03
75-35-4 Dichloroethylene, 1,1-	Vol	2E+01	2E+02	4E+03 ^a	2E+01	2E+02	4E+03 ^a	2.7E+03
78-87-5 Dichloropropane, 1,2-	Vol	2E+02	2E+03	3E+04 ^a		same as adult		2.2E+03
10061-01-5 Dichloropropene, cis-1,3-	Vol	1E+02	1E+03	2E+04 ^a	1E+02	9E+02	2E+04 ^a	2.2E+03
10061-02-6 Dichloropropene, trans-1,3-	Vol	1E+02	1E+03	3E+04 ^a	1E+02	1E+03	2E+04 ^a	2.2E+03
57-97-6 Dimethylbenz(a)anthracene, 7,12-	Semi		NM			NM		4.8E+02
121-14-2 Dinitrotoluene, 2,4-	Semi		NM			NM		2.1E+02
123-91-1 Dioxane, 1,4-	Vol	2E+05 °	no risk b,c	no risk b,c		same as adult		2.3E+05
122-66-7 Diphenylhydrazine, 1,2-	Semi		NM			NM		3.3E+02
106-89-8 Epichlorohydrin	Vol	1E+04	8E+04 ^a	no risk b	8E+03	7E+04 ^a	no risk b	1.6E+04
106-88-7 Epoxybutane, 1,2-	Vol	1E+03	1E+04	2E+05 ^a		same as adult		1.7E+04
111-15-9 Ethoxyethanol acetate (R-R), 2-	Vol	1E+05 °	no risk b,c	no risk b,c		same as adult		2.3E+05
110-80-5 Ethoxyethanol, 2-	Vol	3E+05 a,c	no risk b,c	no risk b,c		same as adult		2.3E+05
100-41-4 Ethylbenzene	Vol	9E+04 ^b	7E+05 ^b	no risk b		same as adult		1.3E+03
106-93-4 Ethylene Dibromide	Vol	5E+01	4E+02	9E+03 ^a	4E+01	4E+02	8E+03 ^a	3.1E+03
107-21-1 Ethylene glycol	Semi		NM			NM		2.3E+05
75-21-8 Ethylene oxide	Vol	4E+01 °	4E+02 °	7E+03 °	4E+01 °	3E+02 °	6E+03 °	8.9E+04
50-00-0 Formaldehyde	Vol	5E+03 °	4E+04 °	8E+05 a,c	4E+03 °	4E+04 °	7E+05 a,c	1.3E+05
98-01-1 Furfural	Vol	3E+04 ^a	3E+05 ^a	no risk b		same as adult		2.7E+04
87-68-3 Hexachloro-1,3-butadiene	Semi		NM			NM		1.0E+03
118-74-1 Hexachlorobenzene	Semi		NM			NM		2.3E+04

Section 4.0

Table 4-1. (continued)

		Adult Resident*			Child Resident 0-3 yr			
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg)
77-47-4 Hexachlorocyclopentadiene	Semi		NM			NM		2.1E+03
67-72-1 Hexachloroethane	Vol	4E+04 ^b	3E+05 ^b	no risk	3E+04 ^b	3E+05 b	no risk ^b	2.6E+03
78-59-1 Isophorone	Semi		NM			NM		6.1E+03
7439-92-1 Lead	Metal	6E+05	no risk	no risk	2E+05	no risk ^a	no risk ^a	1.0E+06
7439-96-5 Manganese	Metal	2E+04	1E+05	no risk		same as adult	_	1.0E+06
7439-97-6 Mercury	Metal	2E+02 ^a	1E+03 b	3E+04 b		same as adult		2.6E+01
67-56-1 Methanol	Vol	no risk b,c	no risk b,c	no risk b,c		same as adult		2.3E+05
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	1E+04 °	1E+05 °	no risk b,c		same as adult	_	2.3E+05
109-86-4 Methoxyethanol, 2-	Vol	3E+04 °	2E+05 °	no risk b,c		same as adult		2.3E+05
74-83-9 Methyl bromide (Bromomethane)	Vol	9E+01	8E+02	1E+04 ^a		same as adult		5.7E+03
74-87-3 Methyl chloride (Chloromethane)	Vol	4E+02	4E+03 ^a	7E+04 ^a	4E+02	3E+03 ^a	6E+04 ^a	1.9E+03
78-93-3 Methyl ethyl ketone	Vol	1E+05 ^a	no risk b	no risk b		same as adult		5.4E+04
108-10-1 Methyl isobutyl ketone	Vol	1E+04 ^a	9E+04 ^a	no risk		same as adult		6.0E+03
80-62-6 Methyl methacrylate	Vol	7E+04 ^a	6E+05 ^b	no risk		same as adult		5.5E+03
1634-04-4 Methyl tert-butyl ether	Vol	2E+05 ^a	no risk b	no risk b		same as adult		2.6E+04
56-49-5 Methylcholanthrene, 3-	Semi		NM			NM		4.0E+01
75-09-2 Methylene chloride	Vol	4E+03	3E+04 ^a	6E+05 b	3E+03	3E+04 ^a	6E+05 ^b	4.6E+03
68-12-2 N,N-Dimethylformamide	Semi		NM			NM		2.3E+05
110-54-3 n-Hexane	Vol	2E+04 ^a	1E+05 ^b	no risk		same as adult		6.4E+02
930-55-2 N-Nitrosopyrrolidine	Vol	7E+02 °	6E+03 °	1E+05 °	6E+02 °	5E+03 °	1E+05 °	2.3E+05
91-20-3 Naphthalene	Vol	2E+03 ^a	1E+04 ^b	2E+05 b		same as adult		3.8E+02
7440-02-0 Nickel	Metal	9E+04	8E+05	no risk	8E+04	7E+05	no risk ^a	1.0E+06
98-95-3 Nitrobenzene	Semi		NM			NM		1.3E+03
79-46-9 Nitropropane, 2-	Vol	3E+00	2E+01	4E+02	2E+00	2E+01	4E+02	4.6E+03
924-16-3 Nitrosodi-n-butylamine	Vol	6E+00	5E+01	1E+03	5E+00	4E+01	9E+02	2.1E+03
55-18-5 Nitrosodiethylamine	Vol	6E-01	6E+00	1E+02	6E-01	5E+00	1E+02	2.3E+04
108-95-2 Phenol	Semi		NM			NM		3.3E+04
85-44-9 Phthalic anhydride	Semi		NM			NM		1.4E+03
75-56-9 Propylene oxide	Vol	1E+03 °	1E+04 °	3E+05 a,c	1E+03 °	1E+04 °	2E+05 a,c	1.1E+05

(continued)

6E+04

3E+05 b

1E+03

3.8E+03

3.4E+03

3.3E+03

2.1E+04

1.0E+06

5.3E+03

1.8E+03

1.5E+03

3E+03

1E+04 a

same as adult

same as adult

same as adult

same as adult

6E+01

same as adult

Adult Resident* Child Resident 0-3 yr C_{sat} **CAS Name** Group 25 m 150 m 1000 m 25 m 150 m 1000 m (mg/kg) 4E+05 ^a 110-86-1 Pyridine Vol 3E+03 2E+04 2.6E+05 same as adult 9E+05 b 1E+05 100-42-5 Styrene Vol no risk 1.5E+03 same as adult 1746-01-6 TCDD, 2,3,7,8-Semi NM NM 3.8E-01 7E+03 a 1E+05 ^b 6E+03 ^a 630-20-6 Tetrachloroethane, 1,1,1,2-Vol 8E+02 7E+02 1E+05 2.8E+03 4E+04 4E+04 Vol 2E+02 2E+03 4.7E+03 79-34-5 Tetrachloroethane, 1,1,2,2-2E+02 2E+03 3E+04 b 3E+03 3E+04 6E+05 127-18-4 Tetrachloroethylene Vol 4E+03 7E+05 5.9E+02 2E+05 b 108-88-3 Toluene Vol 2E+04 no risk 1.7E+03 same as adult 95-53-4 Toluidine, o-Semi NM 5.9E+03 NM no risk b 76-13-1 Trichloro-1,2,2-trifluoroethane, 1,1,2-Vol 4E+05 no risk same as adult 2.1E+03 120-82-1 Trichlorobenzene, 1,2,4-Semi NM NM 1.6E+04 2E+05 b 3E+04 a no risk 71-55-6 Trichloroethane, 1,1,1-Vol 2.7E+03 same as adult

3E+03

1E+04 a

9E+04 b

8E+03

2E+05

8E+04

7E+01

3E+05

6E+04

3E+05

no risk

2E+05 a

no risk

1E+03

no risk

3E+02

2E+03

7E+00

Table 4-1. (continued)

NM = Not modeled for land-based units.

79-00-5 Trichloroethane, 1,1,2-

75-69-4 Trichlorofluoromethane

79-01-6 Trichloroethylene

121-44-8 Triethylamine

108-05-4 Vinyl acetate

1330-20-7 Xylenes (total)

75-01-4 Vinvl chloride

7440-62-2 Vanadium

no risk = the aqueous phase result exceeded 1 million parts per million (1,000,000 mg/kg or mg/L).

Vol

Vol

Vol

Vol

Metal

Vol

Vol

Vol

3E+02

2E+03

1E+04 a

1E+03

2E+04

1E+04

8E+00

4E+04

^{*} For lead only, this column contains results for Child Resident 3-7 years.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^c Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

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Table 4-2. Chronic 90/90 C_w at 25 m to 1,000 m, Risk = $10^{-5}/HQ = 1$ for Land Application Units (mg/kg)

		Ad	lult Resident*		Chil	d Resident 0-3 y	r	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg)
75-07-0 Acetaldehyde	Vol	3E+03 °	9E+03 °	1E+05 °	4E+03 °	1E+04 °	1E+05 °	3.3E+05
67-64-1 Acetone	Vol	no risk ^{b,}	no risk b,c	no risk b,c		same as adult		2.3E+05
75-05-8 Acetonitrile	Vol	7E+03	2E+04	2E+05		same as adult		2.3E+05
107-02-8 Acrolein	Vol	1E+00	4E+00	4E+01		same as adult		5.0E+04
79-06-1 Acrylamide	Semi		NM			NM		1.5E+05
79-10-7 Acrylic acid	Semi		NM			NM		2.4E+05
107-13-1 Acrylonitrile	Vol	5E+01	2E+02	2E+03	5E+01	2E+02	2E+03	1.8E+04
107-05-1 Allyl chloride	Vol	1E+01	4E+01	5E+02		same as adult		1.7E+03
62-53-3 Aniline	Semi		NM			NM		1.0E+04
7440-38-2 Arsenic	Metal	9E+02	3E+03	3E+04	1E+03	3E+03	3E+04	1.0E+06
7440-39-3 Barium	Metal	4E+04	1E+05	no risk		same as adult		1.0E+06
71-43-2 Benzene	Vol	1E+02	5E+02	5E+03 ^a	2E+02	6E+02	5E+03 ^a	1.8E+03
92-87-5 Benzidine	Semi		NM			NM		2.4E+02
50-32-8 Benzo(a)pyrene	Semi		NM			NM		1.5E+02
7440-41-7 Beryllium	Metal	2E+03	5E+03	5E+04	2E+03	6E+03	5E+04	1.0E+06
75-27-4 Bromodichloromethane	Vol	1E+02	4E+02	4E+03	1E+02	4E+02	5E+03	6.3E+03
75-25-2 Bromoform (Tribromomethane)	Vol	7E+03 ^a	3E+04 ^a	2E+05 ^b	8E+03 ^a	3E+04 ^a	3E+05 ^b	4.5E+03
106-99-0 Butadiene, 1,3-	Vol	1E+00	4E+00	4E+01	1E+00	4E+00	4E+01	1.1E+03
7440-43-9 Cadmium	Metal	2E+03	7E+03	6E+04	2E+03	8E+03	7E+04	1.0E+06
75-15-0 Carbon disulfide	Vol	9E+03 ^a	3E+04 ^a	3E+05 b		same as adult		1.3E+03
56-23-5 Carbon tetrachloride	Vol	5E+01	2E+02	1E+03	6E+01	2E+02	2E+03	2.7E+03
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	1E+02	3E+02	3E+03 a		same as adult		1.8E+03
108-90-7 Chlorobenzene	Vol	7E+02	2E+03 ^a	2E+04 ^b		same as adult		2.0E+03
124-48-1 Chlorodibromomethane	Vol	2E+02	7E+02	6E+03 ^a	2E+02	8E+02	7E+03 ^a	2.7E+03
67-66-3 Chloroform	Vol	3E+01	1E+02	1E+03	4E+01	1E+02	1E+03	5.7E+03
95-57-8 Chlorophenol, 2-	Semi		NM			NM		2.2E+04

Table 4-2. (continued)

		Ad	lult Resident*		Chil	d Resident 0-3 yı	r	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg
7440-47-3 Chromium VI	Metal	3E+02	1E+03	9E+03	4E+02	1E+03	1E+04	1.0E+0
7440-48-4 Cobalt	Metal	8E+02	3E+03	2E+04		same as adult		1.0E+0
1319-77-3 Cresols (total)	Semi		NM			NM		5.2E+0
98-82-8 Cumene	Vol	2E+04 ^b	7E+04 ^b	5E+05 ^b		same as adult		1.9E+0
108-93-0 Cyclohexanol	Vol	5E+00	2E+01	2E+02		same as adult		1.6E+
96-12-8 Dibromo-3-chloropropane, 1,2-	Vol	4E+04 ^b	1E+05 ^b	no risk b	4E+04 b	1E+05 b	no risk b	1.8E+
95-50-1 Dichlorobenzene, 1,2-	Vol	6E+04 ^b	2E+05 b	no risk b		same as adult		2.2E+
106-46-7 Dichlorobenzene, 1,4-	Vol	2E+05 ^b	7E+05 ^b	no risk b		same as adult		1.0E+
75-71-8 Dichlorodifluoromethane	Vol	2E+03 ^a	7E+03 ^a	7E+04 a		same as adult		1.2E+
107-06-2 Dichloroethane, 1,2-	Vol	3E+01	1E+02	1E+03	4E+01	2E+02	1E+03	3.4E+
75-35-4 Dichloroethylene, 1,1-	Vol	1E+01	5E+01	5E+02	2E+01	5E+01	5E+02	2.7E+
78-87-5 Dichloropropane, 1,2-	Vol	8E+01	2E+02	2E+03		same as adult		2.2E+
10061-01-5 Dichloropropene, cis-1,3-	Vol	8E+01	3E+02	3E+03 ^a	9E+01	3E+02	3E+03 ^a	2.2E+
10061-02-6 Dichloropropene, trans-1,3-	Vol	9E+01	3E+02	3E+03 ^a	1E+02	3E+02	3E+03 ^a	2.2E+
57-97-6 Dimethylbenz(a)anthracene, 7,12-	Semi		NM			NM		4.8E+
121-14-2 Dinitrotoluene, 2,4-	Semi		NM			NM		2.1E+
123-91-1 Dioxane, 1,4-	Vol	1E+05 °	4E+05 a,c	no risk		same as adult		2.3E+
122-66-7 Diphenylhydrazine, 1,2-	Semi		NM			NM		3.3E+
106-89-8 Epichlorohydrin	Vol	5E+03	2E+04 ^a	2E+05 a	6E+03	2E+04 ^a	2E+05 ^a	1.6E+
106-88-7 Epoxybutane, 1,2-	Vol	1E+03	4E+03	4E+04 a		same as adult		1.7E+
111-15-9 Ethoxyethanol acetate (R-R), 2-	Vol	1E+05 °	3E+05 a,c	no risk b,c		same as adult		2.3E+
110-80-5 Ethoxyethanol, 2-	Vol	3E+05 ^{a,}	8E+05 a,c	no risk b,c		same as adult		2.3E+
100-41-4 Ethylbenzene	Vol	4E+04 ^b	1E+05 ^b	no risk b		same as adult		1.3E+
106-93-4 Ethylene Dibromide	Vol	2E+01	6E+01	6E+02	2E+01	7E+01	7E+02	3.1E+
107-21-1 Ethylene glycol	Semi		NM			NM		2.3E+
75-21-8 Ethylene oxide	Vol	3E+01 °	1E+02 °	1E+03 °	4E+01 °	1E+02 °	1E+03 °	8.9E+
50-00-0 Formaldehyde	Vol	4E+03 °	1E+04 °	1E+05 a,c	5E+03 °	2E+04 °	2E+05 a,c	1.3E+

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Table 4-2. (continued)

		Ad	lult Resident*		Chil	d Resident 0-3 y	r	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg)
98-01-1 Furfural	Vol	4E+03	1E+04	1E+05 ^a		same as adult		2.7E+04
87-68-3 Hexachloro-1,3-butadiene	Semi		NM			NM		1.0E+03
118-74-1 Hexachlorobenzene	Semi		NM			NM		2.3E+04
77-47-4 Hexachlorocyclopentadiene	Semi		NM			NM		2.1E+03
67-72-1 Hexachloroethane	Vol	7E+03 ^b	2E+04 b	2E+05 b	8E+03 ^b	2E+04 b	2E+05 b	2.6E+03
78-59-1 Isophorone	Semi		NM			NM		6.1E+03
7439-92-1 Lead	Metal	1E+05	5E+05	no risk	4E+04	1E+05	no risk ^a	1.0E+06
7439-96-5 Manganese	Metal	4E+03	1E+04	1E+05		same as adult		1.0E+06
7439-97-6 Mercury	Metal	2E+01	5E+01 b	5E+02 b		same as adult		2.6E+01
67-56-1 Methanol	Vol	no risk ^{b,}	no risk b,c	no risk b,c		same as adult		2.3E+05
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	1E+04 ^c	4E+04 °	4E+05 a,c		same as adult		2.3E+05
109-86-4 Methoxyethanol, 2-	Vol	2E+04 °	5E+04 °	7E+05 a,c		same as adult		2.3E+05
74-83-9 Methyl bromide (Bromomethane)	Vol	8E+01	3E+02	3E+03		same as adult		5.7E+03
74-87-3 Methyl chloride (Chloromethane)	Vol	4E+02	1E+03	1E+04 ^a	4E+02	1E+03	1E+04 ^a	1.9E+03
78-93-3 Methyl ethyl ketone	Vol	2E+05 ^a	6E+05 ^a	no risk		same as adult		5.4E+04
108-10-1 Methyl isobutyl ketone	Vol	1E+04 ^a	5E+04 ^a	4E+05 b		same as adult		6.0E+03
80-62-6 Methyl methacrylate	Vol	5E+04 ^a	2E+05 b	no risk b		same as adult		5.5E+03
1634-04-4 Methyl tert-butyl ether	Vol	1E+05 ^a	3E+05 b	no risk b		same as adult		2.6E+04
56-49-5 Methylcholanthrene, 3-	Semi		NM			NM		4.0E+01
75-09-2 Methylene chloride	Vol	2E+03	8E+03 ^a	8E+04 ^a	2E+03	9E+03 ^a	9E+04 ^a	4.6E+03
68-12-2 N,N-Dimethylformamide	Semi		NM			NM		2.3E+05
110-54-3 n-Hexane	Vol	7E+03 ^a	2E+04 b	2E+05 b		same as adult		6.4E+02
930-55-2 N-Nitrosopyrrolidine	Vol	2E+02 °	6E+02 °	7E+03 °	2E+02 °	7E+02 °	7E+03 °	2.3E+05
91-20-3 Naphthalene	Vol	3E+03 ^a	8E+03 b	9E+04 b		same as adult		3.8E+02
7440-02-0 Nickel	Metal	2E+04	5E+04	5E+05	2E+04	6E+04	5E+05	1.0E+06
98-95-3 Nitrobenzene	Semi		NM			NM		1.3E+03
79-46-9 Nitropropane, 2-	Vol	8E-01	3E+00	4E+01	9E-01	4E+00	4E+01	4.6E+03
924-16-3 Nitrosodi-n-butylamine	Vol	4E+00	1E+01	1E+02	4E+00	1E+01	1E+02	2.1E+03

Table 4-2. (continued)

		Ad	lult Resident*		Chil	d Resident 0-3 y	ſ	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg)
55-18-5 Nitrosodiethylamine	Vol	3E-01	8E-01	7E+00	3E-01	1E+00	8E+00	2.3E+04
108-95-2 Phenol	Semi		NM			NM		3.3E+04
85-44-9 Phthalic anhydride	Semi		NM			NM		1.4E+03
75-56-9 Propylene oxide	Vol	1E+03 °	4E+03 °	4E+04 °	1E+03 °	5E+03 °	5E+04 °	1.1E+05
110-86-1 Pyridine	Vol	5E+03	1E+04	2E+05		same as adult		2.6E+05
100-42-5 Styrene	Vol	3E+05 b	9E+05 ^b	no risk b		same as adult		1.5E+03
1746-01-6 TCDD, 2,3,7,8-	Semi		NM			NM		3.8E-01
630-20-6 Tetrachloroethane, 1,1,1,2-	Vol	3E+02	1E+03	1E+04 a	3E+02	1E+03	1E+04 ^a	2.8E+03
79-34-5 Tetrachloroethane, 1,1,2,2-	Vol	1E+02	4E+02	5E+03 ^a	1E+02	5E+02	5E+03 ^a	4.7E+03
127-18-4 Tetrachloroethylene	Vol	1E+03 ^a	5E+03 ^a	4E+04 b	2E+03 ^a	6E+03 ^a	5E+04 b	5.9E+02
108-88-3 Toluene	Vol	1E+04 ^a	4E+04 b	4E+05 b		same as adult		1.7E+03
95-53-4 Toluidine, o-	Semi		NM					5.9E+03
76-13-1 Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol	4E+05 ^b	no risk ^b	no risk b		same as adult		2.1E+03
120-82-1 Trichlorobenzene, 1,2,4-	Semi		NM			NM		1.6E+04
71-55-6 Trichloroethane, 1,1,1-	Vol	2E+04 ^a	5E+04 b	4E+05 b		same as adult		2.7E+03
79-00-5 Trichloroethane, 1,1,2-	Vol	8E+01	2E+02	2E+03	1E+02	3E+02	3E+03	3.8E+03
79-01-6 Trichloroethylene	Vol	5E+02	2E+03	2E+04 a	6E+02	2E+03	2E+04 ^a	3.4E+03
75-69-4 Trichlorofluoromethane	Vol	9E+03 ^a	3E+04 ^a	3E+05 b		same as adult		3.3E+03
121-44-8 Triethylamine	Vol	2E+02	7E+02	6E+03		same as adult		2.1E+04
7440-62-2 Vanadium	Metal	6E+03	2E+04	2E+05		same as adult		1.0E+06
108-05-4 Vinyl acetate	Vol	1E+04 ^a	4E+04 ^a	4E+05		same as adult		5.3E+03
75-01-4 Vinyl chloride	Vol	7E+00	2E+01	2E+02	8E+00	3E+01	3E+02	1.8E+03
1330-20-7 Xylenes (total)	Vol	2E+04 b	6E+04 b	6E+05 b		same as adult		1.5E+03

NM = Not modeled for land-based units.

^{*} For lead only, this column contains results for Child Resident 3-7 years.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^c Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

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Table 4-3. Chronic 90/90 C_w at 25 m to 1,000 m, Risk = $10^{-5}/HQ = 1$ for Wastepiles (mg/kg)

		A	dult Resident*		C	hild Resident 0-3 y	/r	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	- C _{sat} (mg/kg
75-07-0 Acetaldehyde	Vol	8E+03 °	6E+04 °	no risk ^{a,c}	9E+03 °	7E+04 °	no risk ^{a,c}	3.3E+0
67-64-1 Acetone	Vol	no risk b,c	no risk b,c	no risk b,c		same as adult		2.3E+0
75-05-8 Acetonitrile	Vol	2E+04	1E+05	no risk b		same as adult		2.3E+0
107-02-8 Acrolein	Vol	3E+00	2E+01	5E+02		same as adult		5.0E+0
79-06-1 Acrylamide	Semi		NM			NM		1.5E+0
79-10-7 Acrylic acid	Semi		NM			NM		2.4E+0
107-13-1 Acrylonitrile	Vol	1E+02	1E+03	2E+04 ^a	1E+02	1E+03	2E+04 ^a	1.8E+0
107-05-1 Allyl chloride	Vol	3E+01	2E+02	4E+03 ^a		same as adult		1.7E+0
62-53-3 Aniline	Semi		NM			NM		1.0E+0
7440-38-2 Arsenic	Metal	7E+02	6E+03	1E+05	8E+02	7E+03	1E+05	1.0E+
7440-39-3 Barium	Metal	3E+04	3E+05	no risk ^a		same as adult		1.0E+
71-43-2 Benzene	Vol	6E+02	4E+03 ^a	9E+04 b	7E+02	5E+03 ^a	1E+05 b	1.8E+
92-87-5 Benzidine	Semi		NM			NM		2.4E+
50-32-8 Benzo(a)pyrene	Semi		NM			NM		1.5E+
7440-41-7 Beryllium	Metal	1E+03	1E+04	2E+05	1E+03	1E+04	3E+05	1.0E+
75-27-4 Bromodichloromethane	Vol	7E+02	5E+03	1E+05 ^b	8E+02	6E+03	1E+05 b	6.3E+
75-25-2 Bromoform (Tribromomethane)	Vol	8E+04 b	6E+05 ^b	no risk b	9E+04 b	7E+05 ^b	no risk	4.5E+
106-99-0 Butadiene, 1,3-	Vol	1E+00	1E+01	2E+02	1E+00	1E+01	2E+02	1.1E+
7440-43-9 Cadmium	Metal	2E+03	1E+04	3E+05	2E+03	2E+04	4E+05	1.0E+
75-15-0 Carbon disulfide	Vol	2E+04 ^a	1E+05 ^b	no risk b		same as adult		1.3E+
56-23-5 Carbon tetrachloride	Vol	2E+02	1E+03	3E+04 b	2E+02	1E+03	3E+04 b	2.7E+
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	3E+02	2E+03 ^a	4E+04		same as adult		1.8E+0
108-90-7 Chlorobenzene	Vol	4E+03 ^a	3E+04 ^b	5E+05		same as adult		2.0E+0
124-48-1 Chlorodibromomethane	Vol	2E+03	1E+04 ^a	3E+05 b	2E+03	2E+04 ^a	4E+05	2.7E+0
67-66-3 Chloroform	Vol	1E+02	9E+02	2E+04 ^a	1E+02	1E+03	2E+04 ^a	5.7E+0
95-57-8 Chlorophenol, 2-	Semi		NM			NM		2.2E+0

Table 4-3. (continued)

		Į.	Adult Resident*	•	Ch	ild Resident 0-3 y	/r	•
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg
7440-47-3 Chromium VI	Metal	2E+02	2E+03	5E+04	3E+02	2E+03	5E+04	1.0E+0
7440-48-4 Cobalt	Metal	6E+02	5E+03	1E+05		same as adult		1.0E+0
1319-77-3 Cresols (total)	Semi		NM			NM		5.2E+0
98-82-8 Cumene	Vol	3E+04 ^b	2E+05 ^b	no risk b		same as adult		1.9E+0
108-93-0 Cyclohexanol	Vol	1E+01	1E+02	2E+03		same as adult		1.6E+0
96-12-8 Dibromo-3-chloropropane, 1,2-	Vol	1E+05 b	9E+05 b	no risk b	1E+05 b	no risk	no risk	1.8E+0
95-50-1 Dichlorobenzene, 1,2-	Vol	2E+05 b	no risk ^b	no risk b		same as adult		2.2E+0
106-46-7 Dichlorobenzene, 1,4-	Vol	6E+05 ^b	no risk b	no risk b		same as adult		1.0E+0
75-71-8 Dichlorodifluoromethane	Vol	3E+03 ^a	2E+04 ^a	4E+05 ^b		same as adult		1.2E+0
107-06-2 Dichloroethane, 1,2-	Vol	2E+02	1E+03	2E+04 ^a	2E+02	1E+03	3E+04 ^a	3.4E+0
75-35-4 Dichloroethylene, 1,1-	Vol	3E+01	2E+02	5E+03 ^a	4E+01	3E+02	5E+03 ^a	2.7E+0
78-87-5 Dichloropropane, 1,2-	Vol	3E+02	2E+03 ^a	5E+04 b		same as adult		2.2E+0
10061-01-5 Dichloropropene, cis-1,3-	Vol	3E+02	2E+03 ^a	5E+04 ^a	4E+02	3E+03 ^a	5E+04 ^a	2.2E+0
10061-02-6 Dichloropropene, trans-1,3-	Vol	4E+02	3E+03 ^a	6E+04 ^a	4E+02	3E+03 ^a	6E+04 ^b	2.2E+0
57-97-6 Dimethylbenz(a)anthracene, 7,12-	Semi		NM			NM		4.8E+0
121-14-2 Dinitrotoluene, 2,4-	Semi		NM			NM		2.1E+0
123-91-1 Dioxane, 1,4-	Vol	4E+05 a,c	no risk b,c	no risk b,c		same as adult		2.3E+0
122-66-7 Diphenylhydrazine, 1,2-	Semi		NM			NM		3.3E+0
106-89-8 Epichlorohydrin	Vol	2E+04 ^a	1E+05 ^a	no risk b	2E+04 ^a	1E+05 ^a	no risk ^b	1.6E+0
106-88-7 Epoxybutane, 1,2-	Vol	4E+03	2E+04 ^a	5E+05 b		same as adult		1.7E+0
111-15-9 Ethoxyethanol acetate (R-R), 2-	Vol	3E+05 a,c	no risk b,c	no risk		same as adult		2.3E+0
110-80-5 Ethoxyethanol, 2-	Vol	7E+05 a,c	no risk b,c	no risk b,c		same as adult		2.3E+0
100-41-4 Ethylbenzene	Vol	3E+05 b	no risk b	no risk b		same as adult		1.3E+0
106-93-4 Ethylene Dibromide	Vol	2E+02	1E+03	3E+04 ^a	2E+02	1E+03	3E+04 ^a	3.1E+0
107-21-1 Ethylene glycol	Semi		NM			NM		2.3E+0
75-21-8 Ethylene oxide	Vol	8E+01 °	6E+02 °	1E+04 °	9E+01 °	7E+02 °	1E+04 °	8.9E+0
50-00-0 Formaldehyde	Vol	1E+04 °	7E+04 °	no risk a,c	1E+04 °	8E+04 °	no risk a,c	1.3E+0

Table 4-3. (continued)

		-	Adult Resident*		С	hild Resident 0-3 y	Resident 0-3 yr		
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	C _{sat} (mg/kg	
98-01-1 Furfural	Vol	4E+04 a	3E+05 ^a	no risk b		same as adult		2.7E+0	
87-68-3 Hexachloro-1,3-butadiene	Semi		NM			NM		1.0E+0	
118-74-1 Hexachlorobenzene	Semi		NM			NM		2.3E+0	
77-47-4 Hexachlorocyclopentadiene	Semi		NM			NM		2.1E+0	
67-72-1 Hexachloroethane	Vol	2E+05 b	no risk	no risk b	3E+05 b	no risk b	no risk b	2.6E+0	
78-59-1 Isophorone	Semi		NM			NM		6.1E+0	
7439-92-1 Lead	Metal	1E+05	9E+05	no risk ^a	3E+04	3E+05	no risk ^a	1.0E+0	
7439-96-5 Manganese	Metal	3E+03	3E+04	5E+05		same as adult		1.0E+0	
7439-97-6 Mercury	Metal	2E+02 a	1E+03 b	3E+04 b		same as adult		2.6E+0	
67-56-1 Methanol	Vol	no risk b,c	no risk b,c	no risk b,c		same as adult		2.3E+0	
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	3E+04 °	2E+05 °	no risk b,c		same as adult		2.3E+0	
109-86-4 Methoxyethanol, 2-	Vol	5E+04 °	4E+05 a,c	no risk b,c		same as adult		2.3E+0	
74-83-9 Methyl bromide (Bromomethane)	Vol	2E+02	1E+03	3E+04 ^a		same as adult		5.7E+0	
74-87-3 Methyl chloride (Chloromethane)	Vol	5E+02	4E+03 ^a	8E+04 ^a	6E+02	5E+03 ^a	9E+04 ^a	1.9E+0	
78-93-3 Methyl ethyl ketone	Vol	4E+05 ^a	no risk b	no risk b		same as adult		5.4E+0	
108-10-1 Methyl isobutyl ketone	Vol	3E+04 ^a	2E+05 b	no risk b		same as adult		6.0E+0	
80-62-6 Methyl methacrylate	Vol	2E+05 b	no risk b	no risk b		same as adult		5.5E+0	
1634-04-4 Methyl tert-butyl ether	Vol	5E+05 b	no risk b	no risk ^b		same as adult		2.6E+0	
56-49-5 Methylcholanthrene, 3-	Semi		NM			NM		4.0E+0	
75-09-2 Methylene chloride	Vol	6E+03 ^a	5E+04 ^a	no risk b	7E+03 ^a	6E+04 ^a	no risk b	4.6E+0	
68-12-2 N,N-Dimethylformamide	Semi		NM			NM		2.3E+0	
110-54-3 n-Hexane	Vol	4E+04 b	2E+05 b	no risk b		same as adult		6.4E+0	
930-55-2 N-Nitrosopyrrolidine	Vol	8E+02 °	6E+03 °	1E+05 °	9E+02 °	7E+03 °	1E+05 °	2.3E+0	
91-20-3 Naphthalene	Vol	7E+03 ^b	4E+04 b	8E+05 ^b		same as adult		3.8E+0	
7440-02-0 Nickel	Metal	1E+04	1E+05	no risk ^a	1E+04	1E+05	no risk ^a	1.0E+0	
98-95-3 Nitrobenzene	Semi		NM			NM		1.3E+0	
79-46-9 Nitropropane, 2-	Vol	5E+00	4E+01	8E+02	6E+00	4E+01	9E+02	4.6E+0	
924-16-3 Nitrosodi-n-butylamine	Vol	1E+01	1E+02	2E+03	2E+01	1E+02	2E+03 ^a	2.1E+0	

Table 4-3. (continued)

		Į.	Adult Resident	•	Ch	nild Resident 0-3	/r	- C _{sat}
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	(mg/kg)
55-18-5 Nitrosodiethylamine	Vol	9E-01	7E+00	1E+02	1E+00	8E+00	2E+02	2.3E+04
108-95-2 Phenol	Semi		NM			NM		3.3E+04
85-44-9 Phthalic anhydride	Semi		NM			NM		1.4E+03
75-56-9 Propylene oxide	Vol	3E+03 °	2E+04 °	4E+05 a,c	3E+03 °	2E+04 °	5E+05 a,c	1.1E+05
110-86-1 Pyridine	Vol	1E+04	7E+04	no risk b		same as adult		2.6E+05
100-42-5 Styrene	Vol	5E+05 b	no risk b	no risk		same as adult		1.5E+03
1746-01-6 TCDD, 2,3,7,8-	Semi		NM			NM		3.8E-01
630-20-6 Tetrachloroethane, 1,1,1,2-	Vol	2E+03	1E+04 ^a	3E+05 ^b	2E+03	1E+04 ^a	3E+05 ^b	2.8E+03
79-34-5 Tetrachloroethane, 1,1,2,2-	Vol	7E+02	6E+03 ^a	1E+05 ^b	8E+02	7E+03 ^a	1E+05 ^b	4.7E+03
127-18-4 Tetrachloroethylene	Vol	6E+03 b	4E+04 b	9E+05 ^b	6E+03 b	5E+04 b	no risk b	5.9E+02
108-88-3 Toluene	Vol	5E+04 b	4E+05	no risk		same as adult		1.7E+03
95-53-4 Toluidine, o-	Semi		NM			NM		5.9E+03
76-13-1 Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol	9E+05 ^b	no risk	no risk b		same as adult		2.1E+03
120-82-1 Trichlorobenzene, 1,2,4-	Semi		NM			NM		1.6E+04
71-55-6 Trichloroethane, 1,1,1-	Vol	5E+04 ^b	4E+05 b	no risk b		same as adult		2.7E+03
79-00-5 Trichloroethane, 1,1,2-	Vol	4E+02	3E+03	7E+04 ^b	5E+02	4E+03 ^a	8E+04	3.8E+0
79-01-6 Trichloroethylene	Vol	2E+03	2E+04 b	4E+05	3E+03	2E+04 b	4E+05	3.4E+0
75-69-4 Trichlorofluoromethane	Vol	2E+04 ^a	1E+05 b	no risk b		same as adult		3.3E+0
121-44-8 Triethylamine	Vol	1E+03	1E+04	2E+05 ^a		same as adult		2.1E+0
7440-62-2 Vanadium	Metal	4E+03	4E+04	8E+05		same as adult		1.0E+0
108-05-4 Vinyl acetate	Vol	3E+04 ^a	2E+05 b	no risk		same as adult		5.3E+0
75-01-4 Vinyl chloride	Vol	9E+00	7E+01	1E+03	1E+01	8E+01	2E+03	1.8E+0
1330-20-7 Xylenes (total)	Vol	1E+05 ^b	7E+05 b	no risk b		same as adult		1.5E+0

NM = Not modeled for land-based units.

^{*} For lead only, this column contains results for Child Resident 3-7 years.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^c Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

Table 4-4. Chronic 90/90 C_w at 25 m - 1,000 m, Risk = $10^{-5}/HQ = 1$ for Aerated Treatment Tanks (mg/L)

			Adult Resident*		Chil	d Resident 0-3 yr		0-1-1-1111
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
75-07-0 Acetaldehyde	Vol	7E+01	4E+02	9E+03	8E+01	5E+02	1E+04	1.0E+06
67-64-1 Acetone	Vol	1E+05	no risk	no risk ^b		same as adult		1.0E+06
75-05-8 Acetonitrile	Vol	3E+02	2E+03	4E+04		same as adult		1.0E+06
107-02-8 Acrolein	Vol	5E-02	3E-01	6E+00		same as adult		2.1E+05
79-06-1 Acrylamide	Semi	5E+03	4E+04	8E+05 ^b	6E+03	4E+04	9E+05 ^b	6.4E+05
79-10-7 Acrylic acid	Semi	7E+02	5E+03 °	1E+05		same as adult		1.0E+06
107-13-1 Acrylonitrile	Vol	2E+00	1E+01	3E+02	2E+00	1E+01	3E+02	7.4E+04
107-05-1 Allyl chloride	Vol	6E-01	4E+00	9E+01		same as adult		3.4E+03
62-53-3 Aniline	Semi	9E+01	6E+02	1E+04		same as adult		3.6E+04
7440-38-2 Arsenic	Metal		NM			NM		0.0E+00
7440-39-3 Barium	Metal		NM			NM		0.0E+00
71-43-2 Benzene	Vol	4E+00	3E+01	6E+02	5E+00	3E+01	7E+02	1.8E+03
92-87-5 Benzidine	Semi	8E+03 b	5E+04 b	9E+05 ^b	9E+03 ^b	5E+04 b	no risk b	5.0E+02
50-32-8 Benzo(a)pyrene	Semi	2E+02 b	1E+03 b	2E+04 b	2E+02 b	1E+03 b	3E+04 b	2.5E-02
7440-41-7 Beryllium	Metal		NM			NM		0.0E+00
75-27-4 Bromodichloromethane	Vol	3E+00	2E+01	4E+02	4E+00	2E+01	5E+02	6.7E+03
75-25-2 Bromoform (Tribromomethane)	Vol	1E+02	6E+02	1E+04 b	1E+02	7E+02	1E+04 ^b	3.1E+03
106-99-0 Butadiene, 1,3-	Vol	5E-02	4E-01	8E+00	6E-02	4E-01	9E+00	7.4E+02
7440-43-9 Cadmium	Metal		NM			NM		0.0E+00
75-15-0 Carbon disulfide	Vol	4E+02	3E+03 ^a	6E+04 b		same as adult		1.2E+03
56-23-5 Carbon tetrachloride	Vol	2E+00	1E+01	3E+02	2E+00	2E+01	4E+02	7.9E+02
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	4E+00	3E+01	6E+02		same as adult		1.7E+03
108-90-7 Chlorobenzene	Vol	2E+01	1E+02	2E+03 ^a		same as adult		4.7E+02

Table 4-4. (continued)

		,	Adult Resident*		Chil	d Resident 0-3 yı	•	
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
124-48-1 Chlorodibromomethane	Vol	4E+00	2E+01	5E+02	4E+00	2E+01	5E+02	2.6E+03
67-66-3 Chloroform	Vol	2E+00	1E+01	2E+02	2E+00	1E+01	3E+02	7.9E+03
95-57-8 Chlorophenol, 2-	Semi	3E+00	2E+01	3E+02		same as adult		2.2E+04
7440-47-3 Chromium VI	Metal		NM			NM		0.0E+00
7440-48-4 Cobalt	Metal		NM			NM		0.0E+00
1319-77-3 Cresols (total)	Semi	4E+01	3E+02	5E+03		same as adult		2.2E+04
98-82-8 Cumene	Vol	2E+02 ^a	2E+03 b	4E+04 b		same as adult		6.1E+01
108-93-0 Cyclohexanol	Vol	8E-01	4E+00	9E+01		same as adult		3.6E+04
96-12-8 Dibromo-3-chloropropane, 1,2-	Vol	3E+02	2E+03 b	5E+04 b	4E+02	2E+03 b	5E+04 ^b	1.2E+03
95-50-1 Dichlorobenzene, 1,2-	Vol	2E+02 ^a	1E+03 b	3E+04 ^b		same as adult		1.6E+02
106-46-7 Dichlorobenzene, 1,4-	Vol	8E+02 b	5E+03	1E+05 ^b		same as adult		7.4E+01
75-71-8 Dichlorodifluoromethane	Vol	1E+02	8E+02 ^a	2E+04 ^a		same as adult		2.8E+02
107-06-2 Dichloroethane, 1,2-	Vol	2E+00	1E+01	3E+02	2E+00	1E+01	3E+02	8.5E+03
75-35-4 Dichloroethylene, 1,1-	Vol	5E-01	4E+00	9E+01	6E-01	5E+00	1E+02	2.3E+03
78-87-5 Dichloropropane, 1,2-	Vol	3E+00	2E+01	5E+02		same as adult		2.8E+03
10061-01-5 Dichloropropene, cis-1,3-	Vol	1E+00	8E+00	2E+02	1E+00	9E+00	2E+02	2.7E+03
10061-02-6 Dichloropropene, trans-1,3-	Vol	1E+00	1E+01	2E+02	2E+00	1E+01	2E+02	2.7E+03
57-97-6 Dimethylbenz(a)anthracene, 7,12-	Semi	6E+02 b	4E+03 b	8E+04 ^b	8E+02 b	4E+03 ^b	9E+04 ^b	2.5E-02
121-14-2 Dinitrotoluene, 2,4-	Semi	2E+02	2E+03 ^b	4E+04 ^b	3E+02 b	2E+03 ^b	4E+04 ^b	2.7E+02
123-91-1 Dioxane, 1,4-	Vol	1E+04	9E+04	no risk ^b		same as adult		1.0E+06
122-66-7 Diphenylhydrazine, 1,2-	Semi	9E+01 b	5E+02 ^b	1E+04 ^b	1E+02 b	6E+02 b	1E+04 ^b	6.8E+01
106-89-8 Epichlorohydrin	Vol	3E+02	2E+03	4E+04	4E+02	2E+03	5E+04	6.6E+04
106-88-7 Epoxybutane, 1,2-	Vol	3E+01	1E+02	3E+03		same as adult		4.3E+04

(continued)

			Adult Resident*		Chil	d Resident 0-3 yr		O - b - b - b - b - b - b - b - b - b -
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
111-15-9 Ethoxyethanol acetate (R-R), 2-	Vol	1E+04	7E+04	no risk b		same as adult		1.0E+06
110-80-5 Ethoxyethanol, 2-	Vol	6E+04	4E+05	no risk ^b		same as adult		1.0E+06
100-41-4 Ethylbenzene	Vol	7E+02 ^a	5E+03 b	1E+05 ^b		same as adult		1.7E+02
106-93-4 Ethylene Dibromide	Vol	4E-01	2E+00	5E+01	5E-01	3E+00	6E+01	4.2E+03
107-21-1 Ethylene glycol	Semi	no risk ^b	no risk b	no risk ^b		same as adult		1.0E+06
75-21-8 Ethylene oxide	Vol	1E+00	8E+00	2E+02	1E+00	9E+00	2E+02	3.8E+05
50-00-0 Formaldehyde	Vol	7E+02 °	4E+03 °	9E+04 °	8E+02 °	5E+03 °	1E+05 °	5.5E+05
98-01-1 Furfural	Vol	3E+03	1E+04	3E+05 b		same as adult		1.1E+05
87-68-3 Hexachloro-1,3-butadiene	Semi	3E+00	2E+01 b	4E+02 b	3E+00 ^a	2E+01 b	5E+02 b	3.2E+00
118-74-1 Hexachlorobenzene	Semi	9E-01	5E+00	1E+02 ^b	1E+00	6E+00	1E+02 ^b	6.2E+00
77-47-4 Hexachlorocyclopentadiene	Semi	1E-01	7E-01	2E+01 ^b		same as adult		1.8E+00
67-72-1 Hexachloroethane	Vol	3E+01	2E+02 b	4E+03 b	4E+01	2E+02 b	5E+03 ^b	5.0E+01
78-59-1 Isophorone	Semi	4E+02	2E+03	4E+04 ^b		same as adult		1.2E+04
7439-92-1 Lead	Metal		NM			NM		0.0E+00
7439-96-5 Manganese	Metal		NM			NM		0.0E+00
7439-97-6 Mercury	Metal	2E-01 a	2E+00 b	4E+01 b		same as adult		5.6E-02
67-56-1 Methanol	Vol	3E+05	no risk b	no risk b		same as adult		1.0E+06
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	1E+03	9E+03	2E+05		same as adult		1.0E+06
109-86-4 Methoxyethanol, 2-	Vol	5E+03	3E+04	7E+05		same as adult		1.0E+06
74-83-9 Methyl bromide (Bromomethane)	Vol	3E+00	2E+01	5E+02		same as adult		1.5E+04
74-87-3 Methyl chloride (Chloromethane)	Vol	2E+01	1E+02	3E+03	2E+01	1E+02	3E+03	5.3E+03
78-93-3 Methyl ethyl ketone	Vol	4E+03	3E+04	6E+05 b		same as adult		2.2E+05

Table 4-4. (continued)

Table 4-4. (continued)

			Adult Resident*		Chil	d Resident 0-3 yr		0.1.1.111
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
108-10-1 Methyl isobutyl ketone	Vol	2E+02	1E+03	3E+04 b		same as adult	•	1.9E+04
80-62-6 Methyl methacrylate	Vol	1E+03	7E+03	1E+05 ^b		same as adult		1.5E+04
1634-04-4 Methyl tert-butyl ether	Vol	4E+03	2E+04	5E+05 ^b		same as adult		3.9E+04
56-49-5 Methylcholanthrene, 3-	Semi	2E+02 b	1E+03 ^b	2E+04 b	2E+02 b	1E+03 ^b	3E+04 ^b	3.2E-03
75-09-2 Methylene chloride	Vol	8E+01	5E+02	1E+04	9E+01	6E+02	1E+04 ^a	1.3E+04
68-12-2 N,N-Dimethylformamide	Semi	2E+04	1E+05	no risk ^b		same as adult		1.0E+06
110-54-3 n-Hexane	Vol	1E+02 ^a	1E+03 ^a	2E+04 ^a		same as adult		1.2E+01
930-55-2 N-Nitrosopyrrolidine	Vol	4E+02	3E+03	5E+04	4E+02	3E+03	6E+04	1.0E+06
91-20-3 Naphthalene	Vol	6E+00	3E+01 b	7E+02 ^b		same as adult		3.1E+01
7440-02-0 Nickel	Metal		NM			NM		0.0E+00
98-95-3 Nitrobenzene	Semi	2E+01	1E+02	3E+03 b		same as adult		2.1E+03
79-46-9 Nitropropane, 2-	Vol	6E-02	4E-01	7E+00	6E-02	4E-01	8E+00	1.7E+04
924-16-3 Nitrosodi-n-butylamine	Vol	5E-02	3E-01	6E+00	5E-02	3E-01	7E+00	1.3E+03
55-18-5 Nitrosodiethylamine	Vol	3E-02	2E-01	4E+00	3E-02	2E-01	4E+00	9.3E+04
108-95-2 Phenol	Semi	2E+03	1E+04	3E+05 ^b		same as adult		8.3E+04
85-44-9 Phthalic anhydride	Semi	no risk b	no risk b	no risk ^b		same as adult		6.2E+03
75-56-9 Propylene oxide	Vol	4E+01	3E+02	6E+03	5E+01	3E+02	6E+03	4.8E+05
110-86-1 Pyridine	Vol	1E+02	7E+02	2E+04		same as adult		1.0E+06
100-42-5 Styrene	Vol	9E+02 ^a	6E+03 b	1E+05 ^b		same as adult		3.1E+02
1746-01-6 TCDD, 2,3,7,8-	Semi	1E-03 ^b	5E-03 ^b	1E-01 b	1E-03 ^b	6E-03 b	1E-01 ^b	1.9E-05
630-20-6 Tetrachloroethane, 1,1,1,2-	Vol	6E+00	4E+01	9E+02	7E+00	5E+01	1E+03	1.1E+03
79-34-5 Tetrachloroethane, 1,1,2,2-	Vol	2E+00	1E+01	2E+02	2E+00	1E+01	2E+02	3.0E+03
127-18-4 Tetrachloroethylene	Vol	5E+01	4E+02 a	8E+03 ^b	6E+01	4E+02 a	9E+03 ^b	2.0E+02
108-88-3 Toluene	Vol	3E+02	2E+03 a	4E+04 b		same as adult		5.3E+02

Section 4.0

Table 4-4. (continued)

			Adult Resident*			ld Resident 0-3 yı	r	Calmbilita
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
95-53-4 Toluidine, o-	Semi	5E+01	3E+02	7E+03	6E+01	4E+02	8E+03	1.7E+04
76-13-1 Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol	2E+04 ^a	1E+05 ^b	no risk ^b		same as adult		1.7E+02
120-82-1 Trichlorobenzene, 1,2,4-	Semi	3E+02 ^b	2E+03 b	4E+04 ^b		same as adult		3.0E+02
71-55-6 Trichloroethane, 1,1,1-	Vol	6E+02	4E+03 ^a	9E+04 ^b		same as adult		1.3E+03
79-00-5 Trichloroethane, 1,1,2-	Vol	4E+00	2E+01	5E+02	5E+00	3E+01	6E+02	4.4E+03
79-01-6 Trichloroethylene	Vol	2E+01	1E+02	3E+03 ^a	2E+01	2E+02	3E+03 ^a	1.1E+03
75-69-4 Trichlorofluoromethane	Vol	4E+02	3E+03 ^a	6E+04 ^a		same as adult		1.1E+03
121-44-8 Triethylamine	Vol	2E+01	1E+02	2E+03		same as adult		5.5E+04
7440-62-2 Vanadium	Metal		NM			NM		0.0E+00
108-05-4 Vinyl acetate	Vol	3E+02	2E+03	4E+04 ^b		same as adult		2.0E+04
75-01-4 Vinyl chloride	Vol	3E-01	2E+00	5E+01	4E-01	3E+00	6E+01	2.8E+03
1330-20-7 Xylenes (total)	Vol	3E+02 ^a	2E+03 b	4E+04 b		same as adult		1.9E+02

NM = Not modeled for tanks.

^{*} For lead only, this column contains results for Child Resident 3-7 yrs.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25 °C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25 °C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

[°] Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

Table 4-5. Chronic 90/90 C_w at 25 m to 1000 m, Risk = $10^{-5}/HQ = 1$ for Nonaerated Treatment Tanks (mg/L)

			Adult Resident*		CI	nild Resident 0-3	yr	Calubilitu
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
75-07-0 Acetaldehyde	Vol	8E+01	4E+02	7E+03	8E+01	5E+02	8E+03	1.0E+06
67-64-1 Acetone	Vol	1E+05	8E+05	no risk b		same as adult		1.0E+06
75-05-8 Acetonitrile	Vol	3E+02	1E+03	3E+04		same as adult		1.0E+06
107-02-8 Acrolein	Vol	6E-02	3E-01	6E+00		same as adult		2.1E+05
79-06-1 Acrylamide	Semi	6E+03	3E+04	5E+05	7E+03	3E+04	6E+05	6.4E+05
79-10-7 Acrylic acid	Semi	7E+02	3E+03	6E+04		same as adult		1.0E+06
107-13-1 Acrylonitrile	Vol	2E+00	1E+01	2E+02	3E+00	1E+01	3E+02	7.4E+04
107-05-1 Allyl chloride	Vol	1E+00	5E+00	1E+02		same as adult		3.4E+03
62-53-3 Aniline	Semi	8E+01	4E+02	8E+03		same as adult		3.6E+04
7440-38-2 Arsenic	Metal		NM			NM		0.0E+00
7440-39-3 Barium	Metal		NM			NM		0.0E+00
71-43-2 Benzene	Vol	7E+00	4E+01	7E+02	8E+00	4E+01	8E+02	1.8E+03
92-87-5 Benzidine	Semi	7E+03 b	3E+04 b	6E+05 b	8E+03 b	4E+04 b	7E+05	5.0E+02
50-32-8 Benzo(a)pyrene	Semi	2E+02 b	8E+02 b	1E+04 b	2E+02 b	9E+02 b	2E+04	2.5E-02
7440-41-7 Beryllium	Metal		NM			NM		0.0E+00
75-27-4 Bromodichloromethane	Vol	5E+00	2E+01	5E+02	6E+00	3E+01	5E+02	6.7E+03
75-25-2 Bromoform (Tribromomethane)	Vol	1E+02	7E+02	1E+04 ^b	2E+02	8E+02	2E+04 b	3.1E+03
106-99-0 Butadiene, 1,3-	Vol	7E-02	4E-01	9E+00	9E-02	5E-01	1E+01	7.4E+02
7440-43-9 Cadmium	Metal		NM			NM		0.0E+00
75-15-0 Carbon disulfide	Vol	6E+02	3E+03 ^a	7E+04 ^a		same as adult		1.2E+03
56-23-5 Carbon tetrachloride	Vol	3E+00	2E+01	4E+02	4E+00	2E+01	4E+02	7.9E+02
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	7E+00	4E+01	7E+02		same as adult		1.7E+03
108-90-7 Chlorobenzene	Vol	3E+01	1E+02	3E+03 ^a		same as adult		4.7E+02
124-48-1 Chlorodibromomethane	Vol	5E+00	3E+01	5E+02	6E+00	3E+01	6E+02	2.6E+03
67-66-3 Chloroform	Vol	3E+00	1E+01	3E+02	3E+00	2E+01	3E+02	7.9E+03
95-57-8 Chlorophenol, 2-	Semi	4E+00	2E+01	4E+02		same as adult		2.2E+04

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Table 4-5. (continued)

				Adult Resident	*	С	hild Resident 0-3	yr	Solubility
CAS	Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	(mg/L)
7440-47-3	Chromium VI	Metal		NM			NM		0.0E+00
7440-48-4	Cobalt	Metal		NM			NM		0.0E+00
1319-77-3	Cresols (total)	Semi	4E+01	2E+02	3E+03		same as adult		2.2E+04
98-82-8	Cumene	Vol	4E+02 a	2E+03 b	5E+04 b		same as adult		6.1E+01
108-93-0	Cyclohexanol	Vol	6E-01	3E+00	5E+01		same as adult		3.6E+04
96-12-8	Dibromo-3-chloropropane, 1,2-	Vol	4E+02	2E+03 ^b	4E+04 b	5E+02	3E+03 ^b	5E+04 b	1.2E+03
95-50-1	Dichlorobenzene, 1,2-	Vol	4E+02 a	2E+03 b	4E+04 b		same as adult		1.6E+02
106-46-7	Dichlorobenzene, 1,4-	Vol	2E+03 b	8E+03 b	1E+05 b		same as adult		7.4E+01
75-71-8	Dichlorodifluoromethane	Vol	2E+02	1E+03 ^a	2E+04 a		same as adult		2.8E+02
107-06-2	Dichloroethane, 1,2-	Vol	3E+00	2E+01	3E+02	4E+00	2E+01	4E+02	8.5E+03
75-35-4	Dichloroethylene, 1,1-	Vol	9E-01	5E+00	1E+02	1E+00	6E+00	1E+02	2.3E+03
78-87-5	Dichloropropane, 1,2-	Vol	5E+00	3E+01	5E+02		same as adult		2.8E+03
10061-01-5	Dichloropropene, cis-1,3-	Vol	2E+00	1E+01	2E+02	2E+00	1E+01	2E+02	2.7E+03
10061-02-6	Dichloropropene, trans-1,3-	Vol	2E+00	1E+01	2E+02	3E+00	1E+01	2E+02	2.7E+03
57-97-6	Dimethylbenz(a)anthracene, 7,12-	Semi	7E+02 ^b	3E+03 ^b	6E+04 b	8E+02 b	4E+03 ^b	7E+04 b	2.5E-02
121-14-2	Dinitrotoluene, 2,4-	Semi	2E+02	1E+03 ^b	2E+04 b	2E+02	1E+03 ^b	2E+04 b	2.7E+02
123-91-1	Dioxane, 1,4-	Vol	1E+04	5E+04	no risk b		same as adult		1.0E+06
122-66-7	Diphenylhydrazine, 1,2-	Semi	8E+01 b	4E+02 b	7E+03 ^b	9E+01 b	4E+02 b	8E+03 b	6.8E+01
106-89-8	Epichlorohydrin	Vol	3E+02	2E+03	3E+04	3E+02	2E+03	3E+04	6.6E+04
106-88-7	Epoxybutane, 1,2-	Vol	4E+01	2E+02	4E+03		same as adult		4.3E+04
111-15-9	Ethoxyethanol acetate (R-R), 2-	Vol	9E+03	5E+04	9E+05		same as adult		1.0E+06
110-80-5	Ethoxyethanol, 2-	Vol	6E+04	3E+05	no risk b		same as adult		1.0E+06
100-41-4	Ethylbenzene	Vol	1E+03 ^a	7E+03 ^b	1E+05 b		same as adult		1.7E+02
106-93-4	Ethylene Dibromide	Vol	5E-01	3E+00	5E+01	6E-01	3E+00	6E+01	4.2E+03
107-21-1	Ethylene glycol	Semi	no risk b	no risk b	no risk b		same as adult		1.0E+06
75-21-8	Ethylene oxide	Vol	1E+00	7E+00	1E+02	2E+00	9E+00	2E+02	3.8E+05
50-00-0	Formaldehyde	Vol	5E+02 °	3E+03 °	5E+04 °	6E+02 °	3E+03 °	7E+04 °	5.5E+05
98-01-1	Furfural	Vol	2E+03	1E+04	2E+05 b		same as adult		1.1E+05
87-68-3	Hexachloro-1,3-butadiene	Semi	7E+00 a	4E+01 a	8E+02 b	9E+00 a	4E+01 b	8E+02 b	3.2E+00

Table 4-5. (continued)

			Adult Resident*	k	Cł	nild Resident 0-3	yr	Solubility
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	(mg/L)
118-74-1 Hexachlorobenzene	Semi	3E+00	2E+01 b	3E+02 b	4E+00	2E+01 b	3E+02 b	6.2E+00
77-47-4 Hexachlorocyclopentadiene	Semi	3E-01	2E+00	3E+01 b		same as adult		1.8E+00
67-72-1 Hexachloroethane	Vol	7E+01 b	3E+02 b	6E+03 b	7E+01 b	4E+02 b	7E+03 ^b	5.0E+01
78-59-1 Isophorone	Semi	3E+02	1E+03	3E+04 b		same as adult		1.2E+04
7439-92-1 Lead	Metal		NM			NM		0.0E+00
7439-96-5 Manganese	Metal		NM			NM		0.0E+00
7439-97-6 Mercury	Metal	5E-01 a	3E+00 b	5E+01 b		same as adult		5.6E-02
67-56-1 Methanol	Vol	2E+05	no risk b	no risk ^b		same as adult		1.0E+06
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	1E+03	6E+03	1E+05		same as adult		1.0E+06
109-86-4 Methoxyethanol, 2-	Vol	5E+03	2E+04	4E+05		same as adult		1.0E+06
74-83-9 Methyl bromide (Bromomethane)	Vol	5E+00	3E+01	5E+02		same as adult		1.5E+04
74-87-3 Methyl chloride (Chloromethane)	Vol	3E+01	2E+02	3E+03	4E+01	2E+02	4E+03	5.3E+03
78-93-3 Methyl ethyl ketone	Vol	5E+03	2E+04	5E+05 b		same as adult		2.2E+05
108-10-1 Methyl isobutyl ketone	Vol	3E+02	2E+03	3E+04 b		same as adult		1.9E+04
80-62-6 Methyl methacrylate	Vol	2E+03	1E+04	2E+05 ^b		same as adult		1.5E+04
1634-04-4 Methyl tert-butyl ether	Vol	7E+03	3E+04	6E+05 ^b		same as adult		3.9E+04
56-49-5 Methylcholanthrene, 3-	Semi	2E+02 b	9E+02 b	2E+04 b	3E+02 b	1E+03 b	2E+04 b	3.2E-03
75-09-2 Methylene chloride	Vol	1E+02	7E+02	1E+04	2E+02	8E+02	1E+04 ^a	1.3E+04
68-12-2 N,N-Dimethylformamide	Semi	2E+04	8E+04	no risk ^b		same as adult		1.0E+06
110-54-3 n-Hexane	Vol	3E+02 ^a	1E+03 ^a	3E+04 ^a		same as adult		1.2E+01
930-55-2 N-Nitrosopyrrolidine	Vol	4E+02	2E+03	4E+04	5E+02	2E+03	4E+04	1.0E+06
91-20-3 Naphthalene	Vol	1E+01	5E+01 b	1E+03 ^b		same as adult		3.1E+01
7440-02-0 Nickel	Metal		NM			NM		0.0E+00
98-95-3 Nitrobenzene	Semi	2E+01	1E+02	2E+03		same as adult		2.1E+03
79-46-9 Nitropropane, 2-	Vol	7E-02	4E-01	7E+00	8E-02	4E-01	8E+00	1.7E+04
924-16-3 Nitrosodi-n-butylamine	Vol	8E-02	4E-01	7E+00	1E-01	4E-01	8E+00	1.3E+03
55-18-5 Nitrosodiethylamine	Vol	2E-02	1E-01	2E+00	3E-02	1E-01	3E+00	9.3E+04
108-95-2 Phenol	Semi	2E+03	1E+04	2E+05 b		same as adult		8.3E+04

Table 4-5. (continued)

			Adult Resident*	•	С	hild Resident 0-3	yr	Solubility
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	(mg/L)
85-44-9 Phthalic anhydride	Semi	no risk	no risk ^b	no risk b		same as adult		6.2E+03
75-56-9 Propylene oxide	Vol	5E+01	3E+02	5E+03	6E+01	3E+02	6E+03	4.8E+05
110-86-1 Pyridine	Vol	1E+02	5E+02	1E+04		same as adult		1.0E+06
100-42-5 Styrene	Vol	2E+03 ^a	8E+03 b	1E+05 b		same as adult		3.1E+02
1746-01-6 TCDD, 2,3,7,8-	Semi	1E-03	6E-03 b	1E-01 b	1E-03	7E-03 ^b	1E-01	1.9E-05
630-20-6 Tetrachloroethane, 1,1,1,2-	Vol	1E+01	6E+01	1E+03	1E+01	7E+01	1E+03 ^a	1.1E+03
79-34-5 Tetrachloroethane, 1,1,2,2-	Vol	3E+00	1E+01	3E+02	3E+00	2E+01	3E+02	3.0E+03
127-18-4 Tetrachloroethylene	Vol	1E+02	5E+02 ^a	1E+04 b	1E+02	6E+02 ^a	1E+04 ^b	2.0E+02
108-88-3 Toluene	Vol	5E+02	2E+03 ^a	5E+04 b		same as adult		5.3E+02
95-53-4 Toluidine, o-	Semi	5E+01	2E+02	4E+03	5E+01	2E+02	5E+03	1.7E+04
76-13-1 Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol	3E+04 ^a	2E+05 b	no risk b		same as adult		1.7E+02
120-82-1 Trichlorobenzene, 1,2,4-	Semi	7E+02 b	3E+03 b	6E+04 b		same as adult		3.0E+02
71-55-6 Trichloroethane, 1,1,1-	Vol	1E+03	5E+03 ^a	1E+05 b		same as adult		1.3E+03
79-00-5 Trichloroethane, 1,1,2-	Vol	6E+00	3E+01	6E+02	7E+00	4E+01	7E+02	4.4E+03
79-01-6 Trichloroethylene	Vol	3E+01	2E+02	4E+03 ^a	4E+01	2E+02	4E+03 ^a	1.1E+03
75-69-4 Trichlorofluoromethane	Vol	6E+02	3E+03 ^a	7E+04 ^a		same as adult		1.1E+03
121-44-8 Triethylamine	Vol	3E+01	1E+02	2E+03		same as adult		5.5E+04
7440-62-2 Vanadium	Metal		NM			NM		0.0E+00
108-05-4 Vinyl acetate	Vol	5E+02	2E+03	4E+04 ^a		same as adult		2.0E+04
75-01-4 Vinyl chloride	Vol	5E-01	3E+00	5E+01	6E-01	3E+00	6E+01	2.8E+03
1330-20-7 Xylenes (total)	Vol	5E+02 ^a	3E+03 b	5E+04 b		same as adult		1.9E+02

NM = Not modeled for tanks.

^{*} For lead only, this column contains results for Child Resident 3-7 yrs.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

[°] Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

Table 4-6. Chronic 90/90 C_W at 25 m, Risk + $10^{-5}/HQ = 1$ for Storage Tanks (mg/L)

			Adult Resident*			Child Resident 0-3	3 yr	0 - 1 - 1 - 1114
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
75-07-0 Acetaldehyde	Vol	7E+02	5E+03	1E+05	8E+02	5E+03	1E+05	1.0E+06
67-64-1 Acetone	Vol	no risk b	no risk b	no risk		same as adult		1.0E+06
75-05-8 Acetonitrile	Vol	2E+03	2E+04	4E+05		same as adult		1.0E+06
107-02-8 Acrolein	Vol	6E-01	4E+00	9E+01		same as adult		2.1E+05
79-06-1 Acrylamide	Semi	2E+04	2E+05	no risk b	3E+04	2E+05	no risk	6.4E+05
79-10-7 Acrylic acid	Semi	3E+03	3E+04	5E+05		same as adult		1.0E+06
107-13-1 Acrylonitrile	Vol	2E+01	2E+02	3E+03	2E+01	2E+02	4E+03	7.4E+04
107-05-1 Allyl chloride	Vol	3E+01	2E+02	4E+03 ^a		same as adult		3.4E+03
62-53-3 Aniline	Semi	5E+02	3E+03	8E+04 b		same as adult		3.6E+04
7440-38-2 Arsenic	Metal		NM			NM		0.0E+00
7440-39-3 Barium	Metal		NM			NM		0.0E+00
71-43-2 Benzene	Vol	2E+02	1E+03	3E+04 ^a	2E+02	1E+03	3E+04 ^a	1.8E+03
92-87-5 Benzidine	Semi	3E+04 b	2E+05 b	no risk	3E+04 b	3E+05 b	no risk b	5.0E+02
50-32-8 Benzo(a)pyrene	Semi	9E+02 ^b	6E+03 b	1E+05 ^b	1E+03 b	6E+03 b	1E+05 ^b	2.5E-02
7440-41-7 Beryllium	Metal		NM			NM		0.0E+00
75-27-4 Bromodichloromethane	Vol	8E+01	6E+02	1E+04 ^a	9E+01	7E+02	1E+04 ^a	6.7E+03
75-25-2 Bromoform (Tribromomethane)	Vol	1E+03	1E+04 ^b	2E+05 b	2E+03	1E+04 ^b	3E+05 b	3.1E+03
106-99-0 Butadiene, 1,3-	Vol	3E+00	2E+01	4E+02	3E+00	2E+01	5E+02	7.4E+02
7440-43-9 Cadmium	Metal		NM			NM		0.0E+00
75-15-0 Carbon disulfide	Vol	2E+04 ^a	1E+05 ^a	no risk b		same as adult		1.2E+03
56-23-5 Carbon tetrachloride	Vol	1E+02	7E+02	2E+04 ^a	1E+02	9E+02 ^a	2E+04 ^a	7.9E+02
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	2E+02	1E+03	3E+04 ^a		same as adult		1.7E+03
108-90-7 Chlorobenzene	Vol	6E+02 ^a	5E+03 ^a	1E+05 b		same as adult		4.7E+02
124-48-1 Chlorodibromomethane	Vol	6E+01	4E+02	1E+04 b	7E+01	5E+02	1E+04 b	2.6E+03
67-66-3 Chloroform	Vol	6E+01	4E+02	1E+04 ^a	7E+01	5E+02	1E+04 ^a	7.9E+03
95-57-8 Chlorophenol, 2-	Semi	4E+01	3E+02	7E+03		same as adult		2.2E+04

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Table 4-6. (continued)

			Adult Resident*			Child Resident 0-3	3 yr	0 - 1 - 1 - 11 (
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
7440-47-3 Chromium VI	Metal		NM			NM		0.0E+00
7440-48-4 Cobalt	Metal		NM			NM		0.0E+00
1319-77-3 Cresols (total)	Semi	2E+02	1E+03	3E+04 b		same as adult		2.2E+04
98-82-8 Cumene	Vol	2E+04 b	1E+05 b	no risk		same as adult		6.1E+01
108-93-0 Cyclohexanol	Vol	4E+00	3E+01	6E+02		same as adult		3.6E+04
96-12-8 Dibromo-3-chloropropane, 1,2-	Vol	3E+03 b	2E+04 b	6E+05 b	4E+03 b	3E+04 b	6E+05 b	1.2E+03
95-50-1 Dichlorobenzene, 1,2-	Vol	7E+03 b	5E+04 b	no risk b		same as adult		1.6E+02
106-46-7 Dichlorobenzene, 1,4-	Vol	3E+04 ^b	2E+05 b	no risk b		same as adult		7.4E+01
75-71-8 Dichlorodifluoromethane	Vol	7E+03 ^a	5E+04 ^a	no risk b		same as adult		2.8E+02
107-06-2 Dichloroethane, 1,2-	Vol	5E+01	4E+02	9E+03 ^a	6E+01	5E+02	1E+04 ^a	8.5E+03
75-35-4 Dichloroethylene, 1,1-	Vol	3E+01	2E+02	4E+03 ^a	3E+01	2E+02	5E+03 ^a	2.3E+03
78-87-5 Dichloropropane, 1,2-	Vol	1E+02	9E+02	2E+04 a		same as adult		2.8E+03
10061-01-5 Dichloropropene, cis-1,3-	Vol	4E+01	3E+02	6E+03 ^a	4E+01	3E+02	7E+03 ^a	2.7E+03
10061-02-6 Dichloropropene, trans-1,3-	Vol	4E+01	3E+02	6E+03 ^a	4E+01	3E+02	7E+03 ^a	2.7E+03
57-97-6 Dimethylbenz(a)anthracene, 7,12-	Semi	3E+03 b	2E+04 b	5E+05 b	3E+03 b	2E+04 b	5E+05 ^b	2.5E-02
121-14-2 Dinitrotoluene, 2,4-	Semi	9E+02 b	7E+03 b	2E+05 b	1E+03 ^b	8E+03 b	2E+05 ^b	2.7E+02
123-91-1 Dioxane, 1,4-	Vol	7E+04	6E+05	no risk b		same as adult		1.0E+06
122-66-7 Diphenylhydrazine, 1,2-	Semi	4E+02 b	3E+03 b	7E+04 ^b	4E+02 b	3E+03 b	8E+04 b	6.8E+01
106-89-8 Epichlorohydrin	Vol	2E+03	1E+04	3E+05 b	2E+03	2E+04	4E+05 b	6.6E+04
106-88-7 Epoxybutane, 1,2-	Vol	6E+02	4E+03	9E+04 ^a		same as adult		4.3E+04
111-15-9 Ethoxyethanol acetate (R-R), 2-	Vol	5E+04	4E+05	no risk b		same as adult		1.0E+06
110-80-5 Ethoxyethanol, 2-	Vol	3E+05	no risk b	no risk b		same as adult		1.0E+06
100-41-4 Ethylbenzene	Vol	3E+04 b	3E+05 b	no risk		same as adult		1.7E+02
106-93-4 Ethylene Dibromide	Vol	6E+00	5E+01	1E+03	7E+00	5E+01	1E+03	4.2E+03
107-21-1 Ethylene glycol	Semi	no risk b	no risk ^b	no risk b		same as adult		1.0E+06
75-21-8 Ethylene oxide	Vol	1E+01	1E+02	2E+03	2E+01	1E+02	3E+03	3.8E+05
50-00-0 Formaldehyde	Vol	3E+03 °	2E+04 °	5E+05 °	3E+03 °	2E+04 °	5E+05 °	5.5E+05
98-01-1 Furfural	Vol	1E+04	9E+04	no risk		same as adult		1.1E+05
87-68-3 Hexachloro-1,3-butadiene	Semi	2E+02 a	1E+03 b	3E+04 b	2E+02 b	2E+03 b	4E+04 b	3.2E+00

Table 4-6. (continued)

			Adult Resident*			Child Resident 0-3	yr	0 - 1 - 1 - 11 - 11
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
118-74-1 Hexachlorobenzene	Semi	3E+01 b	2E+02 b	4E+03 b	3E+01 b	2E+02 ^b	5E+03 b	6.2E+00
77-47-4 Hexachlorocyclopentadiene	Semi	9E+00 a	7E+01 b	2E+03 b		same as adult		1.8E+00
67-72-1 Hexachloroethane	Vol	6E+02 b	5E+03 b	1E+05 b	7E+02 b	5E+03 b	1E+05 b	5.0E+01
78-59-1 Isophorone	Semi	2E+03	1E+04 b	3E+05 b		same as adult		1.2E+04
7439-92-1 Lead	Metal		NM			NM		0.0E+00
7439-96-5 Manganese	Metal		NM			NM		0.0E+00
7439-97-6 Mercury	Metal	1E+01 b	1E+02 b	2E+03 b		same as adult		5.6E-02
67-56-1 Methanol	Vol	no risk	no risk	no risk		same as adult		1.0E+06
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	7E+03	5E+04	no risk b		same as adult		1.0E+06
109-86-4 Methoxyethanol, 2-	Vol	2E+04	2E+05	no risk b		same as adult		1.0E+06
74-83-9 Methyl bromide (Bromomethane)	Vol	1E+02	9E+02	2E+04 a		same as adult		1.5E+04
74-87-3 Methyl chloride (Chloromethane)	Vol	9E+02	7E+03 ^a	2E+05 a	1E+03	8E+03 ^a	2E+05 ^a	5.3E+03
78-93-3 Methyl ethyl ketone	Vol	4E+04	3E+05 b	no risk b		same as adult		2.2E+05
108-10-1 Methyl isobutyl ketone	Vol	3E+03	2E+04 ^a	5E+05 ^b		same as adult		1.9E+04
80-62-6 Methyl methacrylate	Vol	2E+04 a	2E+05 b	no risk b		same as adult		1.5E+04
1634-04-4 Methyl tert-butyl ether	Vol	9E+04 ^a	6E+05 ^b	no risk b		same as adult		3.9E+04
56-49-5 Methylcholanthrene, 3-	Semi	9E+02 b	6E+03 ^b	1E+05 b	1E+03 b	7E+03 b	2E+05 b	3.2E-03
75-09-2 Methylene chloride	Vol	3E+03	2E+04 ^a	4E+05	3E+03	2E+04 ^a	5E+05 b	1.3E+04
68-12-2 N,N-Dimethylformamide	Semi	8E+04	6E+05	no risk b		same as adult		1.0E+06
110-54-3 n-Hexane	Vol	9E+03 ^a	6E+04 ^a	no risk b		same as adult		1.2E+01
930-55-2 N-Nitrosopyrrolidine	Vol	2E+03	1E+04	3E+05	2E+03	1E+04	3E+05	1.0E+06
91-20-3 Naphthalene	Vol	1E+02 b	8E+02 b	2E+04 b		same as adult		3.1E+01
7440-02-0 Nickel	Metal		NM			NM		0.0E+00
98-95-3 Nitrobenzene	Semi	1E+02	1E+03	2E+04 b		same as adult		2.1E+03
79-46-9 Nitropropane, 2-	Vol	6E-01	4E+00	1E+02	7E-01	5E+00	1E+02	1.7E+04
924-16-3 Nitrosodi-n-butylamine	Vol	1E+00	7E+00	2E+02	1E+00	8E+00	2E+02	1.3E+03
55-18-5 Nitrosodiethylamine	Vol	1E-01	1E+00	2E+01	1E-01	1E+00	3E+01	9.3E+04
108-95-2 Phenol	Semi	1E+04	9E+04 b	no risk		same as adult		8.3E+04

Table 4-6. (continued)

			Adult Resident*			Child Resident 0-3	yr	0 - 1 - 1 - 11/4
CAS Name	Group	25 m	150 m	1000 m	25 m	150 m	1000 m	Solubility (mg/L)
85-44-9 Phthalic anhydride	Semi	no risk b	no risk ^b	no risk b		same as adult		6.2E+03
75-56-9 Propylene oxide	Vol	5E+02	3E+03	7E+04	5E+02	4E+03	8E+04	4.8E+05
110-86-1 Pyridine	Vol	7E+02	5E+03	1E+05		same as adult		1.0E+06
100-42-5 Styrene	Vol	3E+04 b	2E+05 ^b	no risk b		same as adult		3.1E+02
1746-01-6 TCDD, 2,3,7,8-	Semi	6E-03 b	4E-02 b	9E-01 b	7E-03 ^b	4E-02 b	1E+00 ^b	1.9E-05
630-20-6 Tetrachloroethane, 1,1,1,2-	Vol	2E+02	2E+03 ^a	4E+04 b	2E+02	2E+03 ^a	4E+04 b	1.1E+03
79-34-5 Tetrachloroethane, 1,1,2,2-	Vol	3E+01	2E+02	5E+03 ^a	3E+01	2E+02	6E+03 ^a	3.0E+03
127-18-4 Tetrachloroethylene	Vol	3E+03 ^a	2E+04 ^a	4E+05	3E+03 ^a	2E+04 ^a	5E+05 b	2.0E+02
108-88-3 Toluene	Vol	1E+04 ^a	9E+04 b	no risk b		same as adult		5.3E+02
95-53-4 Toluidine, o-	Semi	2E+02	2E+03	4E+04 b	3E+02	2E+03	5E+04 b	1.7E+04
76-13-1 Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol	no risk b	no risk b	no risk b		same as adult		1.7E+02
120-82-1 Trichlorobenzene, 1,2,4-	Semi	9E+03 b	6E+04 b	no risk b		same as adult		3.0E+02
71-55-6 Trichloroethane, 1,1,1-	Vol	3E+04 ^a	2E+05 ^b	no risk b		same as adult		1.3E+03
79-00-5 Trichloroethane, 1,1,2-	Vol	9E+01	7E+02	2E+04 a	1E+02	8E+02	2E+04 ^a	4.4E+03
79-01-6 Trichloroethylene	Vol	9E+02	6E+03 ^a	1E+05 b	1E+03	7E+03 ^a	2E+05 ^b	1.1E+03
75-69-4 Trichlorofluoromethane	Vol	2E+04 ^a	1E+05 ^a	no risk b		same as adult		1.1E+03
121-44-8 Triethylamine	Vol	2E+02	2E+03	4E+04		same as adult		5.5E+04
7440-62-2 Vanadium	Metal		NM			NM		0.0E+00
108-05-4 Vinyl acetate	Vol	6E+03	4E+04 ^a	9E+05 ^b		same as adult		2.0E+04
75-01-4 Vinyl chloride	Vol	1E+01	1E+02	2E+03	2E+01	1E+02	3E+03	2.8E+03
1330-20-7 Xylenes (total)	Vol	1E+04 b	9E+04 b	no risk		same as adult		1.9E+02

NM = Not modeled for tanks.

^{*} For lead only, this column contains results for Child Resident 3-7 yrs.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

[°] Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

Table 4-7. Physical-Chemical Properties and Health Benchmarks

			Inhal CCE	D4C	шс	Solubility.	C
CAS	Name	Group	Inhal CSF (mg/kg-d)-1	RfC (mg/m³)	HLC (atm-m³/mol)	Solubility (mg/L)	C _{sat} (mg/kg)
75-07-0	Acetaldehyde	Vol	8E-03	9E-03	8E-05	1E+06	3E+05
67-64-1	Acetone	Vol	NA	3E+01	4E-05	1E+06	2E+05
75-05-8	Acetonitrile	Vol	NA	6E-02	3E-05	1E+06	2E+05
107-02-8	Acrolein	Vol	NA	2E-05	1E-04	2E+05	5E+04
79-06-1	Acrylamide	Semi	5E+00	NA	1E-09	6E+05	1E+05
79-10-7	Acrylic acid	Semi	NA	1E-03	1E-07	1E+06	2E+05
107-13-1	Acrylonitrile	Vol	2E-01	2E-03	1E-04	7E+04	2E+04
107-05-1	Allyl chloride	Vol	NA	1E-03	1E-02	3E+03	2E+03
62-53-3	Aniline	Semi	NA	1E-03	2E-06	4E+04	1E+04
7440-38-2	Arsenic	Metal	2E+01	NA	0	0	1E+06
7440-39-3	Barium	Metal	NA	5E-04	0	0	1E+06
71-43-2	Benzene	Vol	3E-02	NA	6E-03	2E+03	2E+03
92-87-5	Benzidine	Semi	2E+02	NA	4E-11	5E+02	2E+02
50-32-8	Benzo(a)pyrene	Semi	3E+00	NA	1E-06	3E-02	2E+02
7440-41-7	Beryllium	Metal	8E+00	2E-05	0	0	1E+06
75-27-4	Bromodichloromethane	Vol	6E-02	NA	2E-03	7E+03	6E+03
75-25-2	Bromoform (Tribromomethane)	Vol	4E-03	NA	5E-04	3E+03	5E+03
106-99-0	Butadiene, 1,3-	Vol	2E+00	NA	7E-02	7E+02	1E+03
7440-43-9	Cadmium	Metal	6E+00	NA	0	0	1E+06
75-15-0	Carbon disulfide	Vol	NA	7E-01	3E-02	1E+03	1E+03
56-23-5	Carbon tetrachloride	Vol	5E-02	NA	3E-02	8E+02	3E+03
126-99-8	Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	NA	7E-03	1E-02	2E+03	2E+03
108-90-7	Chlorobenzene	Vol	NA	2E-02	4E-03	5E+02	2E+03
124-48-1	Chlorodibromomethane	Vol	8E-02	NA	8E-04	3E+03	3E+03
67-66-3	Chloroform	Vol	8E-02	NA	4E-03	8E+03	6E+03
95-57-8	Chlorophenol, 2-	Semi	NA	1E-03	4E-04	2E+04	2E+04
7440-47-3	Chromium VI	Metal	4E+01	1E-04	0	0	1E+06
7440-48-4	Cobalt	Metal	NA	1E-05	0	0	1E+06
1319-77-3	Cresols (total)	Semi	NA	4E-04	2E-06	2E+04	5E+03
98-82-8	Cumene	Vol	NA	4E-01	1E+00	6E+01	2E+03
108-93-0	Cyclohexanol	Vol	NA	2E-05	4E-06	4E+04	2E+04
96-12-8	Dibromo-3-chloropropane, 1,2-	Vol	2E-03	2E-04	1E-04	1E+03	2E+03
95-50-1	Dichlorobenzene, 1,2-	Vol	NA	2E-01	2E-03	2E+02	2E+03
106-46-7	Dichlorobenzene, 1,4-	Vol	NA	8E-01	2E-03	7E+01	1E+03
75-71-8	Dichlorodifluoromethane	Vol	NA	2E-01	3E-01	3E+02	1E+03
107-06-2	Dichloroethane, 1,2-	Vol	9E-02	NA	1E-03	9E+03	3E+03
75-35-4	Dichloroethylene, 1,1-	Vol	2E-01	NA	3E-02	2E+03	3E+03
78-87-5	Dichloropropane, 1,2-	Vol	NA	4E-03	3E-03	3E+03	2E+03
10061-01-5	Dichloropropene, cis-1,3-	Vol	1E-01	2E-02	2E-03	3E+03	2E+03
10061-02-6	Dichloropropene, trans-1,3-	Vol	1E-01	2E-02	1E-03	3E+03	2E+03
57-97-6	Dimethylbenz(a)anthracene, 7,12-	Semi	8E+01	NA	3E-08	3E-02	5E+02
121-14-2	Dinitrotoluene, 2,4-	Semi	7E-01	NA	9E-08	3E+02	2E+02
123-91-1	Dioxane, 1,4-	Vol	NA	8E-01	5E-06	1E+06	2E+05
122-66-7	Diphenylhydrazine, 1,2-	Semi	8E-01	NA	2E-06	7E+01	3E+02
							continued)

Table 4-7. (continued)

				_			-
CAS	Name	Group	Inhal CSF (mg/kg-d)-1	RfC (mg/m³)	HLC (atm-m³/mol)	Solubility (mg/L)	C _{sat} (mg/kg)
106-89-8	Epichlorohydrin	Vol	4E-03	1E-03	3E-05	7E+04	2E+04
106-88-7	Epoxybutane, 1,2-	Vol	NA	2E-02	5E-04	4E+04	2E+04
111-15-9	Ethoxyethanol acetate (R-R), 2-	Vol	NA	3E-01	2E-06	1E+06	2E+05
110-80-5	Ethoxyethanol, 2-	Vol	NA	2E-01	4E-07	1E+06	2E+05
100-41-4	Ethylbenzene	Vol	NA	1E+00	8E-03	2E+02	1E+03
106-93-4	Ethylene Dibromide	Vol	8E-01	2E-04	7E-04	4E+03	3E+03
107-21-1	Ethylene glycol	Semi	NA	6E-01	6E-08	1E+06	2E+05
75-21-8	Ethylene oxide	Vol	4E-01	NA	1E-04	4E+05	9E+04
50-00-0	Formaldehyde	Vol	5E-02	NA	3E-07	6E+05	1E+05
98-01-1	Furfural	Vol	NA	5E-02	4E-06	1E+05	3E+04
87-68-3	Hexachloro-1,3-butadiene	Semi	8E-02	NA	8E-03	3E+00	1E+03
118-74-1	Hexachlorobenzene	Semi	2E+00	NA	1E-03	6E+00	2E+04
77-47-4	Hexachlorocyclopentadiene	Semi	NA	7E-05	3E-02	2E+00	2E+03
67-72-1	Hexachloroethane	Vol	1E-02	NA	4E-03	5E+01	3E+03
78-59-1	Isophorone	Semi	NA	1E-02	7E-06	1E+04	6E+03
7439-92-1	Lead	Metal	NA	NA	0	0	1E+06
7439-96-5	Manganese	Metal	NA	5E-05	0	0	1E+06
7439-97-6	Mercury	Metal	NA	3E-04	9E-03	6E-02	3E+01
67-56-1	Methanol	Vol	NA	1E+01	5E-06	1E+06	2E+05
110-49-6	Methoxyethanol acetate (R-R), 2-	Vol	NA	3E-02	2E-06	1E+06	2E+05
109-86-4	Methoxyethanol, 2-	Vol	NA	2E-02	3E-07	1E+06	2E+05
74-83-9	Methyl bromide (Bromomethane)	Vol	NA	5E-03	6E-03	2E+04	6E+03
74-87-3	Methyl chloride (Chloromethane)	Vol	6E-03	NA	9E-03	5E+03	2E+03
78-93-3	Methyl ethyl ketone	Vol	NA	1E+00	6E-05	2E+05	5E+04
108-10-1	Methyl isobutyl ketone	Vol	NA	8E-02	1E-04	2E+04	6E+03
80-62-6	Methyl methacrylate	Vol	NA	7E-01	3E-04	2E+04	6E+03
1634-04-4	Methyl tert-butyl ether	Vol	NA	3E+00	6E-04	4E+04	3E+04
56-49-5	Methylcholanthrene, 3-	Semi	7E+00	NA	9E-07	3E-03	4E+01
75-09-2	Methylene chloride	Vol	2E-03	3E+00	2E-03	1E+04	5E+03
68-12-2	N,N-Dimethylformamide	Semi	NA	3E-02	2E-07	1E+06	2E+05
91-20-3	Naphthalene	Vol	NA	3E-03	5E-04	3E+01	4E+02
110-54-3	n-Hexane	Vol	NA	2E-01	1E-02	1E+01	6E+02
7440-02-0	Nickel	Metal	8E-01	NA	0	0	1E+06
98-95-3	Nitrobenzene	Semi	NA	2E-03	2E-05	2E+03	1E+03
79-46-9	Nitropropane, 2-	Vol	9E+00	2E-02	1E-04	2E+04	5E+03
55-18-5	Nitrosodiethylamine	Vol	2E+02	NA	4E-06	9E+04	2E+04
924-16-3	Nitrosodi-n-butylamine	Vol	6E+00	NA	3E-04	1E+03	2E+03
930-55-2	N-Nitrosopyrrolidine	Vol	2E+00	NA	1E-08	1E+06	2E+05
108-95-2	Phenol	Semi	NA	6E-03	4E-07	8E+04	3E+04
	Phthalic anhydride	Semi	NA	1E-01	2E-08	6E+03	1E+03
75-56-9	·	Vol	1E-02	3E-02	9E-05	5E+05	1E+05
110-86-1	Pyridine	Vol	NA	7E-03	9E-06	1E+06	3E+05
100-42-5		Vol	NA	1E+00	3E-03	3E+02	2E+03
	TCDD, 2,3,7,8-	Semi	2E+05	NA	2E-05	2E-05	4E-01
							oontinued)

CAS	Name	Group	Inhal CSF (mg/kg-d)-1	RfC (mg/m³)	HLC (atm-m³/mol)	Solubility (mg/L)	C _{sat} (mg/kg)
630-20-6	Tetrachloroethane, 1,1,1,2-	Vol	3E-02	NA	2E-03	1E+03	3E+03
79-34-5	Tetrachloroethane, 1,1,2,2-	Vol	2E-01	NA	3E-04	3E+03	5E+03
127-18-4	Tetrachloroethylene	Vol	2E-03	3E-01	2E-02	2E+02	6E+02
108-88-3	Toluene	Vol	NA	4E-01	7E-03	5E+02	2E+03
95-53-4	Toluidine, o-	Semi	2E-01	NA	3E-06	2E+04	6E+03
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol	NA	3E+01	5E-01	2E+02	2E+03
120-82-1	Trichlorobenzene, 1,2,4-	Semi	NA	2E-01	1E-03	3E+02	2E+04
71-55-6	Trichloroethane, 1,1,1-	Vol	NA	1E+00	2E-02	1E+03	3E+03
79-00-5	Trichloroethane, 1,1,2-	Vol	6E-02	NA	9E-04	4E+03	4E+03
79-01-6	Trichloroethylene	Vol	6E-03	NA	1E-02	1E+03	3E+03
75-69-4	Trichlorofluoromethane	Vol	NA	7E-01	1E-01	1E+03	3E+03
121-44-8	Triethylamine	Vol	NA	7E-03	1E-04	6E+04	2E+04
7440-62-2	Vanadium	Metal	NA	7E-05	0	0	1E+06
108-05-4	Vinyl acetate	Vol	NA	2E-01	5E-04	2E+04	5E+03
75-01-4	Vinyl chloride	Vol	3E-01	NA	3E-02	3E+03	2E+03
1330-20-7	Xylenes (total)	Vol	NA	4E-01	6E-03	2E+02	2E+03

 $C_{\rm w}$'s that are greater than their solubility limit at a neutral pH and temperature of 20 to 25° C. As discussed in Section 3.2.4.3, the actual solubility of a constituent in a waste stream is dependent on site-specific conditions including pH , temperature, and the presence of other chemicals and metals. Therefore, the solubility limit will differ from the limit at a neutral pH and temperature of 20 to 25° C under different conditions. When a constituent exceeded its solubility limit at a neutral pH and temperature of 20 to 25° C, we modeled the constituent as pure component. Both 2,3,7,8-TCDD and mercury, when modeled as pure component did not exceed the target risk level of 10^{-5} or an HQ of 1.

4.3 Effect of WMU Type

The C_w 's from the different WMU types rank fairly consistently across chemicals. Treatment tanks (both aerated and nonaerated) typically result in the lowest C_w as a result of the aeration, which aids volatilization, or the generally large areas associated with nonaerated treatment tanks. Storage tanks and land application units (LAUs) are next, usually about two orders of magnitude higher; these

Ranking of C_w by WMU Type (low to high)

Aerated tank/Nonaerated tank
Storage tank/LAU
Landfill
Wastepile

are typically within an order of magnitude of each other. Landfills are next. These usually differ from LAUs by less than an order of magnitude. The wastepiles usually give the highest results, about an order of magnitude higher than the landfills.

Table 4-8 shows the number of chemicals for which each WMU type was most limiting. For most chemicals modeled in the tanks, aerated or nonaerated treatment tanks were the most

		Number of	Chemicals	
WMU Type	Volatiles	Semivolatiles	Metals	Total
All tanks	69	21	Not modeled	90
Aerated Treatment	52	6	Not modeled	58
Nonaerated Treatment	17	15	Not modeled	32
Storage	0	0	Not modeled	0
LAU	1	Not modeled	10	11
Landfill	0	Not modeled	0	0
Wastepile	0	Not modeled	0	0
No risk	1	2	0	3
Total	71	23	10	104

Table 4-8. Most Limiting WMU Type

limiting unit. The only exception is N-Nitrosopyrrolidene in LAUs. Three chemicals had no risk in any of the WMU types (see Section 4.2).

4.4 Effect of Exposure Factors

The effect of exposure factors can be seen by comparing results for different receptors. This is possible only for carcinogens, as exposure factors are not explicitly accounted for in the calculations for noncarcinogens. For all WMU types except landfills, adult resident receptors give the lowest results. This is because while child receptors have a higher inhalation rate relative to body weight than the adult, they are exposed for a shorter duration. The off-site worker is also exposed fewer days per year. For landfills, exposure duration is capped at 20 years, the assumed

Rank of Receptors (low to high)

Adult resident Child resident 0-3 yrs Child resident 4-10 yrs Child resident 11-18 yrs Offsite worker

operating life of the unit. This affects adults more significantly than children, so that adults are similar to children age 4-10 years in the ranking. The offsite worker gives higher results than the residents by less than an order of magnitude. The worker has a higher hourly inhalation rate than an adult resident but is exposed only 8 hours per day instead of 24.

4.5 Subchronic and Acute Results

This section presents the results for subchronic and acute exposures and compares them to the results for chronic exposures. Not all chemicals were included in the assessments for the subchronic and acute exposures due to limitations in available health benchmarks. Sixty-two chemicals were included in the assessment of subchronic exposures, and 35 chemicals were included in the assessment of acute exposures. All subchronic and acute results described in this section are for a resident; because acute and subchronic health benchmarks are expressed as an air concentration, exposure factors are not used, so there is no distinction between receptors. Waste concentration levels $(C_w$'s) are presented based on the most exposed individual at each site. From the cumulative distribution of this concentration across all sites, the 90^{th} percentile

was chosen; thus the results presented here protect 100 percent of the receptors at 90 percent of the sites at an HQ of 1 (100/90 levels). The detailed results that are summarized here, as well as results for additional distances, are presented in Volume III: *Results*.

Tables 4-9 and 4-10 show acute and subchronic results by distance for LAUs and wastepiles, respectively. The range of the most protective (i.e., lowest) $C_{\rm w}$ values for both WMUs was similar for subchronic, acute, and chronic. The results for subchronic and acute exposures show patterns similar to those for chronic exposures with respect to chemical properties, toxicity, and WMUs, except that toxicity was more important and there was less difference between the results for LAUs and wastepiles.

Tables 4-11 and 4-12 compare the subchronic and acute $C_{\rm w}$'s to the chronic 90/90 values for adults for all chemicals for LAUs and wastepiles, respectively. All values shown are for 75 m.

Several factors are likely to affect the subchronic and acute results differently than the chronic results:

- 1. Acute and subchronic health benchmarks were taken from several sources and so do not reflect a consistent methodology across chemicals.
- 2. The hazard posed by a chemical is likely to vary with exposure duration, i.e., some chemicals will have greater hazard at chronic exposures, others at acute and subchronic exposures.
- 3. Biodegradation is less likely to be an important factor for subchronic and acute exposures.

For both WMUs, the acute results, subchronic results, and chronic results show no pattern with respect to each other. Any of the three results may be lowest depending on the chemical, and the difference ranges from negligible up to two orders of magnitude. No pattern is apparent based on source of the toxicity benchmark or other physical-chemical properties affecting volatility or biodegradation.

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Table 4-9. Acute and Subchronic 100/90 C_w at 25 to 75 m for HQ = 1 for Land Application Units (mg/kg)

			Subchronic			Acute		
CAS Name	Group	25 m	50 m	75 m	25 m	50 m	75 m	C _{sat} (mg/kg)
75-07-0 Acetaldehyde	Vol	5E+03 °	6E+03 °	7E+03 °		NMA		3.3E+05
67-64-1 Acetone	Vol	no risk b,c	no risk b,c	no risk b,c	2E+05 °	3E+05 a,c	3E+05 a,c	2.3E+05
75-05-8 Acetonitrile	Vol	2E+03	3E+03	3E+03		NMA		2.3E+05
107-02-8 Acrolein	Vol	5E+00	6E+00	7E+00	1E-01	2E-01	2E-01	5.0E+04
79-06-1 Acrylamide	Semi		NM			NM		1.5E+05
79-10-7 Acrylic acid	Semi		NM			NM		2.4E+05
107-13-1 Acrylonitrile	Vol	5E+02	7E+02	8E+02	3E+02	4E+02	5E+02	1.8E+04
107-05-1 Allyl chloride	Vol	7E+01	1E+02	1E+02		NMA		1.7E+03
62-53-3 Aniline	Semi		NM			NM		1.0E+04
7440-38-2 Arsenic	Metal		NMS		6E+03	7E+03	8E+03	1.0E+06
7440-39-3 Barium	Metal	2E+05	3E+05	3E+05		NMA		1.0E+06
71-43-2 Benzene	Vol	1E+02	2E+02	2E+02	1E+02	2E+02	2E+02	1.8E+03
92-87-5 Benzidine	Semi		NM			NM		2.4E+02
50-32-8 Benzo(a)pyrene	Semi		NM			NM		1.5E+02
7440-41-7 Beryllium	Metal		NMS			NMA		1.0E+06
75-27-4 Bromodichloromethane	Vol		NMS			NMA		6.3E+03
75-25-2 Bromoform (Tribromomethane)	Vol		NMS			NMA		4.5E+03
106-99-0 Butadiene, 1,3-	Vol		NMS			NMA		1.1E+03
7440-43-9 Cadmium	Metal		NMS			NMA		1.0E+06
75-15-0 Carbon disulfide	Vol	5E+03 ^a	7E+03 ^a	9E+03 ^a	1E+04 ^a	1E+04 ^a	2E+04 ^a	1.3E+03
56-23-5 Carbon tetrachloride	Vol	2E+03	3E+03 ^a	4E+03 ^a	9E+02	1E+03	1E+03	2.7E+03
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol		NMS			NMA		1.8E+03
108-90-7 Chlorobenzene	Vol	3E+03 ^a	4E+03 a	4E+03 ^a		NMA		2.0E+03
124-48-1 Chlorodibromomethane	Vol		NMS			NMA		2.7E+03
67-66-3 Chloroform	Vol	2E+03	3E+03	3E+03	4E+02	5E+02	6E+02	5.7E+03
95-57-8 Chlorophenol, 2-	Semi		NM			NM		2.2E+04
7440-47-3 Chromium VI	Metal	2E+04	3E+04	3E+04		NMA		1.0E+06

Table 4-9. (continued)

				Subchronic			Acute		C
CAS	Name	Group	25 m	50 m	75 m	25 m	50 m	75 m	C _{sat} (mg/kg)
7440-48-4 Coba	lt	Metal	1E+03	2E+03	2E+03		NMA		1.0E+06
1319-77-3 Creso	ols (total)	Semi		NM			NM		5.2E+03
98-82-8 Cume	ene	Vol	8E+04 b	1E+05 b	1E+05 ^b		NMA		1.9E+03
108-93-0 Cyclo	hexanol	Vol	2E+01	2E+01	2E+01		NMA		1.6E+04
96-12-8 Dibro	mo-3-chloropropane, 1,2-	Vol	3E+02	4E+02	5E+02		NMA		1.8E+03
95-50-1 Dichlo	orobenzene, 1,2-	Vol	2E+05 ^b	3E+05 b	3E+05 b		NMA		2.2E+03
106-46-7 Dichlo	orobenzene, 1,4-	Vol	2E+05 ^b	3E+05 b	3E+05 b	2E+04 b	3E+04 b	3E+04 ^b	1.0E+03
75-71-8 Dichlo	orodifluoromethane	Vol	1E+04 ^a	2E+04 ^a	2E+04 ^a		NMA		1.2E+03
107-06-2 Dichlo	proethane, 1,2-	Vol	6E+03 ^a	1E+04 ^a	1E+04 ^a	6E+02	9E+02	1E+03	3.4E+03
75-35-4 Dichlo	oroethylene, 1,1-	Vol	6E+02	8E+02	1E+03		NMA		2.7E+03
78-87-5 Dichlo	oropropane, 1,2-	Vol	1E+02	1E+02	2E+02	2E+02	2E+02	3E+02	2.2E+03
10061-01-5 Dichlo	oropropene, cis-1,3-	Vol	4E+02	5E+02	6E+02		NMA		2.2E+03
10061-02-6 Dichlo	oropropene, trans-1,3-	Vol	4E+02	5E+02	7E+02		NMA		2.2E+03
57-97-6 Dimet	thylbenz(a)anthracene, 7,12-	Semi		NM			NM		4.8E+02
121-14-2 Dinitro	otoluene, 2,4-	Semi		NM			NM		2.1E+02
123-91-1 Dioxa	ine, 1,4-	Vol	3E+05 a,c	4E+05 a,c	5E+05 a,c	1E+04 °	1E+04 °	1E+04 °	2.3E+05
122-66-7 Diphe	enylhydrazine, 1,2-	Semi		NM			NM		3.3E+02
106-89-8 Epich	lorohydrin	Vol	6E+02	7E+02	8E+02	7E+03	1E+04	1E+04	1.6E+04
106-88-7 Epoxy	ybutane, 1,2-	Vol	4E+03	5E+03	6E+03		NMA		1.7E+04
111-15-9 Ethox	yethanol acetate (R-R), 2-	Vol		NMS		2E+03 °	3E+03 °	4E+03 °	2.3E+05
110-80-5 Ethox	xyethanol, 2-	Vol	8E+05 a,c	no risk ^{a,c}	no risk ^{a,c}	2E+04 °	3E+04 °	3E+04 °	2.3E+05
100-41-4 Ethylk	penzene	Vol	1E+04 ^a	2E+04 b	2E+04 ^b		NMA		1.3E+03
106-93-4 Ethyle	ene Dibromide	Vol	4E+01	6E+01	7E+01		NMA		3.1E+03
107-21-1 Ethyle	ene glycol	Semi		NM			NM		2.3E+05
75-21-8 Ethyle	ene oxide	Vol	5E+03 °	6E+03 °	7E+03 °		NMA		8.9E+04
50-00-0 Forma	aldehyde	Vol	4E+03 °	5E+03 °	7E+03 °	2E+03 °	2E+03 °	2E+03 °	1.3E+05
98-01-1 Furfu	ral	Vol	1E+04	2E+04	2E+04		NMA		2.7E+04
87-68-3 Hexa	chloro-1,3-butadiene	Semi		NM			NM		1.0E+03
118-74-1 Hexa	chlorobenzene	Semi		NM			NM		2.3E+04

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Table 4-9. (continued)

			Subchronic			Acute		•
CAS Name	Group	25 m	50 m	75 m	25 m	50 m	75 m	C _{sat} (mg/kg)
77-47-4 Hexachlorocyclopentadiene	Semi		NM			NM		2.1E+03
67-72-1 Hexachloroethane	Vol	no risk ^b	no risk b	no risk b	5E+05 b	8E+05 b	9E+05 ^b	2.6E+03
78-59-1 Isophorone	Semi		NM			NM		6.1E+03
7439-92-1 Lead	Metal		NMS			NMA		1.0E+06
7439-96-5 Manganese	Metal	2E+04	3E+04	3E+04		NMA		1.0E+06
7439-97-6 Mercury	Metal	9E+00	1E+01	1E+01	5E+00	7E+00	8E+00	2.6E+01
67-56-1 Methanol	Vol		NMS		2E+05 a,c	3E+05 a,c	4E+05 a,c	2.3E+05
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol		NMS			NMA		2.3E+05
109-86-4 Methoxyethanol, 2-	Vol	5E+04 °	7E+04 °	8E+04 °	5E+02 °	6E+02 °	7E+02 ^c	2.3E+05
74-83-9 Methyl bromide (Bromomethane)	Vol	1E+03	2E+03	2E+03	1E+02	2E+02	2E+02	5.7E+03
74-87-3 Methyl chloride (Chloromethane)	Vol	3E+03 ^a	4E+03 ^a	5E+03 ^a	6E+02	7E+02	7E+02	1.9E+03
78-93-3 Methyl ethyl ketone	Vol	6E+04 ^a	7E+04 ^a	9E+04 ^a	1E+05 ^a	1E+05 ^a	2E+05 ^a	5.4E+04
108-10-1 Methyl isobutyl ketone	Vol	5E+04 ^a	6E+04 ^a	7E+04 ^a		NMA		6.0E+03
80-62-6 Methyl methacrylate	Vol	2E+05 a	2E+05 b	3E+05 b		NMA		5.5E+03
1634-04-4 Methyl tert-butyl ether	Vol	2E+04	4E+04 a	5E+04 ^a	8E+03	9E+03	1E+04	2.6E+04
56-49-5 Methylcholanthrene, 3-	Semi		NM			NM		4.0E+01
75-09-2 Methylene chloride	Vol	2E+04 ^a	3E+04 ^a	4E+04 a	6E+03 ^a	9E+03 ^a	1E+04 ^a	4.6E+03
68-12-2 N,N-Dimethylformamide	Semi		NM			NM		2.3E+05
110-54-3 n-Hexane	Vol	3E+03 a	4E+03 a	5E+03 ^a		NMA		6.4E+02
930-55-2 N-Nitrosopyrrolidine	Vol		NMS			NMA		2.3E+05
91-20-3 Naphthalene	Vol	8E+03 ^b	1E+04 b	1E+04 ^b		NMA		3.8E+02
7440-02-0 Nickel	Metal	8E+04	1E+05	1E+05	1E+05	2E+05	2E+05	1.0E+06
98-95-3 Nitrobenzene	Semi		NM			NM		1.3E+03
79-46-9 Nitropropane, 2-	Vol	3E+02	4E+02	6E+02		NMA		4.6E+03
924-16-3 Nitrosodi-n-butylamine	Vol		NMS			NMA		2.1E+03
55-18-5 Nitrosodiethylamine	Vol		NMS			NMA		2.3E+04
108-95-2 Phenol	Semi		NM			NM		3.3E+04
85-44-9 Phthalic anhydride	Semi		NM			NM		1.4E+03
75-56-9 Propylene oxide	Vol	1E+03 °	1E+03 °	2E+03 °	1E+04 °	2E+04 °	2E+04 °	1.1E+05

Table 4-9. (continued)

				Subchronic Acute					
CAS	Name	Group	25 m	50 m	75 m	25 m	50 m	75 m	C _{sat} (mg/kg)
110-86-1	Pyridine	Vol		NMS			NMA		2.6E+05
100-42-5	Styrene	Vol	3E+05 ^b	3E+05 b	4E+05 ^b	1E+05 b	2E+05 b	2E+05 b	1.5E+03
1746-01-6	TCDD, 2,3,7,8-	Semi		NM			NM		3.8E-01
630-20-6	Tetrachloroethane, 1,1,1,2-	Vol		NMS			NMA		2.8E+03
79-34-5	Tetrachloroethane, 1,1,2,2-	Vol	8E+04 ^b	1E+05 b	1E+05 ^b		NMA		4.7E+03
127-18-4	Tetrachloroethylene	Vol	2E+04 b	3E+04 b	4E+04 b	1E+03 a	1E+03 ^a	1E+03 ^a	5.9E+02
108-88-3	Toluene	Vol	4E+04 ^b	6E+04 b	8E+04 ^b	2E+04 a	2E+04 ^a	2E+04 ^a	1.7E+03
95-53-4	Toluidine, o-	Semi		NM			NM		5.9E+03
76-13-1	Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol		NMS			NMA		2.1E+03
120-82-1	Trichlorobenzene, 1,2,4-	Semi		NM			NM		1.6E+04
71-55-6	Trichloroethane, 1,1,1-	Vol	3E+04 ^a	4E+04 ^a	5E+04 ^a	7E+03 ^a	9E+03 ^a	1E+04 ^a	2.7E+03
79-00-5	Trichloroethane, 1,1,2-	Vol		NMS			NMA		3.8E+03
79-01-6	Trichloroethylene	Vol	5E+03 ^a	6E+03 ^a	8E+03 ^a	1E+04 a	1E+04 ^a	1E+04 ^a	3.4E+03
75-69-4	Trichlorofluoromethane	Vol	5E+04 ^a	7E+04 ^a	9E+04 ^a		NMA		3.3E+03
121-44-8	Triethylamine	Vol	9E+02	1E+03	2E+03		NMA		2.1E+04
7440-62-2	Vanadium	Metal	3E+03	4E+03	4E+03	1E+04	1E+04	1E+04	1.0E+06
108-05-4	Vinyl acetate	Vol	5E+03	6E+03 ^a	7E+03 ^a		NMA		5.3E+03
75-01-4	Vinyl chloride	Vol	6E+02	7E+02	1E+03	7E+02	9E+02	1E+03	1.8E+03
1330-20-7	Xylenes (total)	Vol	5E+04 ^b	7E+04 ^b	9E+04 ^b	6E+03 ^a	7E+03 ^a	8E+03 ^a	1.5E+03

NM = Not modeled for land-based units.

NMA = Not modeled for acute.

NMS = Not modeled for subchronic.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25 °C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25 °C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^c Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

Table 4-10. Acute and Subchronic 100/90 C_W at 25 to 75 m for HQ = 1 for Wastepiles (mg/kg)

				Subchronic			Acute		75 m C _s
CAS	Name	Group	25 m	50 m	75 m	25 m	50 m	75m	(mg/kg)
75-07-0 Acetalo	dehyde	Vol	2E+03 °	3E+03 °	5E+03 °	_	NMA		3.3E+05
67-64-1 Aceton	ne	Vol	6E+05 a,c	no risk b,c	no risk b,c	8E+04 °	1E+05 °	2E+05 °	2.3E+05
75-05-8 Acetor	nitrile	Vol	1E+03	2E+03	3E+03		NMA		2.3E+05
107-02-8 Acrolei	in	Vol	3E+00	5E+00	7E+00	9E-02	1E-01	2E-01	5.0E+04
79-06-1 Acrylai	mide	Semi		NM			NM		1.5E+05
79-10-7 Acrylic	acid	Semi		NM			NM		2.4E+05
107-13-1 Acrylor	nitrile	Vol	3E+02	5E+02	7E+02	2E+02	3E+02	4E+02	1.8E+04
107-05-1 Allyl ch	nloride	Vol	3E+01	4E+01	6E+01		NMA		1.7E+03
62-53-3 Aniline	}	Semi		NM			NM		1.0E+04
7440-38-2 Arseni	С	Metal		NMS		2E+03	3E+03	3E+03	1.0E+06
7440-39-3 Barium	1	Metal	7E+04	1E+05	2E+05		NMA		1.0E+06
71-43-2 Benzei	ne	Vol	8E+01	1E+02	2E+02	6E+01	1E+02	1E+02	1.8E+03
92-87-5 Benzid	line	Semi		NM			NM		2.4E+02
50-32-8 Benzo	(a)pyrene	Semi		NM			NM		1.5E+02
7440-41-7 Berylliu	um	Metal		NMS			NMA		1.0E+06
75-27-4 Bromo	dichloromethane	Vol		NMS			NMA		6.3E+03
75-25-2 Bromo	form (Tribromomethane)	Vol		NMS			NMA		4.5E+03
106-99-0 Butadi	ene, 1,3-	Vol		NMS			NMA		1.1E+03
7440-43-9 Cadmi	um	Metal		NMS			NMA		1.0E+06
75-15-0 Carbor	n disulfide	Vol	2E+03 ^a	3E+03 ^a	4E+03 ^a	3E+03 ^a	5E+03 ^a	7E+03 ^a	1.3E+03
56-23-5 Carbor	n tetrachloride	Vol	2E+03	3E+03	4E+03 ^a	4E+02	6E+02	9E+02	2.7E+03
126-99-8 Chloro	-1,3-butadiene, 2- (Chloroprene)	Vol		NMS			NMA		1.8E+03
108-90-7 Chloro	benzene	Vol	3E+03 ^a	5E+03 ^a	8E+03 ^a		NMA		2.0E+03
124-48-1 Chloro	dibromomethane	Vol		NMS			NMA		2.7E+03
67-66-3 Chloro	form	Vol	1E+03	2E+03	4E+03	2E+02	3E+02	4E+02	5.7E+03
95-57-8 Chloro	phenol, 2-	Semi		NM			NM		2.2E+04
7440-47-3 Chrom	iium VI	Metal	7E+03	1E+04	2E+04		NMA		1.0E+06
7440-48-4 Cobalt		Metal	4E+02	7E+02	1E+03		NMA		1.0E+06

Table 4-10. (continued)

				Subchronic			Acute		75 m C _{sat}
CAS	Name	Group	25 m	50 m	75 m	25 m	50 m	75m	(mg/kg)
1319-77-3 Creso	ls (total)	Semi		NM			NM		5.2E+03
98-82-8 Cume	ne	Vol	1E+04 b	2E+04 b	3E+04 b		NMA		1.9E+03
108-93-0 Cycloh	nexanol	Vol	8E+00	1E+01	2E+01		NMA		1.6E+04
96-12-8 Dibron	no-3-chloropropane, 1,2-	Vol	2E+02	3E+02	4E+02		NMA		1.8E+03
95-50-1 Dichlo	robenzene, 1,2-	Vol	8E+04 b	1E+05 ^b	2E+05 ^b		NMA		2.2E+03
106-46-7 Dichlo	probenzene, 1,4-	Vol	9E+04 b	1E+05 ^b	2E+05 b	1E+04 b	2E+04 b	2E+04 b	1.0E+03
75-71-8 Dichlo	orodifluoromethane	Vol	3E+03 ^a	5E+03 ^a	8E+03 ^a		NMA		1.2E+03
107-06-2 Dichlo	roethane, 1,2-	Vol	7E+03 ^a	1E+04 ^a	2E+04 ^a	5E+02	6E+02	9E+02	3.4E+03
75-35-4 Dichlo	roethylene, 1,1-	Vol	2E+02	4E+02	6E+02		NMA		2.7E+03
78-87-5 Dichlo	ropropane, 1,2-	Vol	1E+02	2E+02	3E+02	1E+02	2E+02	2E+02	2.2E+03
10061-01-5 Dichlo	ropropene, cis-1,3-	Vol	2E+02	4E+02	6E+02		NMA		2.2E+03
10061-02-6 Dichlo	ropropene, trans-1,3-	Vol	3E+02	5E+02	7E+02		NMA		2.2E+03
57-97-6 Dimet	hylbenz(a)anthracene, 7,12-	Semi		NM			NM		4.8E+02
121-14-2 Dinitro	otoluene, 2,4-	Semi		NM			NM		2.1E+02
123-91-1 Dioxar	ne, 1,4-	Vol	3E+05 a,c	6E+05 a,c	8E+05 a,c	2E+04 °	2E+04 °	3E+04 ^c	2.3E+05
122-66-7 Dipher	nylhydrazine, 1,2-	Semi		NM			NM		3.3E+02
106-89-8 Epichl	orohydrin	Vol	3E+02	5E+02	8E+02	6E+03	9E+03	1E+04	1.6E+04
106-88-7 Epoxy	butane, 1,2-	Vol	2E+03	3E+03	5E+03		NMA		1.7E+04
111-15-9 Ethoxy	yethanol acetate (R-R), 2-	Vol		NMS		2E+03 °	2E+03 °	3E+03 °	2.3E+05
110-80-5 Ethoxy	yethanol, 2-	Vol	5E+05 a,c	9E+05 a,c	no risk ^{a,c}	1E+04 °	2E+04 °	3E+04 °	2.3E+05
100-41-4 Ethylb	enzene	Vol	1E+04 ^a	2E+04 b	3E+04 b		NMA		1.3E+03
106-93-4 Ethyle	ene dibromide	Vol	6E+01	1E+02	2E+02		NMA		3.1E+03
107-21-1 Ethyle	ne glycol	Semi		NM			NM		2.3E+05
75-21-8 Ethyle	ne oxide	Vol	2E+03 °	3E+03 °	5E+03 °		NMA		8.9E+04
50-00-0 Forma	aldehyde	Vol	2E+03 °	4E+03 °	6E+03 °	7E+02 °	1E+03 °	1E+03 °	1.3E+05
98-01-1 Furfur	al	Vol	5E+04 ^a	9E+04 ^a	1E+05 [°]		NMA		2.7E+04
87-68-3 Hexac	chloro-1,3-butadiene	Semi		NM			NM		1.0E+03
118-74-1 Hexac	chlorobenzene	Semi		NM			NM		2.3E+04
77-47-4 Hexac	chlorocyclopentadiene	Semi		NM			NM		2.1E+03
67-72-1 Hexac	chloroethane	Vol	no risk	no risk b	no risk b	no risk b	no risk b	no risk b	2.6E+03

Section 4.

Table 4-10. (continued)

				Subchronic			75 m C _{sat}		
CAS	Name	Group	25 m	50 m	75 m	25 m	50 m	75m	(mg/kg)
78-59-1 Isopho	orone	Semi		NM			NM		6.1E+03
7439-92-1 Lead		Metal		NMS		_	NMA		1.0E+06
7439-96-5 Manga	anese	Metal	7E+03	1E+04	2E+04		NMA		1.0E+06
7439-97-6 Mercu	ıry	Metal	4E+01 a	6E+01 ^a	9E+01 a	2E+01	2E+01	3E+01 b	2.6E+01
67-56-1 Metha	nol	Vol		NMS		1E+05 °	1E+05 °	2E+05 °	2.3E+05
110-49-6 Metho	exyethanol acetate (R-R), 2-	Vol		NMS			NMA		2.3E+05
109-86-4 Metho	xyethanol, 2-	Vol	6E+04 °	9E+04 °	1E+05 °	4E+02 °	5E+02 °	7E+02 °	2.3E+05
74-83-9 Methy	I bromide (Bromomethane)	Vol	7E+02	1E+03	2E+03	4E+01	6E+01	9E+01	5.7E+03
74-87-3 Methy	d chloride (Chloromethane)	Vol	9E+02	1E+03	2E+03 ^a	1E+02	2E+02	3E+02	1.9E+03
78-93-3 Methy	l ethyl ketone	Vol	2E+04	4E+04	6E+04 ^a	4E+04	7E+04 ^a	9E+04 ^a	5.4E+04
108-10-1 Methy	l isobutyl ketone	Vol	2E+04 ^a	3E+04 ^a	4E+04 ^a		NMA		6.0E+03
80-62-6 Methy	l methacrylate	Vol	1E+05 ^a	2E+05 b	3E+05 b		NMA		5.5E+03
1634-04-4 Methy	l tert-butyl ether	Vol	3E+04 ^a	5E+04 ^a	8E+04 ^a	6E+03	9E+03	1E+04	2.6E+04
56-49-5 Methy	rlcholanthrene, 3-	Semi		NM			NM		4.0E+01
75-09-2 Methy	lene chloride	Vol	2E+04 ^a	2E+04 ^a	4E+04 ^a	3E+03	5E+03 ^a	7E+03 ^a	4.6E+03
68-12-2 N,N-D	Dimethylformamide	Semi		NM			NM		2.3E+05
110-54-3 n-Hex	ane	Vol	4E+03 ^a	5E+03 ^a	8E+03 ^a		NMA		6.4E+02
930-55-2 N-Nitr	osopyrrolidine	Vol		NMS			NMA		2.3E+05
91-20-3 Napht	halene	Vol	2E+03 ^a	4E+03 ^a	6E+03 ^a		NMA		3.8E+02
7440-02-0 Nickel		Metal	3E+04	5E+04	7E+04	5E+04	6E+04	7E+04	1.0E+06
98-95-3 Nitrob	enzene	Semi		NM			NM		1.3E+03
79-46-9 Nitrop	ropane, 2-	Vol	4E+02	6E+02	9E+02		NMA		4.6E+03
924-16-3 Nitros	odi-n-butylamine	Vol		NMS			NMA		2.1E+03
55-18-5 Nitros	odiethylamine	Vol		NMS			NMA		2.3E+04
108-95-2 Pheno	ol	Semi		NM			NM		3.3E+04
85-44-9 Phtha	lic anhydride	Semi		NM			NM		1.4E+03
75-56-9 Propy	lene oxide	Vol	5E+02 °	9E+02 °	1E+03 °	6E+03 °	9E+03 °	1E+04 ^c	1.1E+05
110-86-1 Pyridi	ne	Vol		NMS			NMA		2.6E+05
100-42-5 Styrer	ne	Vol	6E+04 b	1E+05 ^b	2E+05 b	3E+04 b	4E+04 b	5E+04 ^b	1.5E+03

Table 4-10. (continued)

				Subchronic			Acute		75 m C _{sat}
CAS	Name	Group	25 m	50 m	75 m	25 m	50 m	75m	(mg/kg)
1746-01-6 TCDD	, 2,3,7,8-	Semi		NM			NM		3.8E-01
630-20-6 Tetrac	hloroethane, 1,1,1,2-	Vol		NMS			NMA		2.8E+03
79-34-5 Tetrac	hloroethane, 1,1,2,2-	Vol	9E+04 ^b	1E+05 b	2E+05 b		NMA		4.7E+03
127-18-4 Tetrac	hloroethylene	Vol	2E+04 b	3E+04 b	4E+04 b	6E+02	8E+02 ^a	1E+03 ^a	5.9E+02
108-88-3 Toluer	ne	Vol	4E+04 b	7E+04 ^b	9E+04 b	1E+04 ^a	1E+04 ^a	2E+04 ^a	1.7E+03
95-53-4 Toluid	ine, o-	Semi		NM			NM		5.9E+03
76-13-1 Trichlo	oro-1,2,2-trifluoroethane, 1,1,2-	Vol		NMS			NMA		2.1E+03
120-82-1 Trichlo	probenzene, 1,2,4-	Semi		NM			NM		1.6E+04
71-55-6 Trichlo	proethane, 1,1,1-	Vol	2E+04 ^a	3E+04 ^a	5E+04 b	4E+03 a	5E+03 ^a	8E+03 ^a	2.7E+03
79-00-5 Trichlo	proethane, 1,1,2-	Vol		NMS			NMA		3.8E+03
79-01-6 Trichlo	proethylene	Vol	4E+03 ^a	7E+03 ^a	1E+04 ^a	6E+03 ^a	8E+03 ^a	1E+04 ^a	3.4E+03
75-69-4 Trichlo	profluoromethane	Vol	2E+04 ^a	3E+04 ^a	5E+04 ^a		NMA		3.3E+03
121-44-8 Triethy	/lamine	Vol	2E+03	3E+03	4E+03		NMA		2.1E+04
7440-62-2 Vanac	ium	Metal	1E+03	2E+03	2E+03	3E+03	4E+03	5E+03	1.0E+06
108-05-4 Vinyl a	acetate	Vol	2E+03	3E+03	5E+03		NMA		5.3E+03
75-01-4 Vinyl o	chloride	Vol	1E+02	2E+02	4E+02	2E+02	2E+02	3E+02	1.8E+03
1330-20-7 Xylene	es (total)	Vol	5E+04 b	9E+04 ^b	1E+05 b	5E+03 ^a	7E+03 ^a	9E+03 ^a	1.5E+03

NM = Not modeled for land-based units.

NMA = Not modeled for acute.

NMS = Not modeled for subchronic.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25 °C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^c Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

 $\label{eq:comparison} \begin{array}{ll} Table \ 4\text{-}11. & Comparison \ of \ C_w \ for \ Chronic, Subchronic, and \ Acute \ Averaging \\ Times \ for \ HQ = 1 \ for \ Land \ Application \ Units \ (mg/kg) \end{array}$

CAS Name	Group	Chronic	Subchronic	Acute	C _{sat} (mg/kg)
75-07-0 Acetaldehyde	Vol	6E+03 ^a	7E+03 °	NM	3.3E+05
67-64-1 Acetone	Vol	no risk ^c	no risk b,c	3E+05 a,c	2.3E+05
75-05-8 Acetonitrile	Vol	1E+04	3E+03	NM	2.3E+05
107-02-8 Acrolein	Vol	2E+00	7E+00	2E-01	5.0E+04
79-06-1 Acrylamide	Semi	NM	NM	NM	1.5E+05
79-10-7 Acrylic acid	Semi	NM	NM	NM	2.4E+05
107-13-1 Acrylonitrile	Vol	9E+01	8E+02	5E+02	1.8E+04
107-05-1 Allyl chloride	Vol	2E+01	1E+02	NM	1.7E+03
62-53-3 Aniline	Semi	NM	NM	NM	1.0E+04
7440-38-2 Arsenic	Metal	2E+03	NM	8E+03	1.0E+06
7440-39-3 Barium	Metal	8E+04	3E+05	NM	1.0E+06
71-43-2 Benzene	Vol	3E+02	2E+02	2E+02	1.8E+03
92-87-5 Benzidine	Semi	NM	NM	NM	2.4E+02
50-32-8 Benzo(a)pyrene	Semi	NM	NM	NM	1.5E+02
7440-41-7 Beryllium	Metal	3E+03	NM	NM	1.0E+06
75-27-4 Bromodichloromethane	Vol	2E+02	NM	NM	6.3E+03
75-25-2 Bromoform (Tribromomethane)	Vol	1E+04 b	NM	NM	4.5E+03
106-99-0 Butadiene, 1,3-	Vol	2E+00	NM	NM	1.1E+03
7440-43-9 Cadmium	Metal	4E+03	NM	NM	1.0E+06
75-15-0 Carbon disulfide	Vol	2E+04 b	9E+03 ^a	2E+04 ^a	1.3E+03
56-23-5 Carbon tetrachloride	Vol	1E+02	4E+03 a	1E+03	2.7E+03
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	2E+02	NM	NM	1.8E+03
108-90-7 Chlorobenzene	Vol	1E+03	4E+03 a	NM	2.0E+03
124-48-1 Chlorodibromomethane	Vol	4E+02	NM	NM	2.7E+03
67-66-3 Chloroform	Vol	7E+01	3E+03	6E+02	5.7E+03
95-57-8 Chlorophenol, 2-	Semi	NM	NM	NM	2.2E+04
7440-47-3 Chromium VI	Metal	6E+02	3E+04	NM	1.0E+06
7440-48-4 Cobalt	Metal	2E+03	2E+03	NM	1.0E+06
1319-77-3 Cresols (total)	Semi	NM	NM .	NM	5.2E+03
98-82-8 Cumene	Vol	4E+04 °	1E+05 ^b	NM	1.9E+03
108-93-0 Cyclohexanol	Vol	8E+00	2E+01	NM	1.6E+04
96-12-8 Dibromo-3-chloropropane, 1,2-	Vol	8E+04 °	5E+02	NM	1.8E+03
95-50-1 Dichlorobenzene, 1,2-	Vol	1E+05 °	3E+05 b	NM	2.2E+03
106-46-7 Dichlorobenzene, 1,4-	Vol	4E+05 °	3E+05 b	3E+04 b	1.0E+03
75-71-8 Dichlorodifluoromethane	Vol	4E+03 b	2E+04 a	NM	1.2E+03
107-06-2 Dichloroethane, 1,2-	Vol	7E+01	1E+04 ^a	1E+03	3.4E+03
75-35-4 Dichloroethylene, 1,1-	Vol	3E+01	1E+03	NM	2.7E+03
78-87-5 Dichloropropane, 1,2-	Vol	1E+02	2E+02	3E+02	2.2E+03
10061-01-5 Dichloropropene, cis-1,3-	Vol	1E+02	6E+02	NM	2.2E+03
		•			(continued)

Table 4-11. (continued)

					C _{sat}
CAS Name	Group	Chronic	Subchronic	Acute	(mg/kg)
10061-02-6 Dichloropropene, trans-1,3-	Vol	1E+02	7E+02	NM	2.2E+03
57-97-6 Dimethylbenz(a)anthracene, 7,12-	Semi	NM	NM	NM	4.8E+02
121-14-2 Dinitrotoluene, 2,4-	Semi	NM	NM	NM	2.1E+02
123-91-1 Dioxane, 1,4-	Vol	2E+05 a	5E+05 a,c	1E+04 °	2.3E+05
122-66-7 Diphenylhydrazine, 1,2-	Semi	NM	NM	NM	3.3E+02
106-89-8 Epichlorohydrin	Vol	1E+04	8E+02	1E+04	1.6E+04
106-88-7 Epoxybutane, 1,2-	Vol	2E+03	6E+03	NM	1.7E+04
111-15-9 Ethoxyethanol acetate (R-R), 2-	Vol	2E+05 a	NM	4E+03 ^c	2.3E+05
110-80-5 Ethoxyethanol, 2-	Vol	5E+05 a,b	no risk a,c	3E+04 °	2.3E+05
100-41-4 Ethylbenzene	Vol	8E+04 °	2E+04 b	NM	1.3E+03
106-93-4 Ethylene Dibromide	Vol	3E+01	7E+01	NM	3.1E+03
107-21-1 Ethylene glycol	Semi	NM	NM	NM	2.3E+05
75-21-8 Ethylene oxide	Vol	6E+01 ^a	7E+03 °	NM	8.9E+04
50-00-0 Formaldehyde	Vol	8E+03 ^a	7E+03 °	2E+03 ^c	1.3E+05
98-01-1 Furfural	Vol	7E+03	2E+04	NM	2.7E+04
87-68-3 Hexachloro-1,3-butadiene	Semi	NM	NM	NM	1.0E+03
118-74-1 Hexachlorobenzene	Semi	NM	NM	NM	2.3E+04
77-47-4 Hexachlorocyclopentadiene	Semi	NM	NM	NM	2.1E+03
67-72-1 Hexachloroethane	Vol	1E+04 °	no risk b	9E+05 b	2.6E+03
78-59-1 Isophorone	Semi	NM	NM	NM	6.1E+03
7439-92-1 Lead*	Metal	3E+05	NM	NM	1.0E+06
7439-96-5 Manganese	Metal	8E+03	3E+04	NM	1.0E+06
7439-97-6 Mercury	Metal	3E+01 °	1E+01	8E+00	2.6E+01
67-56-1 Methanol	Vol	no risk ^c	NM	4E+05 a,c	2.3E+05
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	2E+04 a	NM	NM	2.3E+05
109-86-4 Methoxyethanol, 2-	Vol	3E+04 a	8E+04 °	7E+02 ^c	2.3E+05
74-83-9 Methyl bromide (Bromomethane)	Vol	2E+02	2E+03	2E+02	5.7E+03
74-87-3 Methyl chloride (Chloromethane)	Vol	7E+02	5E+03 ^a	7E+02	1.9E+03
78-93-3 Methyl ethyl ketone	Vol	4E+05 b	9E+04 ^a	2E+05 a	5.4E+04
108-10-1 Methyl isobutyl ketone	Vol	2E+04 ^b	7E+04 ^a	NM	6.0E+03
80-62-6 Methyl methacrylate	Vol	9E+04 b	3E+05 b	NM	5.5E+03
1634-04-4 Methyl tert-butyl ether	Vol	2E+05 b	5E+04 ^a	1E+04	2.6E+04
56-49-5 Methylcholanthrene, 3-	Semi	NM	NM	NM	4.0E+01
75-09-2 Methylene chloride	Vol	4E+03	4E+04 a	1E+04 a	4.6E+03
68-12-2 N,N-Dimethylformamide	Semi	NM	NM	NM	2.3E+05
110-54-3 n-Hexane	Vol	1E+04 b	5E+03 ^a	NM	6.4E+02
930-55-2 N-Nitrosopyrrolidine	Vol	3E+02 ^a	NM	NM	2.3E+05
91-20-3 Naphthalene	Vol	5E+03 b	1E+04 b	NM	3.8E+02
7440-02-0 Nickel	Metal	3E+04	1E+05	2E+05	1.0E+06
98-95-3 Nitrobenzene	Semi	NM	NM	NM	1.3E+03
79-46-9 Nitropropane, 2-	Vol	2E+00	6E+02	NM	4.6E+03

Table 4-11. (continued)

CAS Name	Gro	ıp C	hronic	Subchronic	Acute	C _{sat} (mg/kg)
924-16-3 Nitrosodi-n-butylamine	Vo	I 6	6E+00	NM	NM	2.1E+03
55-18-5 Nitrosodiethylamine	Vo	I	5E-01	NM	NM	2.3E+04
108-95-2 Phenol	Sen	ni	NM	NM	NM	3.3E+04
85-44-9 Phthalic anhydride	Sen		NM	NM	NM	1.4E+03
75-56-9 Propylene oxide	Vo	1 2	2E+03 ^a	2E+03 °	2E+04 °	1.1E+05
110-86-1 Pyridine	Vo	1 8	8E+03	NM	NM	2.6E+05
100-42-5 Styrene	Vo	l (5E+05 °	4E+05 b	2E+05 b	1.5E+03
1746-01-6 TCDD, 2,3,7,8-	Sen	ni	NM	NM	NM	3.8E-01
630-20-6 Tetrachloroethane, 1,1,1,2	- Vo	I 6	6E+02	NM	NM	2.8E+03
79-34-5 Tetrachloroethane, 1,1,2,2	- Vo	1 2	2E+02	1E+05 ^b	NM	4.7E+03
127-18-4 Tetrachloroethylene	Vo		3E+03 ^b	4E+04 ^b	1E+03 ^a	5.9E+02
108-88-3 Toluene	Vo	1 2	2E+04 ^c	8E+04 b	2E+04 ^a	1.7E+03
95-53-4 Toluidine, o-	Sen		NM	NM	NM	5.9E+03
76-13-1 Trichloro-1,2,2-trifluoroetha	ne, 1,1,2- Vo	1 8	8E+05 °	NM	NM	2.1E+03
120-82-1 Trichlorobenzene, 1,2,4-	Sen		NM	NM	NM	1.6E+04
71-55-6 Trichloroethane, 1,1,1-	Vo	I :	3E+04 ^b	5E+04 ^a	1E+04 ^a	2.7E+03
79-00-5 Trichloroethane, 1,1,2-	Vo		1E+02	NM	NM	3.8E+03
79-01-6 Trichloroethylene	Vo		1E+03	8E+03 ^a	1E+04 ^a	3.4E+03
75-69-4 Trichlorofluoromethane	Vo	1 2	2E+04 ^b	9E+04 ^a	NM	3.3E+03
121-44-8 Triethylamine	Vo		4E+02	2E+03	NM	2.1E+04
7440-62-2 Vanadium	Met	al '	1E+04	4E+03	1E+04	1.0E+06
108-05-4 Vinyl acetate	Vo	1 2	2E+04 ^b	7E+03 ^a	NM	5.3E+03
75-01-4 Vinyl chloride	Vo	1 '	1E+01	1E+03	1E+03	1.8E+03
1330-20-7 Xylenes (total)	Vo	l (3E+04 [°]	9E+04 b	8E+03 ^a	1.5E+03

no risk = the aqueous phase result exceeded 1 million parts per million (1,000,000 mg/kg or mg/L).

NM = Not modeled; if all 3 columns say NM, the chemical was not modeled for land-based unit. If only one or two columns say NM, the chemical was not modeled for that averaging time.

^{*} For lead, this value is Child aged 3-7 years.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^c Organic-phase emissions greater than aqueous-phase emissions; therefore, the C_w will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

Table 4-12. Comparison of $C_{\rm w}$ for Chronic, Subchronic, and Acute Averaging Times for HQ = 1 for Wastepiles (mg/kg)

			0. !		
CAS Name	Group	Chronic	Sub- chronic	Acute	C _{sat} (mg/kg)
75-07-0 Acetaldehyde	Vol	2E+04 a	5E+03 °	NM	3.3E+05
67-64-1 Acetone	Vol	no risk °	no risk b,c	2E+05 °	2.3E+05
75-05-8 Acetonitrile	Vol	5E+04	3E+03	NM	2.3E+05
107-02-8 Acrolein	Vol	1E+01	7E+00	2E-01	5.0E+04
79-06-1 Acrylamide	Semi	NM	NM	NM	1.5E+05
79-10-7 Acrylic acid	Semi	NM	NM	NM	2.4E+05
107-13-1 Acrylonitrile	Vol	4E+02	7E+02	4E+02	1.8E+04
107-05-1 Allyl chloride	Vol	9E+01	6E+01	NM	1.7E+03
62-53-3 Aniline	Semi	NM	NM	NM	1.0E+04
7440-38-2 Arsenic	Metal	2E+03	NM	3E+03	1.0E+06
7440-39-3 Barium	Metal	1E+05	2E+05	NM	1.0E+06
71-43-2 Benzene	Vol	2E+03	2E+02	1E+02	1.8E+03
92-87-5 Benzidine	Semi	NM	NM	NM	2.4E+02
50-32-8 Benzo(a)pyrene	Semi	NM	NM	NM	1.5E+02
7440-41-7 Beryllium	Metal	4E+03	NM	NM	1.0E+06
75-27-4 Bromodichloromethane	Vol	2E+03	NM	NM	6.3E+03
75-25-2 Bromoform (Tribromomethane)	Vol	3E+05 °	NM	NM	4.5E+03
106-99-0 Butadiene, 1,3-	Vol	4E+00	NM	NM	1.1E+03
7440-43-9 Cadmium	Metal	5E+03	NM	NM	1.0E+06
75-15-0 Carbon disulfide	Vol	5E+04 b	4E+03 a	7E+03 a	1.3E+03
56-23-5 Carbon tetrachloride	Vol	5E+02	4E+03 a	9E+02	2.7E+03
126-99-8 Chloro-1,3-butadiene, 2- (Chloroprene)	Vol	8E+02	NM	NM	1.8E+03
108-90-7 Chlorobenzene	Vol	1E+04 °	8E+03 ^a		2.0E+03
124-48-1 Chlorodibromomethane	Vol	6E+03 b	NM	NM	2.7E+03
67-66-3 Chloroform	Vol	4E+02	4E+03	4E+02	5.7E+03
95-57-8 Chlorophenol, 2-	Semi	NM	NM	NM	2.2E+04
7440-47-3 Chromium VI	Metal	8E+02	2E+04	NM	1.0E+06
7440-48-4 Cobalt	Metal	2E+03	1E+03	NM	1.0E+06
1319-77-3 Cresols (total)	Semi	NM	NM	NM	5.2E+03
98-82-8 Cumene	Vol	8E+04 °	3E+04 b	NM	1.9E+03
108-93-0 Cyclohexanol	Vol	4E+01	2E+01	NM	1.6E+04
96-12-8 Dibromo-3-chloropropane, 1,2-	Vol	4E+05 °	4E+02	NM	1.8E+03
95-50-1 Dichlorobenzene, 1,2-	Vol	5E+05 °	2E+05 b	NM	2.2E+03
106-46-7 Dichlorobenzene, 1,4-	Vol	no risk °	2E+05 b	2E+04 b	1.0E+03
75-71-8 Dichlorodifluoromethane	Vol	8E+03 b	8E+03 ^a	NM	1.2E+03
107-06-2 Dichloroethane, 1,2-	Vol	5E+02	2E+04 a	9E+02	3.4E+03
75-35-4 Dichloroethylene, 1,1-	Vol	9E+01	6E+02	NM	2.7E+03
78-87-5 Dichloropropane, 1,2-	Vol	9E+02	3E+02	2E+02	2.2E+03
10061-01-5 Dichloropropene, cis-1,3-	Vol	1E+03	6E+02	NM	2.2E+03
10061-02-6 Dichloropropene, trans-1,3-	Vol	1E+03	7E+02	NM	2.2E+03
					

Table 4-12. (continued)

			0		•
CAS Name	Group	Chronic	Sub- chronic	Acute	C _{sat} (mg/kg)
57-97-6 Dimethylbenz(a)anthracene, 7,12-	Semi	NM	NM	NM	4.8E+02
121-14-2 Dinitrotoluene, 2,4-	Semi	NM	NM	NM	2.1E+02
123-91-1 Dioxane, 1,4-	Vol	no risk a,b	8E+05 a,c	3E+04 ^c	2.3E+05
122-66-7 Diphenylhydrazine, 1,2-	Semi	NM	NM	NM	3.3E+02
106-89-8 Epichlorohydrin	Vol	5E+04 b	8E+02	1E+04	1.6E+04
106-88-7 Epoxybutane, 1,2-	Vol	1E+04	5E+03	NM	1.7E+04
111-15-9 Ethoxyethanol acetate (R-R), 2-	Vol	8E+05 a,b	NM	3E+03 °	2.3E+05
110-80-5 Ethoxyethanol, 2-	Vol	no risk °	no risk a,c	3E+04 °	2.3E+05
100-41-4 Ethylbenzene	Vol	no risk ^c	3E+04 b	NM	1.3E+03
106-93-4 Ethylene Dibromide	Vol	4E+02	2E+02	NM	3.1E+03
107-21-1 Ethylene glycol	Semi	NM	NM	NM	2.3E+05
75-21-8 Ethylene oxide	Vol	2E+02 a	5E+03 °	NM	8.9E+04
50-00-0 Formaldehyde	Vol	3E+04 a	6E+03 °	1E+03 °	1.3E+05
98-01-1 Furfural	Vol	1E+05 b	1E+05 a	NM	2.7E+04
87-68-3 Hexachloro-1,3-butadiene	Semi	NM	NM	NM	1.0E+03
118-74-1 Hexachlorobenzene	Semi	NM	NM	NM	2.3E+04
77-47-4 Hexachlorocyclopentadiene	Semi	NM	NM	NM	2.1E+03
67-72-1 Hexachloroethane	Vol	7E+05 °	no risk b	no risk b	2.6E+03
78-59-1 Isophorone	Semi	NM	NM	NM	6.1E+03
7439-92-1 Lead	Metal	3E+05	NM	NM	1.0E+06
7439-96-5 Manganese	Metal	1E+04	2E+04	NM	1.0E+06
7439-97-6 Mercury	Metal	5E+02°	9E+01 a	3E+01 b	2.6E+01
67-56-1 Methanol	Vol	no risk ^c	NM	2E+05 °	2.3E+05
110-49-6 Methoxyethanol acetate (R-R), 2-	Vol	9E+04 a	NM	NM	2.3E+05
109-86-4 Methoxyethanol, 2-	Vol	1E+05 a	1E+05 °	7E+02 °	2.3E+05
74-83-9 Methyl bromide (Bromomethane)	Vol	5E+02	2E+03	9E+01	5.7E+03
74-87-3 Methyl chloride (Chloromethane)	Vol	2E+03	2E+03 a	3E+02	1.9E+03
78-93-3 Methyl ethyl ketone	Vol	no risk ^c	6E+04 ^a	9E+04 a	5.4E+04
108-10-1 Methyl isobutyl ketone	Vol	1E+05 b	4E+04 a	NM	6.0E+03
80-62-6 Methyl methacrylate	Vol	5E+05 °	3E+05 b	NM	5.5E+03
1634-04-4 Methyl tert-butyl ether	Vol	no risk °	8E+04 a	1E+04	2.6E+04
56-49-5 Methylcholanthrene, 3-	Semi	NM	NM	NM	4.0E+01
75-09-2 Methylene chloride	Vol	2E+04 b	4E+04 a	7E+03 ^a	4.6E+03
68-12-2 N,N-Dimethylformamide	Semi	NM	NM	NM	2.3E+05
110-54-3 n-Hexane	Vol	1E+05 °	8E+03 ^a	NM	6.4E+02
930-55-2 N-Nitrosopyrrolidine	Vol	2E+03 a	NM	NM	2.3E+05
91-20-3 Naphthalene	Vol	2E+04 °	6E+03 ^a	NM	3.8E+02
7440-02-0 Nickel	Metal	4E+04	7E+04	7E+04	1.0E+06
98-95-3 Nitrobenzene	Semi	NM	NM	NM	1.3E+03
79-46-9 Nitropropane, 2-	Vol	2E+01	9E+02	NM	4.6E+03
					(continued)

Table 4-12. (continued)

			Sub-		C _{sat}
CAS Name	Group	Chronic	chronic	Acute	(mg/kg)
924-16-3 Nitrosodi-n-butylamine	Vol	4E+01	NM	NM	2.1E+03
55-18-5 Nitrosodiethylamine	Vol	3E+00	NM	NM	2.3E+04
108-95-2 Phenol	Semi	NM	NM	NM	3.3E+04
85-44-9 Phthalic anhydride	Semi	NM	NM	NM	1.4E+03
75-56-9 Propylene oxide	Vol	8E+03 ^a	1E+03 °	1E+04 °	1.1E+05
110-86-1 Pyridine	Vol	3E+04	NM	NM	2.6E+05
100-42-5 Styrene	Vol	no risk ^c	2E+05 b	5E+04 b	1.5E+03
1746-01-6 TCDD, 2,3,7,8-	Semi	NM	NM	NM	3.8E-01
630-20-6 Tetrachloroethane, 1,1,1,2-	Vol	5E+03 b	NM	NM	2.8E+03
79-34-5 Tetrachloroethane, 1,1,2,2-	Vol	2E+03	2E+05 b	NM	4.7E+03
127-18-4 Tetrachloroethylene	Vol	2E+04 °	4E+04 b	1E+03 a	5.9E+02
108-88-3 Toluene	Vol	2E+05 °	9E+04 b	2E+04 a	1.7E+03
95-53-4 Toluidine, o-	Semi	NM	NM	NM	5.9E+03
76-13-1 Trichloro-1,2,2-trifluoroethane, 1,1,2-	Vol	no risk °	NM	NM	2.1E+03
120-82-1 Trichlorobenzene, 1,2,4-	Semi	NM	NM	NM	1.6E+04
71-55-6 Trichloroethane, 1,1,1-	Vol	2E+05 °	5E+04 b	8E+03 a	2.7E+03
79-00-5 Trichloroethane, 1,1,2-	Vol	1E+03	NM	NM	3.8E+03
79-01-6 Trichloroethylene	Vol	7E+03 b	1E+04 a	1E+04 a	3.4E+03
75-69-4 Trichlorofluoromethane	Vol	5E+04 °	5E+04 ^a	NM	3.3E+03
121-44-8 Triethylamine	Vol	4E+03	4E+03	NM	2.1E+04
7440-62-2 Vanadium	Metal	1E+04	2E+03	5E+03	1.0E+06
108-05-4 Vinyl acetate	Vol	8E+04 b	5E+03	NM	5.3E+03
75-01-4 Vinyl chloride	Vol	3E+01	4E+02	3E+02	1.8E+03
1330-20-7 Xylenes (total)	Vol	3E+05 °	1E+05 b	9E+03 ^a	1.5E+03

no risk = the aqueous phase result exceeded 1 million parts per million (1,000,000 mg/kg or mg/L). NM = Not modeled; if all 3 columns say NM, the chemical was not modeled for land-based unit. If only one or two columns say NM, the chemical was not modeled for that averaging time.

^{*} For lead, this value is Child aged 3-7 years.

^a Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk exceeds 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

^b Aqueous-phase result exceeds solubility or C_{sat} at a neutral pH and temperature of 25°C. When modeled as a pure, organic-phase component, the risk is less than 1E-5 or HQ = 1. See Section 3.2.4.3 for more details.

 $^{^{\}rm c}$ Organic-phase emissions greater than aqueous-phase emissions; therefore, the C $_{\rm w}$ will be lower if this constituent is modeled in an organic waste matrix. The organic-phase results are shown in Volume III. See Section 3.2.4.3 for more details.

5.0 Characterization of Significant Findings

5.1 Introduction

The 1998 Air Characteristic Study began evaluating the merits of developing an air characteristic by integrating the results of the regulatory analysis, occurrence evaluation, and risk analysis. The purpose of combining these three analyses in the 1998 study was to provide preliminary information on where gaps in regulatory coverage may exist, what concentrations in waste are protective for direct inhalation risk, and, to a limited extent, the occurrence of these constituents in nonhazardous waste. Because integrating these analyses is an important step in identifying gaps and determining the significance of risk results, this section draws on the results of the revised risk assessment and the regulatory gaps analysis and occurrence analysis done for the May 1998 Air Characteristic Study.

This section presents an overview of the significant findings and begins to construct a comparison of the results to focus on potential gaps and the associated risks. This integration will assist in identifying areas warranting further study. This section serves two primary purposes: (1) to focus on the potential need to develop an air characteristic that would expand the present definition of hazardous waste, and (2) to evaluate the effectiveness of current RCRA regulations in reducing air emissions once a waste is defined as hazardous.

To address these dual purposes, Section 5.2 focuses on the extent to which the current RCRA hazardous waste characteristics and listings or CAA coverage already capture wastes that might pose hazardous air emissions, and Section 5.3 evaluates the extent to which current RCRA regulations would reduce the risks of air emissions from waste if an air characteristic were promulgated. This section examines the risks and gaps associated with wastewater tanks independently from the land-based waste management units. Results from this analysis indicate that the concentrations in tanks that would be protective of human health are considerably lower than the concentrations in waste that would be protective in the land-based units. The division between tanks and land-based units is also apparent in the level of direct air emission controls required.

This section focuses on a subset of the constituents addressed in the risk analysis. Constituents discussed in this chapter are primarily those that have protective concentrations in wastes of 100 ppm or less at a risk of 10⁻⁵ or HQ of 1 and a distance of 150 m from the site. **This is only for presentation purposes and does not exclude other risk levels, distances, or constituents from consideration in determining the need for an air characteristic.**

Aqueous-phase risk results presented in this section that exceed solubility or soil saturation concentrations under standard conditions (temperature of 20-25°C and neutral pH) are

footnoted; also noted is whether pure component modeled using organic phase produced risks or hazard quotients (HQs) over the target risk level of 10⁻⁵ or target HQ of 1. The difference between aqueous-phase and organic-phase emissions modeling is discussed in Sections 3 and 4.

5.2 Gaps in Constituent Coverage

In keeping with the charge of this study to determine the need for an air characteristic, the basis for constituent coverage under both RCRA and CAA was evaluated and compared to the constituents addressed in this study.

5.2.1 RCRA Gaps in Constituent Coverage

In RCRA, hazardous wastes are defined as solid wastes that exhibit a characteristic of hazardous waste or are listed as hazardous wastes. In recent years, the listing process has been based on multimedia considerations that include the investigation of air pathway risks. Characteristics more broadly define hazardous wastes because they are not limited to particular waste management processes. Currently, only one of the characteristics is based on constituent toxicity, the Toxicity Characteristic (TC). The TC was designed to identify wastes that pose significant potential risk through contamination of groundwater. The relationship of the constituents in this study to both the TC and listings was explored to ascertain the degree to which the 105 constituents may already be regulated as hazardous wastes.

An investigation of the TC list identified that 22 TC constituents overlap with the constituents in this study. A direct comparison was made between the milligram/liter TC level in the regulations to the waste concentration results

for wastewaters in tanks. For nonwastewaters, except metals, ¹ EPA used the Organic Leaching Model to derive a waste concentration in parts per million from the regulatory leachate level. This

Tank Waste Leve	ls That Are Mo (mg/L	0	than TC Level
Constituent	TC level	Aerated	Nonaerated

Constituent	TC level	Aerated	Nonaerated
Chlorobenzene Cresols, total	100 200	99	 160
Cicsois, total	200	_	100

Land-based Unit Waste Levels That Are More Stringent Than TC Level for Chronic Exposures (mg/kg)

Constituent	TC (Waste)	LAU Waste	WP Waste	LF Waste
Chlorobenzene	270,000	2,000ª	25,000 ^b	14,000°
1,1-Dichloroethylene	74	50	_	_

^a These values exceed the soil saturation concentration; pure component results in risk at or above 10⁻⁵.

^b These values exceed the soil saturation concentration; pure component results in risk below 10⁻⁵.

¹ Metals were calculated using empirical leaching data.

comparison, shown in Appendix A, indicates that the current TC levels are generally protective of potential direct inhalation risks.² In other words, the waste would already be hazardous and

LAU and WP Waste Levels That Are More Stringent Than TC Level for Subchronic Exposures

Constituent	TC (Waste)	LAU Waste	WP Waste	Scenario
Chlorobenzene	270,000	4,000 ^a	8,000 ^a	Subchronic

^a These values exceed the soil saturation concentration; pure component results in risk at or above 10⁻⁵.

potentially subject to air emission controls at a concentration lower than the concentration that would present a direct inhalation risk. Four constituents had protective waste concentrations lower than the TC or TC-derived waste concentration. Two constituents had TC levels that may not be protective of air pathway risks for tanks. Two constituents had waste concentrations more stringent than TC levels for land application units (LAUs); one of these had waste concentrations more stringent than TC levels in all the land-based units. The magnitude of the difference between the TC and the estimated air characteristic, C_w , varied according to the waste management unit and the constituent. In some circumstances, the difference was negligible, such as chlorobenzene in aerated tanks. The largest difference was less than a factor of 135 for chlorobenzene in LAUs.

A comparison was then made of the constituents that are associated with a hazardous waste listing. Listed wastes associated with constituents examined in this study were investigated using the best available national-level waste characterization data. Because of data

limitations, the listed waste code data provide a rough representation of the extent to which constituents in this study may be present in listed waste. The comparison was done to identify study constituents that did not appear to ever be associated with a listing or had very few listing associations. It can be argued that constituents that are associated with multiple listings are the least likely to present gaps. Sixteen of the study

Two Constituents Not Associated with a Listing or the TC with $C_{\rm w}s$ in Tanks Less than 100 ppm

1,3- Butadiene # Cyclohexanol

constituents were neither associated with a listing nor on the TC list. Two of these constituents had concentrations in all three types of tanks less than 100 ppm.

5.2.2 Clean Air Act Gaps in Constituent Coverage

Under the CAA, most direct regulation of air toxics is applied to those constituents on the list of hazardous air pollutants (HAPs). Of the 105 constituents addressed in this study, 84 are HAPs. The HAP list is strongly weighted to include the most toxic, volatile, and persistent of the chemicals known to be released to air from industrial operations. The CAA program controls HAPs under the authority of Section 112 for major sources that have been designated as source

² TC levels could change in the future; if a TC were significantly raised, it might no longer be protective of air risks.

categories. There is the possibility for a gap if the constituent is not on the HAP list. Twenty-one of the constituents addressed in this study are not on the HAP list. Of the 21 that are not HAPs, 10 had $C_{\rm w}$'s in tanks less than 100 mg/L and 4 had $C_{\rm w}$'s in land-based units less than 100 mg/kg. Of those that are on the HAP list, 25 had concentration levels ($C_{\rm w}$'s) in tanks less than 100 mg/L and 8 had $C_{\rm w}$'s less than 100 ppm in land-based units. The comparison of $C_{\rm w}$'s for the land units considered subchronic and acute exposures in addition to chronic exposures.

The potential exists for a constituent to be on the HAP list but not subject to CAA controls if it is contained in a waste that is managed at (1) a facility that is not an identified source category, (2) a facility that is in a source category but is not a major or area source, or (3) a facility that is in a source category and is a major source but the unit does not have a MACT standard. It would be difficult to determine the extent to which a constituent may be managed under one of

Listings, Not on the TC, and in Tanks Less th	nd Not a I	HAP with C _w 's
Constituent	<u>Cw</u>	
Nitrosodiethylamine	0.1	(nonaerated)
Bromodichloromethane	20	(aerated)
Cyclohexanol	3	(nonaerated)

20

(aerated)

0.3 (aerated)

these situations and not controlled under the MACT program.

5.2.3 Summary of Gaps in Regulation

It is apparent that the most significant gaps in regulatory coverage exist if the constituent is not a HAP, is not on the TC list, and is not associated with a listing. Out of the evaluated constituents, two were not regulated by any of these authorities: 3,4-dimethylphenol and cyclohexanol. The lowest $C_{\rm w}$ for cyclohexanol was 3 mg/L for nonaerated tanks. As noted previously, a $C_{\rm w}$ was not calculated for 3,4-dimethylphenol because of the lack of data appropriate for developing a health benchmark.

Chlorodibromomethane

Nitrosodi-*n*-butylamine

5.3 Air Emission Controls Under RCRA

This section evaluates the likelihood that air emissions from wastes subject to RCRA by an air characteristic would be reduced by current RCRA controls. Although there are provisions in RCRA for controlling air emissions for the purpose of reducing the resulting human health risk, qualifying factors place certain waste management units outside the scope of regulatory coverage.

5.3.1 Air Emission Controls for Tanks Under RCRA

It is apparent from the risk analysis that wastewater tanks are of most concern for risks from the inhalation pathway. This analysis focuses on the RCRA Subpart CC rules because they apply most directly to tanks. Subpart CC design requirements control air emissions from tanks, surface impoundments, and containers in which hazardous wastes have been placed. Current exclusions to the applicability of RCRA raise the potential that an air characteristic could

regulate wastes as hazardous, but current RCRA regulations may not reduce air emissions. Two of these exemptions are noteworthy in this chapter.

Foremost among the exclusions is the exemption of wastewater treatment tanks. Because available data indicate that wastewater treatment systems manage approximately 89 percent of industrial nonhazardous wastes, a large proportion of wastes that might be newly identified as hazardous by an air characteristic would escape regulation for air releases under RCRA Subpart CC rules (although some of these may be regulated by CAA requirements).

Another Subpart CC exclusion are wastes below the 500 ppm threshold for total organic content. Approximately 35 constituents were identified in the risk analysis as having protective waste concentrations in tanks of less than 100 ppm. Because the 500-ppm limit does not apply on a constituent-by-constituent basis and data are limited on the total volatiles that are typically co-managed, it is difficult to forecast the significance of this finding. Additionally, this 500-ppm threshold limit was established to be protective of human health and the environment.

Table 5-1 compares risk and occurrence data for the constituents that had a protective concentration in tanks of less than 100 ppm. This table indicates the protective waste concentration, if the constituent is a HAP, the TC level, if the TC level was protective, the number of listed waste associations, the total TRI releases and off-site transfers, and the estimated production volume. In addition, the table shows the number of two-digit SIC codes associated with the use of each constituent.

5.3.2 Air Emission Controls for Land Units Under RCRA

Air emissions from land-based units (i.e., landfills, land application units, wastepiles) are controlled directly or indirectly by land disposal restrictions and waste management unit standards. The protective concentrations calculated in the risk assessment are higher for land-based units than those for tanks, implying (without consideration of controls) that the risks associated with these units may be lower relative to those associated with tanks.

In general, the land disposal restriction treatment standards will greatly reduce the concentration of organics and will require the stabilization of metal-bearing wastes (reducing the wind-dispersal characteristics) of any regulated hazardous wastes

Land-based Unit Waste Levels That Are More Stringent Than UTS Level for Chronic Exposures (mg/kg)								
Constituent	UTS (Waste)	LAU Waste	WP Waste	LF Waste				
Nitrosodiethylamine Nitrosodi-n-butylamine	30 20	0.8 10	7	6				

placed in land-based units. A comparison between the LDRs and the protective concentrations in waste was made and indicates that the treatment standards are not always below the levels at which there are potential air risks. Two constituents had concentrations in waste for chronic exposures that were below the LDR treatment levels—one by more than tenfold. No constituents had concentrations in waste that were below the LDR treatment levels for acute or subchronic exposures. The tables contained in Appendix A show this comparison for each constituent.

Table 5-1. Regulatory, Occurrence, and Risk Comparison for Tanks— Constituents with Waste Concentrations Less than 100 ppm

CAS	Name	Waste Concentration (mg/kg)	Unit Type	Scenario	НАР	TC (leachate) Level (mg/kg)	Number of Listed Waste Associations	Total 1995 TRI Releases and Off-site Transfers (lb)	Estimated Annual Production (lb)	SIC Frequency
1746-01-6	TCDD, 2,3,7,8-	5.E-03 ^a	ATANK	Chronic	Н		11			
55-18-5	Nitrosodiethylamine	1.E-01	NTANK	Chronic			2			3
924-16-3	Nitrosodi-n-butylamine	3.E-01	ATANK	Chronic			2			2
107-02-8	Acrolein	3.E-01	ATANK	Chronic	Н		2	2.1E+05	6.1E+07	13
79-46-9	Nitropropane, 2-	4.E-01	ATANK	Chronic	Н		2	3.5E+04	5.0E+03	5
106-99-0	Butadiene, 1,3-	4.E-01	ATANK	Chronic	Н		0	1.0E+07	3.1E+09	7
77-47-4	Hexachlorocyclopentadiene	7.E-01	ATANK	Chronic	Н		9			
7439-97-6	Mercury	2.E+00 ^a	ATANK	Chronic	Н	0.2	24	3.0E+05	7.7E+06	12
75-01-4	Vinyl chloride	2.E+00	ATANK	Chronic	Н	2E-01	13	1.3E+06	1.4E+10	10
106-93-4	Ethylene Dibromide	2.E+00	ATANK	Chronic	Н		5	8.5E+04		7
108-93-0	Cyclohexanol	3.E+00	NTANK	Chronic			0		3.0E+03	5
75-35-4	Dichloroethylene, 1,1-	4.E+00	ATANK	Chronic		7E-01	8		6.7E+04	2
107-05-1	Allyl chloride	4.E+00	ATANK	Chronic	Н		3	4.7E+05	5.3E+05	1
118-74-1	Hexachlorobenzene	5.E+00	ATANK	Chronic	Н	0.13	12			
75-21-8	Ethylene oxide	7.E+00	NTANK	Chronic	Н		3	7.2E+04	5.7E+08	9
0061-01-5	Dichloropropene, cis-1,3-	8.E+00	ATANK	Chronic	Н		3		1.0E+06	3
0061-02-6	Dichloropropene, trans-1,3-	1.E+01	ATANK	Chronic	Н		3		1.0E+06	3
67-66-3	Chloroform	1.E+01	ATANK	Chronic	Н		27	1.3E+07	4.3E+07	10
79-34-5	Tetrachloroethane, 1,1,2,2-	1.E+01	ATANK	Chronic	Н		13	2.4E+06	1.0E+06	9
107-13-1	Acrylonitrile	1.E+01	NTANK	Chronic	Н		7	8.3E+06	3.5E+08	4
107-06-2	Dichloroethane, 1,2-	1.E+01	ATANK	Chronic		0.5	13	1.9E+07	1.0E+10	10
56-23-5	Carbon tetrachloride	1.E+01	ATANK	Chronic	Н	5E-01	18	1.6E+06	9.6E+06	13
95-57-8	Chlorophenol, 2-	2.E+01	ATANK	Chronic			4			

^a These values exceed the solubility; pure component results in risk below 10⁻⁵.

Table 5-2 compares risk and occurrence data for the constituents that had a protective concentration in tanks of less than 100 ppm. This table indicates the protective waste concentration, if the constituent is a HAP, the TC level, if the TC level was protective, the number of listed waste associations, the total TRI releases and off-site transfers, and the estimated production volume. In addition, the table shows the number of two-digit SIC codes associated with the use of each constituent.

Table 5-2. Regulatory, Occurrence, and Risk Comparison for Land-based Units— Constituents with Waste Concentrations Less than 100 ppm

CAS	Name	Waste Concentration (mg/kg)	Unit Type	Scenario	НАР	TC (waste) Level (mg/kg)	Number of Listed Waste Associations	Total 1995 TRI Releases and Off-site Transfers (lb)	Estimated Annual Production (lb)	SIC Frequency
107-02-8	Acrolein	2.E-01	WP	Acute	Н		2	2.1E+05	6.1E+07	13
55-18-5	Nitrosodiethylamine	8.E-01	LAU	Chronic			2			3
79-46-9	Nitropropane, 2-	3.E+00	LAU	Chronic	Н		4	3.5E+04	5.0E+03	5
106-99-0	Butadiene, 1,3-	4.E+00	LAU	Chronic	Н		0	1.0E+07	3.1E+09	7
7439-97-6	Mercury	8.E+00	LAU	Acute	Н	4	18	3.0E+05	7.7E+06	12
924-16-3	Nitrosodi-n-butylamine	1.E+01	LAU	Chronic			2			2
108-93-0	Cyclohexanol	2.E+01	LAU	Chronic			0		3.0E+03	5
75-01-4	Vinyl chloride	2.E+01	LAU	Chronic	Н	10	11	1.3E+06	1.4E+10	10
107-05-1	Allyl chloride	4.E+01	LAU	Chronic	Н		3	4.7E+05	5.3E+05	1
75-35-4	Dichloroethylene, 1,1-	5.E+01	LAU	Chronic		70	11		6.7E+04	2
106-93-4	Ethylene dibromide	6.E+01	LAU	Chronic	Н		5	8.5E+04		7
74-83-9	Methyl bromide (Bromomethane)	9.E+01	WP	Acute	Н		7	2.6E+06	4.5E+07	11

6.0 References

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Appendix A

Comparison with LDR Universal Treatment Standards and Toxicity Characteristic

Table A-1. Comparison of LDR Universal Treatment Standards with Chronic Waste Concentrations for Land-Based Units

Constituent Name	UTS Level (Nonwastewater) (mg/kg)	Wastepile Level (mg/kg)	Landfill Level (mg/kg)	Land Application Unit Level (mg/kg)
Acetaldehyde		6E+04 [°]	3E+04 ^c	9E+03 °
Acetone	2E+02	1E+07 b,c	1E+07 b,c	1E+07
Acetonitrile	4E+01	1E+05	6E+04	2E+04
Acrolein		2E+01	1E+01	4E+00
Acrylamide	2E+01			
Acrylic acid				
Acrylonitrile	8E+01	1E+03	6E+02	2E+02
Allyl chloride	3E+01	2E+02	1E+02	4E+01
Aniline	1E+01			
Arsenic	5E+00	6E+03	4E+04	3E+03
Barium	8E+00	3E+05	1E+07	1E+05
Benzene	1E+01	4E+03 ^a	2E+03 ^a	5E+02
Benzidine				
Benzo(a)pyrene	3E+00			
Beryllium	1E-02	1E+04	8E+04	5E+03
Bromodichloromethane	2E+01	5E+03	3E+03	4E+02
Bromoform (Tribromomethane)	2E+01	6E+05 ^b	2E+05 ^b	3E+04
Butadiene, 1,3-		1E+01	9E+00	4E+00
Cadmium	2E-01	1E+04	1E+05	7E+03
Carbon disulfide	5E+00	1E+05 ^b	9E+04 ^b	3E+04
Carbon tetrachloride	6E+00	1E+03	9E+02	2E+02
Chloro-1,3-butadiene, 2- (Chloroprene)	3E-01	2E+03 ^a	1E+03	3E+02
Chlorobenzene	6E+00	3E+04 ^b	1E+04 ^a	2E+03
Chlorodibromomethane	2E+01	1E+04 ^a	5E+03 ^a	7E+02
Chloroform	6E+00	9E+02	8E+02	1E+02
Chlorophenol, 2-	6E+00			
Chromium VI	9E-01	2E+03	2E+04	1E+03
Cobalt		5E+03	3E+04	3E+03
Cresols (total)	6E+00			
Cumene		2E+05 ^b	5E+04 ^b	7E+04
Cyclohexanol		1E+02	5E+01	2E+01
Dibromo-3-chloropropane, 1,2-	2E+01	9E+05 ^b	4E+05 ^b	1E+05
Dichlorobenzene, 1,2-	6E+00	1E+07 ^b	4E+05 ^b	2E+05
Dichlorobenzene, 1,4-	6E+00	1E+07 ^b	1E+07 ^b	7E+05
Dichlorodifluoromethane	7E+00	2E+04 ^a	2E+04 ^a	7E+03
Dichloroethane, 1,2-	6E+00	1E+03	1E+03	1E+02
Dichloroethylene, 1,1-	6E+00	2E+02	2E+02	5E+01
Dichloropropane, 1,2-	2E+01	2E+03 ^a	2E+03	2E+02

Table A-1. (continued)

Constituent Name	UTS Level (Nonwastewater) (mg/kg)	Wastepile Level (mg/kg)	Landfill Level (mg/kg)	Land Application Unit Level (mg/kg)
Dichloropropene, cis-1,3-	2E+01	2E+03 ^a	1E+03	3E+02
Dichloropropene, trans-1,3-	2E+01	3E+03 ^a	1E+03	3E+02
Dimethylbenz(a)anthracene, 7,12-				
Dimethylphenol, 3,4-				
Dinitrotoluene, 2,4-	1E+02			
Dioxane, 1,4-	2E+02	1E+07 b,c	1E+07 b,c	4E+05
Diphenylhydrazine, 1,2-				
Epichlorohydrin		1E+05 ^a	8E+04 ^a	2E+04
Epoxybutane, 1,2-		2E+04 ^a	1E+04	4E+03
Ethoxyethanol acetate (R-R), 2-		1E+07 b,c	1E+07 b,c	3E+05
Ethoxyethanol, 2-		1E+07 b,c	1E+07 b,c	8E+05
Ethylbenzene	1E+01	1E+07 ^b	7E+05 ^b	1E+05
Ethylene Dibromide	2E+01	1E+03	4E+02	6E+01
Ethylene glycol				
Ethylene oxide		6E+02 °	4E+02 °	1E+02
Formaldehyde		7E+04 ^c	4E+04 °	1E+04
Furfural		3E+05 ^a	3E+05 ^a	1E+04
Hexachloro-1,3-butadiene	6E+00			
Hexachlorobenzene	1E+01			
Hexachlorocyclopentadiene	2E+00			
Hexachloroethane	3E+01	1E+07 ^b	3E+05 ^b	2E+04
Isophorone				
Lead	4E-01	9E+05	1E+07	5E+05
Manganese		3E+04	1E+05	1E+04
Mercury	3E-01	1E+03 b	1E+03 ^b	5E+01
Methanol	8E-01	1E+07 b,c	1E+07 b,c	1E+07
Methoxyethanol acetate (R-R), 2-		2E+05 °	1E+05 °	4E+04
Methoxyethanol, 2-		4E+05 a,c	2E+05 °	5E+04
Methyl bromide (Bromomethane)	2E+01	1E+03	8E+02	3E+02
Methyl chloride (Chloromethane)	3E+01	4E+03 ^a	4E+03 ^a	1E+03
Methyl ethyl ketone	4E+01	1E+07 ^b	1E+07 b	6E+05
Methyl isobutyl ketone	3E+01	2E+05 ^b	9E+04 ^a	5E+04
Methyl methacrylate	2E+02	1E+07 b	6E+05 b	2E+05
Methyl tert-butyl ether		1E+07 ^b	1E+07 ^b	3E+05
Methylcholanthrene, 3-	2E+01			
Methylene chloride	3E+01	5E+04 ^a	3E+04 ^a	8E+03
N,N-Dimethylformamide				
Naphthalene	6E+00	4E+04 b	1E+04 ^b	8E+03
n-Hexane		2E+05 b	1E+05 ^b	2E+04

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Table A-1. (continued)

Constituent Name	UTS Level (Nonwastewater) (mg/kg)	Wastepile Level (mg/kg)	Landfill Level (mg/kg)	Land Application Unit Level (mg/kg)
Nickel	5E+00	1E+05	8E+05	5E+04
Nitrobenzene	1E+01			
Nitropropane, 2-		4E+01	2E+01	3E+00
Nitrosodiethylamine	3E+01	7E+00	6E+00	8E-01
Nitrosodi-n-butylamine	2E+01	1E+02	5E+01	1E+01
N-Nitrosopyrrolidine	4E+01	6E+03 °	6E+03 °	6E+02
Phenol	6E+00			
Phthalic anhydride	3E+01			
Propylene oxide		2E+04 ^c	1E+04 [°]	4E+03
Pyridine	2E+01	7E+04	2E+04	1E+04
Styrene		1E+07 ^b	9E+05 ^b	9E+05
TCDD, 2,3,7,8-	1E-03			
Tetrachloroethane, 1,1,1,2-	6E+00	1E+04 ^a	7E+03 ^a	1E+03
Tetrachloroethane, 1,1,2,2-	6E+00	6E+03 ^a	2E+03	4E+02
Tetrachloroethylene	6E+00	4E+04 ^b	3E+04 b	5E+03
Toluene	1E+01	4E+05 ^b	2E+05 ^b	4E+04
Toluidine, o-				
Trichloro-1,2,2-trifluoroethane, 1,1,2-	3E+01	1E+07 ^b	1E+07 ^b	1E+07
Trichlorobenzene, 1,2,4-	2E+01			
Trichloroethane, 1,1,1-	6E+00	4E+05 ^b	2E+05 ^b	5E+04
Trichloroethane, 1,1,2-	6E+00	3E+03	3E+03	2E+02
Trichloroethylene	6E+00	2E+04 ^b	1E+04 ^a	2E+03
Trichlorofluoromethane	3E+01	1E+05 ^b	9E+04 ^b	3E+04
Triethylamine		1E+04	8E+03	7E+02
Vanadium	2E-01	4E+04	2E+05	2E+04
Vinyl acetate		2E+05 ^b	8E+04 ^a	4E+04
Vinyl chloride	6E+00	7E+01	7E+01	2E+01
Xylenes (total)	3E+01	7E+05 ^b	3E+05 ^b	6E+04

^a Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component exceeds 1e-5 or

HQ = 1. See text for more details.

b Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component less than 1e-5 or HQ = 1. See text for more details.

^c Organic phase emissions greater than aqueous phase emissions, so result based on organic phase would be lower. See text & v. III for more details.

Table A-2. Comparison of Toxicity Characteristic Levels with Chronic Waste Concentrations for Land-Based Units

Constituent Name	TC Solids Level (mg/kg)	Wastepile Level (mg/kg)	Landfill Level (mg/kg)	Land Application Unit Level (mg/kg)
Acetaldehyde		6E+04 [°]	3E+04 [°]	9E+03 [°]
Acetone		1E+07 b,c	1E+07 b,c	1E+07 b,c
Acetonitrile		1E+05	6E+04	2E+04
Acrolein		2E+01	1E+01	4E+00
Acrylamide				
Acrylic acid				
Acrylonitrile		1E+03	6E+02	2E+02
Allyl chloride		2E+02	1E+02	4E+01
Aniline				
Arsenic	2E+02	6E+03	4E+04	3E+03
Barium	2E+03	3E+05	1E+07	1E+05
Benzene	5E+01	4E+03 ^a	2E+03 ^a	5E+02
Benzidine				
Benzo(a)pyrene				
Beryllium		1E+04	8E+04	5E+03
Bromodichloromethane		5E+03	3E+03	4E+02
Bromoform (Tribromomethane)		6E+05 ^b	2E+05 ^b	3E+04 ^a
Butadiene, 1,3-		1E+01	9E+00	4E+00
Cadmium		1E+04	1E+05	7E+03
Carbon disulfide		1E+05 ^b	9E+04 ^b	3E+04 ^a
Carbon tetrachloride	8E+01	1E+03	9E+02	2E+02
Chloro-1,3-butadiene, 2- (Chloroprene)		2E+03 ^a	1E+03	3E+02
Chlorobenzene	3E+05	3E+04 ^b	1E+04 ^a	2E+03 ^a
Chlorodibromomethane		1E+04 ^a	5E+03 ^a	7E+02
Chloroform		9E+02	8E+02	1E+02
Chlorophenol, 2-				
Chromium VI	1E+03	2E+03	2E+04	1E+03
Cobalt		5E+03	3E+04	3E+03
Cresols (total)	7E+04			
Cumene		2E+05 b	5E+04 ^b	7E+04 ^b
Cyclohexanol		1E+02	5E+01	2E+01
Dibromo-3-chloropropane, 1,2-		9E+05 b	4E+05 b	1E+05 ^b
Dichlorobenzene, 1,2-		1E+07 b	4E+05 b	2E+05 ^b
Dichlorobenzene, 1,4-	2E+04	1E+07 b	1E+07 ຶ	7E+05 ^b
Dichlorodifluoromethane		2E+04 ^a	2E+04 ^a	7E+03 ^a
Dichloroethane, 1,2-	2E+01	1E+03	1E+03	1E+02
Dichloroethylene, 1,1-	7E+01	2E+02	2E+02	5E+01
Dichloropropane, 1,2-		2E+03 ^a	2E+03	2E+02
Dichloropropene, cis-1,3-		2E+03 ^a	1E+03	3E+02
1 1 -7 7-				(continued

Table A-2. (continued)

Constituent Name	TC Solids Level (mg/kg)	Wastepile Level (mg/kg)	Landfill Level (mg/kg)	Land Application Unit Level (mg/kg)
Dichloropropene, trans-1,3-		3E+03 ^a	1E+03	3E+02
Dimethylbenz(a)anthracene, 7,12-				
Dimethylphenol, 3,4-				
Dinitrotoluene, 2,4-	8E+00			
Dioxane, 1,4-		1E+07 b,c	1E+07 b,c	4E+05 a,c
Diphenylhydrazine, 1,2-				
Epichlorohydrin		1E+05 ^a	8E+04 ^a	2E+04 ^a
Epoxybutane, 1,2-		2E+04 ^a	1E+04	4E+03
Ethoxyethanol acetate (R-R), 2-		1E+07 b,c	1E+07 b,c	3E+05 a,c
Ethoxyethanol, 2-		1E+07 b,c	1E+07 b,c	8E+05 a,c
Ethylbenzene		1E+07 ^b	7E+05 ^b	1E+05 ^b
Ethylene Dibromide		1E+03	4E+02	6E+01
Ethylene glycol				
Ethylene oxide		6E+02 °	4E+02 °	1E+02 °
Formaldehyde		7E+04 °	4E+04 °	1E+04 °
Furfural		3E+05 ^a	3E+05 ^a	1E+04
Hexachloro-1,3-butadiene	9E+03			
Hexachlorobenzene	7E+03			
Hexachlorocyclopentadiene				
Hexachloroethane	5E+03	1E+07 ^b	3E+05 b	2E+04 b
Isophorone				
Lead	2E+03	9E+05	1E+07	5E+05
Manganese		3E+04	1E+05	1E+04
Mercury	4E+00	1E+03 ^b	1E+03 ^b	5E+01 ^b
Methanol		1E+07 b,c	1E+07 b,c	1E+07 b,c
Methoxyethanol acetate (R-R), 2-		2E+05 °	1E+05 °	4E+04 ^c
Methoxyethanol, 2-		4E+05 a,c	2E+05 °	5E+04 °
Methyl bromide (Bromomethane)		1E+03	8E+02	3E+02
Methyl chloride (Chloromethane)		4E+03 ^a	4E+03 ^a	1E+03
Methyl ethyl ketone	2E+04	1E+07 ^b	1E+07 ^b	6E+05 ^a
Methyl isobutyl ketone		2E+05 ^b	9E+04 ^a	5E+04 ^a
Methyl methacrylate		1E+07 ^b	6E+05 b	2E+05 ^b
Methyl tert-butyl ether		1E+07 ^b	1E+07 ^b	3E+05 ^b
Methylcholanthrene, 3-				
Methylene chloride		5E+04 ^a	3E+04 ^a	8E+03 ^a
N,N-Dimethylformamide				
Naphthalene		4E+04 b	1E+04 b	8E+03 ^b
n-Hexane		2E+05 b	1E+05 b	2E+04 ^b
Nickel		1E+05	8E+05	5E+04
Nitrobenzene	4E+02			

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Table A-2. (continued)

Constituent Name	TC Solids Level (mg/kg)	Wastepile Level (mg/kg)	Landfill Level (mg/kg)	Land Application Unit Level (mg/kg)
Nitropropane, 2-		4E+01	2E+01	3E+00
Nitrosodiethylamine		7E+00	6E+00	8E-01
Nitrosodi-n-butylamine		1E+02	5E+01	1E+01
N-Nitrosopyrrolidine		6E+03 °	6E+03 °	6E+02 °
Phenol				
Phthalic anhydride				
Propylene oxide		2E+04 °	1E+04 [°]	4E+03 °
Pyridine	3E+02	7E+04	2E+04	1E+04
Styrene		1E+07 ^b	9E+05 ^b	9E+05 ^b
TCDD, 2,3,7,8-				
Tetrachloroethane, 1,1,1,2-		1E+04 ^a	7E+03 ^a	1E+03
Tetrachloroethane, 1,1,2,2-		6E+03 ^a	2E+03	4E+02
Tetrachloroethylene	3E+02	4E+04 ^b	3E+04 ^b	5E+03 ^a
Toluene		4E+05 ^b	2E+05 ^b	4E+04 ^b
Toluidine, o-				
Trichloro-1,2,2-trifluoroethane, 1,1,2-		1E+07 ^b	1E+07 ^b	1E+07 ^b
Trichlorobenzene, 1,2,4-				
Trichloroethane, 1,1,1-		4E+05 ^b	2E+05 ^b	5E+04 ^b
Trichloroethane, 1,1,2-		3E+03	3E+03	2E+02
Trichloroethylene	7E+01	2E+04 ^b	1E+04 ^a	2E+03
Trichlorofluoromethane		1E+05 ^b	9E+04 ^b	3E+04 ^a
Triethylamine		1E+04	8E+03	7E+02
Vanadium		4E+04	2E+05	2E+04
Vinyl acetate		2E+05 ^b	8E+04 ^a	4E+04 ^a
Vinyl chloride	1E+01	7E+01	7E+01	2E+01
Xylenes (total)		7E+05 ^b	3E+05 ^b	6E+04 ^b

^a Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component exceeds 1e-5 or HQ = 1. See text for more details.

b Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component less than 1e-5 or

 $H\dot{Q} = 1$. See text for more details.

^c Organic phase emissions greater than aqueous phase emissions, so result based on organic phase would be lower. See text & v. III for more details.

Table A-3. Comparison of LDR Universal Treatment Standards (UTS) with Acute and Subchronic Waste Concentrations for Land Application Units and Wastepiles

Constituent	UTS Level (Nonwastewater) (mg/kg)	Land Application Unit Level (mg/kg)	Scenario	Wastepile Level (mg/kg)	Scenario
Acetaldehyde		7E+03 °	Subchronic	5E+03 °	Subchronic
Acetone	2E+02	3E+05 a,c	Acute	2E+05 °	Acute
Acetonitrile	4E+01	3E+03	Subchronic	3E+03	Subchronic
Acrolein		2E-01	Acute	2E-01	Acute
Acrylamide	2E+01				
Acrylic acid					
Acrylonitrile	8E+01	5E+02	Acute	4E+02	Acute
Allyl chloride	3E+01	1E+02	Subchronic	6E+01	Subchronic
Aniline	1E+01				
Arsenic	5E+00	8E+03	Acute	3E+03	Acute
Barium	8E+00	3E+05	Subchronic	2E+05	Subchronic
Benzene	1E+01	2E+02	Acute	1E+02	Acute
Benzidine					
Benzo(a)pyrene	3E+00				
Beryllium	1E-02				
Bromodichloromethane	2E+01				
Bromoform (Tribromomethane)	2E+01				
Butadiene, 1,3-					
Cadmium	2E-01				
Carbon disulfide	5E+00	9E+03 ^a	Subchronic	4E+03 ^a	Subchronic
Carbon tetrachloride	6E+00	1E+03	Acute	9E+02	Acute
Chloro-1,3-butadiene, 2- (Chloroprene)	3E-01				
Chlorobenzene	6E+00	4E+03 ^a	Subchronic	8E+03 ^a	Subchronic
Chlorodibromomethane	2E+01				
Chloroform	6E+00	6E+02	Acute	4E+02	Acute
Chlorophenol, 2-	6E+00				
Chromium VI	9E-01	3E+04	Subchronic	2E+04	Subchronic
Cobalt		2E+03	Subchronic	1E+03	Subchronic
Cresols (total)	6E+00				
Cumene		1E+05 ^b	Subchronic	3E+04 b	Subchronic
Cyclohexanol		2E+01	Subchronic	2E+01	Subchronic
Dibromo-3-chloropropane, 1,2-	2E+01	5E+02	Subchronic	4E+02	Subchronic
Dichlorobenzene, 1,2-	6E+00	3E+05 ^b	Subchronic	2E+05 b	Subchronic
Dichlorobenzene, 1,4-	6E+00	3E+04 ^b	Acute	2E+04 ^⁵	Acute
Dichlorodifluoromethane	7E+00	2E+04 ^a	Subchronic	8E+03 ^a	Subchronic
Dichloroethane, 1,2-	6E+00	1E+03	Acute	9E+02	Acute
Dichloroethylene, 1,1-	6E+00	1E+03	Subchronic	6E+02	Subchronic
Dichloropropane, 1,2-	2E+01	2E+02	Subchronic	2E+02	Acute
Dichloropropene, cis-1,3-	2E+01	6E+02	Subchronic	6E+02	Subchronic
Dichloropropene, trans-1,3-	2E+01	7E+02	Subchronic	7E+02	Subchronic
Dimethylbenz(a)anthracene, 7,12-					

Table A-3. (continued)

Constituent	UTS Level (Nonwastewater) (mg/kg)	Land Application Unit Level (mg/kg)	Scenario	Wastepile Level (mg/kg)	Scenario
Dimethylphenol, 3,4-					
Dinitrotoluene, 2,4-	1E+02				
Dioxane, 1,4-	2E+02	1E+04 °	Acute	3E+04 °	Acute
Diphenylhydrazine, 1,2-					
Epichlorohydrin		8E+02	Subchronic	8E+02	Subchronic
Epoxybutane, 1,2-		6E+03	Subchronic	5E+03	Subchronic
Ethoxyethanol acetate (R-R), 2-		4E+03 °	Acute	3E+03 °	Acute
Ethoxyethanol, 2-		3E+04 °	Acute	3E+04 °	Acute
Ethylbenzene	1E+01	2E+04 b	Subchronic	3E+04 b	Subchronic
Ethylene Dibromide	2E+01	7E+01	Subchronic	2E+02	Subchronic
Ethylene glycol					
Ethylene oxide		7E+03 °	Subchronic	5E+03 °	Subchronic
Formaldehyde		2E+03 °	Acute	1E+03 °	Acute
Furfural		2E+04	Subchronic	1E+05 ^a	Subchronic
Hexachloro-1,3-butadiene	6E+00				
Hexachlorobenzene	1E+01				
Hexachlorocyclopentadiene	2E+00				
Hexachloroethane	3E+01	9E+05 b	Acute	1E+07 b	Subchronic
Isophorone					
Lead	4E-01				
Manganese		3E+04	Subchronic	2E+04	Subchronic
Mercury	3E-01	8E+00	Acute	3E+01 b	Acute
Methanol	8E-01	4E+05 a,c	Acute	2E+05 °	Acute
Methoxyethanol acetate (R-R), 2-					
Methoxyethanol, 2-		7E+02 °	Acute	7E+02 °	Acute
Methyl bromide (Bromomethane)	2E+01	2E+02	Acute	9E+01	Acute
Methyl chloride (Chloromethane)	3E+01	7E+02	Acute	3E+02	Acute
Methyl ethyl ketone	4E+01	9E+04 ^a	Subchronic	6E+04 ^a	Subchronic
Methyl isobutyl ketone	3E+01	7E+04 ^a	Subchronic	4E+04 ^a	Subchronic
Methyl methacrylate	2E+02	3E+05 ^b	Subchronic	3E+05 b	Subchronic
Methyl tert-butyl ether		1E+04	Acute	1E+04	Acute
Methylcholanthrene, 3-	2E+01				
Methylene chloride	3E+01	1E+04 ^a	Acute	7E+03 ^a	Acute
N,N-Dimethylformamide					
Naphthalene	6E+00	1E+04 ^b	Subchronic	6E+03 ^a	Subchronic
n-Hexane		5E+03 ^a	Subchronic	8E+03 ^a	Subchronic
Nickel	5E+00	1E+05	Subchronic	7E+04	Subchronic
Nitrobenzene	1E+01				
Nitropropane, 2-		6E+02	Subchronic	9E+02	Subchronic
Nitrosodiethylamine	3E+01				
Nitrosodi-n-butylamine	2E+01				
N-Nitrosopyrrolidine	4E+01				

Table A-3. (continued)

Constituent	UTS Level (Nonwastewater) (mg/kg)	Land Application Unit Level (mg/kg)	Scenario	Wastepile Level (mg/kg)	Scenario
Phenol	6E+00				
Phthalic anhydride	3E+01				
Propylene oxide		2E+03 °	Subchronic	1E+03 [°]	Subchronic
Pyridine	2E+01				
Styrene		2E+05 ^b	Acute	5E+04 ^b	Acute
TCDD, 2,3,7,8-	1E-03				
Tetrachloroethane, 1,1,1,2-	6E+00				
Tetrachloroethane, 1,1,2,2-	6E+00	1E+05 ^b	Subchronic	2E+05 ^b	Subchronic
Tetrachloroethylene	6E+00	1E+03 ^a	Acute	1E+03 ^a	Acute
Toluene	1E+01	2E+04 ^a	Acute	2E+04 ^a	Acute
Toluidine, o-					
Trichloro-1,2,2-trifluoroethane, 1,1,2-	3E+01				
Trichlorobenzene, 1,2,4-	2E+01				
Trichloroethane, 1,1,1-	6E+00	1E+04 ^a	Acute	8E+03 ^a	Acute
Trichloroethane, 1,1,2-	6E+00				
Trichloroethylene	6E+00	8E+03 ^a	Subchronic	1E+04 ^a	Subchronic
Trichlorofluoromethane	3E+01	9E+04 ^a	Subchronic	5E+04 ^a	Subchronic
Triethylamine		2E+03	Subchronic	4E+03	Subchronic
Vanadium	2E-01	4E+03	Subchronic	2E+03	Subchronic
Vinyl acetate		7E+03 ^a	Subchronic	5E+03	Subchronic
Vinyl chloride	6E+00	1E+03	Subchronic	3E+02	Acute
Xylenes (total)	3E+01	8E+03 ^a	Acute	9E+03 ^a	Acute

^a Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component exceeds 1e-5 or HQ = 1.

See text for more details.

^b Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component less than 1e-5 or HQ = 1.

See text for more details.

^c Organic phase emissions greater than aqueous phase emissions, so result based on organic phase would be lower. See text & v. III for more details.

Table A-4. Comparison of Toxicity Characteristic Levels (TC) with Acute and Subchronic Waste Concentrations for Land Application Units and Wastepiles

Constituent	TC (Waste) Level (mg/kg)	Land Application Unit Level (mg/kg)	Scenario	Wastepile Level (mg/kg)	Scenario
Acetaldehyde		7E+03 °	Subchronic	5E+03 °	Subchronic
Acetone		3E+05 a,c	Acute	2E+05 °	Acute
Acetonitrile		3E+03	Subchronic	3E+03	Subchronic
Acrolein		2E-01	Acute	2E-01	Acute
Acrylamide					
Acrylic acid					
Acrylonitrile		5E+02	Acute	4E+02	Acute
Allyl chloride		1E+02	Subchronic	6E+01	Subchronic
Aniline					
Arsenic	2E+02	8E+03	Acute	3E+03	Acute
Barium	2E+03	3E+05	Subchronic	2E+05	Subchronic
Benzene	5E+01	2E+02	Acute	1E+02	Acute
Benzidine					
Benzo(a)pyrene					
Beryllium					
Bromodichloromethane					
Bromoform (Tribromomethane)					
Butadiene, 1,3-					
Cadmium					
Carbon disulfide		9E+03 ^a	Subchronic	4E+03 ^a	Subchronic
Carbon tetrachloride	8E+01	1E+03	Acute	9E+02	Acute
Chloro-1,3-butadiene, 2- (Chloroprene)					
Chlorobenzene	3E+05	4E+03 ^a	Subchronic	8E+03 ^a	Subchronic
Chlorodibromomethane					
Chloroform		6E+02	Acute	4E+02	Acute
Chlorophenol, 2-					
Chromium VI	1E+03	3E+04	Subchronic	2E+04	Subchronic
Cobalt		2E+03	Subchronic	1E+03	Subchronic
Cresols (total)	7E+04				
Cumene		1E+05 [°]	Subchronic	3E+04 b	Subchronic
Cyclohexanol		2E+01	Subchronic	2E+01	Subchronic
Dibromo-3-chloropropane, 1,2-		5E+02	Subchronic	4E+02	Subchronic
Dichlorobenzene, 1,2-		3E+05 ^b	Subchronic	2E+05 [°]	Subchronic
Dichlorobenzene, 1,4-	2E+04	3E+04 ^⁵	Acute	2E+04 ^b	Acute
Dichlorodifluoromethane		2E+04 ^a	Subchronic	8E+03 ^a	Subchronic
Dichloroethane, 1,2-	2E+01	1E+03	Acute	9E+02	Acute
Dichloroethylene, 1,1-	7E+01	1E+03	Subchronic	6E+02	Subchronic
Dichloropropane, 1,2-		2E+02	Subchronic	2E+02	Acute
Dichloropropene, cis-1,3-		6E+02	Subchronic	6E+02	Subchronic
Dichloropropene, trans-1,3-		7E+02	Subchronic	7E+02	Subchronic
					(continued)

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Table A-4. (continued)

Constituent	TC (Waste) Level (mg/kg)	Land Application Unit Level (mg/kg)	Scenario	Wastepile Level (mg/kg)	Scenario
Dimethylbenz(a)anthracene, 7,12-					
Dimethylphenol, 3,4-					
Dinitrotoluene, 2,4-	8E+00				
Dioxane, 1,4-		1E+04 ^c	Acute	3E+04 °	Acute
Diphenylhydrazine, 1,2-					
Epichlorohydrin		8E+02	Subchronic	8E+02	Subchronic
Epoxybutane, 1,2-		6E+03	Subchronic	5E+03	Subchronic
Ethoxyethanol acetate (R-R), 2-		4E+03 ^c	Acute	3E+03 °	Acute
Ethoxyethanol, 2-		3E+04 °	Acute	3E+04 °	Acute
Ethylbenzene		2E+04 b	Subchronic	3E+04 b	Subchronic
Ethylene Dibromide		7E+01	Subchronic	2E+02	Subchronic
Ethylene glycol					
Ethylene oxide		7E+03 °	Subchronic	5E+03 °	Subchronic
Formaldehyde		2E+03 °	Acute	1E+03 °	Acute
Furfural		2E+04	Subchronic	1E+05 ^a	Subchronic
Hexachloro-1,3-butadiene	9E+03				
Hexachlorobenzene	7E+03				
Hexachlorocyclopentadiene					
Hexachloroethane	5E+03	9E+05 b	Acute	1E+07 b	Subchronic
Isophorone					
Lead	2E+03				
Manganese		3E+04	Subchronic	2E+04	Subchronic
Mercury	4E+00	8E+00	Acute	3E+01	Acute
Methanol		4E+05 a,c	Acute	2E+05 ^c	
Methoxyethanol acetate (R-R), 2-					
Methoxyethanol, 2-		7E+02 °	Acute	7E+02 °	Acute
Methyl bromide (Bromomethane)		2E+02	Acute	9E+01	Acute
Methyl chloride (Chloromethane)		7E+02	Acute	3E+02	Acute
Methyl ethyl ketone	2E+04	9E+04 ^a	Subchronic	6E+04 ^a	Subchronic
Methyl isobutyl ketone		7E+04 ^a	Subchronic	4E+04 a	Subchronic
Methyl methacrylate		3E+05 ^b	Subchronic	3E+05 b	Subchronic
Methyl tert-butyl ether		1E+04	Acute	1E+04	Acute
Methylcholanthrene, 3-					
Methylene chloride		1E+04 ^a	Acute	7E+03 ^a	Acute
N,N-Dimethylformamide					
Naphthalene		1E+04 b	Subchronic	6E+03 ^a	Subchronic
n-Hexane		5E+03 ^a	Subchronic	8E+03 ^a	Subchronic
Nickel		1E+05	Subchronic	7E+04	Subchronic
Nitrobenzene	4E+02				
Nitropropane, 2-		6E+02	Subchronic	9E+02	Subchronic
Nitrosodiethylamine		02102	300011101110	02.02	345071101110

Table A-4. (continued)

Constituent	TC (Waste) Level (mg/kg)	Land Application Unit Level (mg/kg)	Scenario	Wastepile Level (mg/kg)	Scenario
Nitrosodi-n-butylamine					
N-Nitrosopyrrolidine					
Phenol					
Phthalic anhydride					
Propylene oxide		2E+03 °	Subchronic	1E+03 [°]	Subchronic
Pyridine	3E+02				
Styrene		2E+05 ^b	Acute	5E+04 b	Acute
TCDD, 2,3,7,8-					
Tetrachloroethane, 1,1,1,2-					
Tetrachloroethane, 1,1,2,2-		1E+05 ^b	Subchronic	2E+05 b	Subchronic
Tetrachloroethylene	3E+02	1E+03 ^a	Acute	1E+03 ^a	Acute
Toluene		2E+04 ^a	Acute	2E+04 ^a	Acute
Toluidine, o-					
Trichloro-1,2,2-trifluoroethane, 1,1,2-					
Trichlorobenzene, 1,2,4-					
Trichloroethane, 1,1,1-		1E+04 ^a	Acute	8E+03 ^a	Acute
Trichloroethane, 1,1,2-					
Trichloroethylene	7E+01	8E+03 ^a	Subchronic	1E+04 ^a	Subchronic
Trichlorofluoromethane		9E+04 ^a	Subchronic	5E+04 ^a	Subchronic
Triethylamine		2E+03	Subchronic	4E+03	Subchronic
Vanadium		4E+03	Subchronic	2E+03	Subchronic
Vinyl acetate		7E+03 ^a	Subchronic	5E+03	Subchronic
Vinyl chloride	1E+01	1E+03	Subchronic	3E+02	Acute
Xylenes (total)		8E+03 ^a	Acute	9E+03 ^a	Acute

^a Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component exceeds 1e-5 or HQ = 1. See text for more details.

^b Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component less than 1e-5 or HQ = 1. See text for more details.

[°] Organic phase emissions greater than aqueous phase emissions, so result based on organic phase would be lower. See text & v. III for more details.

Table A-5. Comparison of Toxicity Characteristic Levels with Chronic Waste Concentrations for Tanks

Constituent Name	TC (Leachate) (mg/L)	Aerated Treatment Tank Level (mg/L)	Nonaerated Treatment Tank Level (mg/L)	Storage Tank Level (mg/L)
Acetaldehyde		4E+02	4E+02	5E+03
Acetone		1E+07	8E+05	1E+07 ^b
Acetonitrile		2E+03	1E+03	2E+04
Acrolein		3E-01	3E-01	4E+00
Acrylamide		4E+04	3E+04	2E+05
Acrylic acid		5E+03 [°]	3E+03	3E+04
Acrylonitrile		1E+01	1E+01	2E+02
Allyl chloride		4E+00	5E+00	2E+02
Aniline		6E+02	4E+02	3E+03
Arsenic	5E+00			
Barium	1E+02			
Benzene	5E-01	3E+01	4E+01	1E+03
Benzidine		5E+04 ^b	3E+04 ^b	2E+05 ^b
Benzo(a)pyrene		1E+03 ^b	8E+02 b	6E+03 b
Beryllium				
Bromodichloromethane		2E+01	2E+01	6E+02
Bromoform (Tribromomethane)		6E+02	7E+02	1E+04 ^b
Butadiene, 1,3-		4E-01	4E-01	2E+01
Cadmium				
Carbon disulfide		3E+03 ^a	3E+03 ^a	1E+05 ^a
Carbon tetrachloride	5E-01	1E+01	2E+01	7E+02
Chloro-1,3-butadiene, 2- (Chloroprene)		3E+01	4E+01	1E+03
Chlorobenzene	1E+02	1E+02	1E+02	5E+03 ^a
Chlorodibromomethane		2E+01	3E+01	4E+02
Chloroform		1E+01	1E+01	4E+02
Chlorophenol, 2-		2E+01	2E+01	3E+02
Chromium VI	5E+00			
Cobalt				
Cresols (total)	2E+02	3E+02	2E+02	1E+03
Cumene		2E+03 [°]	2E+03 [°]	1E+05 [°]
Cyclohexanol		4E+00	3E+00	3E+01
Dibromo-3-chloropropane, 1,2-		2E+03 ^b	2E+03 [°]	2E+04 ^b
Dichlorobenzene, 1,2-		1E+03 ^b	2E+03 ^b	5E+04 ^b
Dichlorobenzene, 1,4-	8E+00	5E+03 ^b	8E+03 ^b	2E+05 ^b
Dichlorodifluoromethane		8E+02 ^a	1E+03 ^a	5E+04 ^a
Dichloroethane, 1,2-	5E-01	1E+01	2E+01	4E+02
Dichloroethylene, 1,1-	7E-01	4E+00	5E+00	2E+02
Dichloropropane, 1,2-		2E+01	3E+01	9E+02

Table A-5. (continued)

Constituent Name	TC (Leachate) (mg/L)	Aerated Treatment Tank Level (mg/L)	Nonaerated Treatment Tank Level (mg/L)	Storage Tank Level (mg/L)
Dichloropropene, cis-1,3-		8E+00	1E+01	3E+02
Dichloropropene, trans-1,3-		1E+01	1E+01	3E+02
Dimethylbenz(a)anthracene, 7,12-		4E+03 ^b	3E+03 ^b	2E+04 b
Dimethylphenol, 3,4-				
Dinitrotoluene, 2,4-	1E-01	2E+03 ^b	1E+03 ^b	7E+03 ^b
Dioxane, 1,4-		9E+04	5E+04	6E+05
Diphenylhydrazine, 1,2-		5E+02 b	4E+02 b	3E+03 b
Epichlorohydrin		2E+03	2E+03	1E+04
Epoxybutane, 1,2-		1E+02	2E+02	4E+03
Ethoxyethanol acetate (R-R), 2-		7E+04	5E+04	4E+05
Ethoxyethanol, 2-		4E+05	3E+05	1E+07 b
Ethylbenzene		5E+03 ^b	7E+03 ^b	3E+05 ^b
Ethylene Dibromide		2E+00	3E+00	5E+01
Ethylene glycol		1E+07 ^b	1E+07 ^b	1E+07 b
Ethylene oxide		8E+00	7E+00	1E+02
Formaldehyde		4E+03 ^c	3E+03 [°]	2E+04 °
- Furfural		1E+04	1E+04	9E+04
Hexachloro-1,3-butadiene	5E-01	2E+01 b	4E+01 a	1E+03 b
Hexachlorobenzene	1E-01	5E+00	2E+01 b	2E+02 b
Hexachlorocyclopentadiene		7E-01	2E+00	7E+01 b
Hexachloroethane	3E+00	2E+02 b	3E+02 ^b	5E+03 b
sophorone		2E+03	1E+03	1E+04 ^b
_ead	5E+00			
Manganese				
Mercury	2E-01	2E+00 ^b	3E+00 ^b	1E+02 ^b
Methanol		1E+07 ^b	1E+07 ^b	1E+07 ^b
Methoxyethanol acetate (R-R), 2-		9E+03	6E+03	5E+04
Methoxyethanol, 2-		3E+04	2E+04	2E+05
Methyl bromide (Bromomethane)		2E+01	3E+01	9E+02
Methyl chloride (Chloromethane)		1E+02	2E+02	7E+03 ^a
Methyl ethyl ketone	2E+02	3E+04	2E+04	3E+05 ^b
Methyl isobutyl ketone		1E+03	2E+03	2E+04 ^a
Methyl methacrylate		7E+03	1E+04	2E+05 ^b
Methyl tert-butyl ether		2E+04	3E+04	6E+05 ^b
Methylcholanthrene, 3-		1E+03 ^b	9E+02 ^b	6E+03 ^b
Methylene chloride		5E+02	7E+02	2E+04 ^a
N,N-Dimethylformamide		1E+05	8E+04	6E+05
Naphthalene		3E+01 ^b	5E+01 ^b	8E+02 ^b
n-Hexane		1E+03 ^a	1E+03 ^a	6E+04 ^a

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Table A-5. (continued)

Constituent Name	TC (Leachate) (mg/L)	Aerated Treatment Tank Level (mg/L)	Nonaerated Treatment Tank Level (mg/L)	Storage Tank Level (mg/L)
Nickel				
Nitrobenzene	2E+00	1E+02	1E+02	1E+03
Nitropropane, 2-		4E-01	4E-01	4E+00
Nitrosodiethylamine		2E-01	1E-01	1E+00
Nitrosodi-n-butylamine		3E-01	4E-01	7E+00
N-Nitrosopyrrolidine		3E+03	2E+03	1E+04
Phenol		1E+04	1E+04	9E+04 ^b
Phthalic anhydride		1E+07 ^b	1E+07 ^b	1E+07 b
Propylene oxide		3E+02	3E+02	3E+03
Pyridine	5E+00	7E+02	5E+02	5E+03
Styrene		6E+03 ^b	8E+03 b	2E+05 b
TCDD, 2,3,7,8-		5E-03 ^b	6E-03 ^b	4E-02 b
Tetrachloroethane, 1,1,1,2-		4E+01	6E+01	2E+03 ^a
Tetrachloroethane, 1,1,2,2-		1E+01	1E+01	2E+02
Tetrachloroethylene	7E-01	4E+02 ^a	5E+02 ^a	2E+04 a
Toluene		2E+03 ^a	2E+03 ^a	9E+04 ^b
Toluidine, o-		3E+02	2E+02	2E+03
Trichloro-1,2,2-trifluoroethane, 1,1,2-		1E+05 ^b	2E+05 ^b	1E+07 b
Trichlorobenzene, 1,2,4-		2E+03 b	3E+03 b	6E+04 b
Trichloroethane, 1,1,1-		4E+03 ^a	5E+03 ^a	2E+05 b
Trichloroethane, 1,1,2-		2E+01	3E+01	7E+02
Trichloroethylene	5E-01	1E+02	2E+02	6E+03 ^a
Trichlorofluoromethane		3E+03 ^a	3E+03 ^a	1E+05 ^a
Triethylamine		1E+02	1E+02	2E+03
Vanadium				
Vinyl acetate		2E+03	2E+03	4E+04 ^a
Vinyl chloride	2E-01	2E+00	3E+00	1E+02
Xylenes (total)		2E+03	3E+03 b	9E+04 ^b

 ^a Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component exceeds 1e-5 or HQ = 1. See text for more details.
 ^b Aqueous-phase result exceeds solubility or Csat. Risk for pure, organic-phase component less than

¹e-5 or HQ = 1. See text for more details.

^c Organic phase emissions greater than aqueous phase emissions, so result based on organic phase would be lower. See text & v. III for more details.